

FINAL

**INTERIM ASSESSMENT OF THE PRESENCE AND CAUSES
OF SEDIMENT TOXICITY IN
BRYAN MUNICIPAL LAKE, SEGMENT 1209A
AND
FINFEATHER LAKE, SEGMENT 1209B**

Prepared For

TOTAL MAXIMUM DAILY LOAD PROGRAM

**TEXAS NATURAL RESOURCES CONSERVATION COMMISSION
P.O. BOX 13087, MC - 150
AUSTIN, TEXAS 78711-3087**

Prepared By

PARSONS

**PROJECT LEAD ORGANIZATION
8000 CENTRE PARK DR., SUITE 200
AUSTIN, TEXAS 78754
512-719-6000**

FEBRUARY 2003

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**PREPARED IN COOPERATION WITH THE TEXAS COMMISSION ON
ENVIRONMENTAL QUALITY AND THE U.S. ENVIRONMENTAL PROTECTION AGENCY**

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Protection Agency through the Texas Commission on Environmental Quality**

EXECUTIVE SUMMARY

Bryan Municipal Lake Segment 1209A and Finfeather Lake Segment 1209B Toxicity in Sediment

The Texas Commission on Environmental Quality (TCEQ) is responsible for administering provisions of the constitution and laws of the State of Texas to promote judicious use and the protection of the quality of waters in the State. A major aspect of this responsibility is the continuous monitoring and assessment of water quality to evaluate compliance with state water quality standards which are established within Texas Water Code, '26.023 and Title 30 Texas Administrative Code, '307.1-307.10. Texas Surface Water Quality Standards 30 TAC 370.4(d) specify that surface waters will not be toxic to aquatic life. Pursuant to the federal Clean Water Act '303(d), states must establish Total Maximum Daily Loads (TMDLs) for pollutants contributing to violations of water quality standards. The purpose of this TMDL Study was to assess the presence and causes of ambient toxicity in seven Texas waterbodies listed on the Draft 2000 Federal Clean Water Act (CWA) '303(d) List in an effort to comply with Texas law.

In order to assess the waterbodies, this study provided goals as follows:

- Confirmation that toxicity is present more than 10% of the time, through the collection of up to date toxicity testing.
- The identification of the substance(s) or factors causing the toxicity where present.
- The identification of the sources of the toxicant(s).
- Confirmation, via chemical analysis, that water quality standards are being maintained.

This study was limited to the following seven waterbodies of concern:

1. Alligator Bayou (Segment 0702A) in Jefferson County (toxicity in water and sediment),
2. Bryan Municipal Lake (Segment 1209A) in Brazos County (toxicity in sediment),
3. Finfeather Lake (Segment 1209B) in Brazos County (toxicity in sediment),
4. Vince Bayou (Segment 1007A) in Harris County (toxicity in sediment),
5. Arroyo Colorado Tidal (Segment 2201) in Cameron County (toxicity in sediment),
6. Rio Grande (Segment 2304) in Kinney, Maverick, and Webb Counties (toxicity in water), and
7. Rio Grande (Segment 2306) in Presidio County (toxicity in water).

The TCEQ selected Parsons to conduct a more thorough and intensive assessment of the existence of toxicity and identification of likely toxicants in the waterbodies. The Texas Surface Water Quality Standards specify that surface waters will not be toxic to aquatic life. Pursuant to the federal Clean Water Act §303(d), States must establish total maximum daily loads (TMDLs) for pollutants contributing to violations of surface water quality standards. Ambient toxicity testing complements routine chemical monitoring to identify waterbodies

with aquatic life impairment. The waterbody assessments are each described in six different reports. Finfeather Lake (FFL) and Bryan Municipal Lake (BML) are described in the same report due to their close proximity and likely cause.

Both FFL and BML were sampled for a total of six events each during an initial 5-month period from April 2001 to August 2001. Once TIE studies began, routine whole sediment toxicity testing ceased. Detections of aluminum, arsenic, copper, and zinc at Station 11799 of FFL were above the corresponding sediment screening levels. Station 11798 of FFL indicated detections of aluminum, arsenic, cadmium, chromium, copper, lead, nickel and zinc were also above the corresponding screening levels. Detections of aluminum, arsenic, cadmium, copper, lead, and zinc at Station 11793 of BML were above the corresponding screening levels.

Toxicity test results for sediment samples collected in April and May 2001 indicated the sediments were significantly toxic due to lethality at Stations 11798 and 11800 in FFL and Station 11793 in BML to *Chironmus tentans* and *Hyallolella azteca* species using whole sediment toxicity test methods. Statistically significant sublethal effects were also observed in sediment collected from Station 11800 in FFL and Stations 11792, 11793, and 11794 in BML. Due to the toxicity of the sediments, a TIE was initiated for both FFL and BML. Phase I of the TIE was initiated at Station 11798 in FFL and Station 11793 in BML. Since it is very likely that the same contaminant is affecting both lakes, it was decided to focus on the most toxic site first (in FFL) for the TIE. Refer to the Table ES.1 and ES.2 for details.

**Table ES.1
 Sediment Toxicity Test Results**

| | | % Survival | | Sub-Lethal Effect | |
|--|-----------|-------------------|----------------|-------------------|----------------|
| | | | | Growth | |
| | | Chironmus tentans | Hyaella azteca | Chironmus tentans | Hyaella azteca |
| Bryan Municipal Lake 1209A April 19,2001 | Control | 81 | 91 | 0.706 | 0.112 |
| | 11792 | 69 | 79 | 0.455 | 0.091 |
| | 11793 | 31 | 84 | NA | 0.086 |
| | 11794 | 71 | 85 | 0.367 | 0.115 |
| Bryan Municipal Lake 1209A May 21, 2001 | Control | 86 | 99 | CW | 0.167 |
| | 11792 | 66 | 84 * | CW | 0.128 |
| | 11793 | 70 | 92 | NA | 0.109 |
| | 11794 | 84 | 90 * | CW | 0.145 |
| Finfeather Lake 1209B April 19,2001 | Control | 74 | 91 | CW | 0.112 |
| | 11798 | 33 | 54 | NA | NA |
| | 11799 | 58 | 86 | CW | 0.107 |
| | 11800 | 46 | 89 | NA | 0.094 |
| | 11800-Dup | 49 | 88 | NA | 0.071 |
| Finfeather Lake 1209B May 21, 2001 | Control | 75 | 99 | CW | 0.167 |
| | 11798 | 23 | 84 | NA | 0.091 |
| | 11799 | 79 | 80 | CW | 0.146 |
| | 11800 | 69 | 56 | CW | 0.112 |

Bold/Shaded cell - denotes significant difference from the control; duplicate is for quality control purposes only
 * Note that while statistically significant mortality effects were observed, the results did not exceed recommended criteria.
 CW- Control weight below minimum of 0.48 mg AFDW
 NA = Not Analyzed

**Table ES.2
 Summary of Sediment Toxicity Test Results**

| Station | Lethal <i>C. tentans</i> | Lethal <i>H. azteca</i> | Sublethal <i>C. tentans</i> | Sublethal <i>H. azteca</i> |
|---------|-----------------------------|----------------------------|--------------------------------|-------------------------------|
| 11792 | 1/2 | 0/2 | 1/1 | 1/2 |
| 11793 | 1/2 | 0/2 | 1/1 | 2/2 |
| 11794 | 0/2 | 0/2 | 1/1 | 0/2 |
| 11798 | 2/2 | 1/2 | 0/0 | 2/2 |
| 11799 | 1/2 | 0/2 | 0/0 | 0/2 |
| 11800 | 1/2 | 1/2 | 0/0 | 1/2 |

U.S. EPA has not finalized sediment pore water or whole sediment Toxicity Identification Evaluation (TIE) methodology. Draft sediment TIE guidelines are available for pore waters and elutriates (EPA 1991) and closely follow effluent TIE procedures. Some whole sediment procedures for reducing toxicity of specific toxicant classes have been reported in the literature; however, whole sediment TIE procedures are not published in guideline format (Ho et al. 2002). Therefore, a tiered approach based on pore water tests was employed in this project (Ankley and Schubauer-Berigan 1995). Additional whole sediment TIE procedures were performed on FFL. Generally, 40-60% of sediment volume was isolated as pore water. *C. dubia* was chosen for pore water testing because of test volume requirements. *Hyalella azteca* and *Chironomus tentans* were also used to test whole sediments. Table ES.3 provides the TIE toxicity test results.

Table ES.3
***C. dubia*, 7-Day Toxicity Tests using Sediment Pore Water**

| Sample Date | Test Date | Station | Treatment | Results | | |
|-------------|------------|---------|----------------|--------------|---------|------------|
| | | | | Reproduction | Std Dev | % Survival |
| 6/5/2001 | 7/31/2001 | 11798 | RHW (Control) | 26.4 | 2.19 | |
| | | | Baseline 100% | 15 | 2.55 | |
| 6/5/2001 | 7/31/2001 | 11793 | RHW (Control) | 26.4 | 2.19 | |
| | | | Baseline 100% | 15.2 | 3.11 | |
| 6/5/2001 | 8/25/2001 | 11798 | RHW (Control) | 24.2 | 2.28 | |
| | | | Baseline 100% | 14.6 | 1.82 | |
| 10/30/2001 | 2/2/2002 | 11798 | RMHW (Control) | 23 | 3.16 | 100 |
| | | | Baseline 100% | 11.4 | 0.894 | 100 |
| 10/30/2001 | 6/6/2002 | 11798 | RMHW (Control) | 25.8 | 3.03 | 100 |
| | | | Baseline 100% | 0 | 0 | 20 |
| 10/30/2001 | 6/26/2002 | 11798 | RMHW (Control) | 23.8 | 2.57 | 100 |
| | | | Baseline 100% | 10.6 | 4.9 | 100 |
| 10/30/2002 | 12/12/2002 | 11798 | RMHW (Control) | 27.4 | 2.07 | 100 |
| | | | Baseline 100% | 2 | 0.82 | 80 |

A summary of all TIEs performed in this study is provided in Table ES.4.

Table ES.4
Sediment Toxicity Identification Evaluation Procedures

| Test Date | Test Type | Station | Organism | Effective Treatment |
|-------------------------|----------------|---------|------------------|---------------------|
| July 13-23, 2001 | Pore Water | 11793 | <i>C. dubia</i> | EDTA |
| July 13-23, 2001 | Pore Water | 11798 | <i>C. dubia</i> | None |
| Aug. 25 – Sept. 4, 2001 | Pore Water | 11798 | <i>C. dubia</i> | SIR300, SIR900 |
| February 2-12, 2002 | Pore Water | 11798 | <i>C. dubia</i> | SIR300, SIR900 |
| March 10-20, 2002 | Whole Sediment | 11798 | <i>H. azteca</i> | SIR900* |
| June 6-16, 2002 | Pore Water | 11798 | <i>C. dubia</i> | EDTA, SIR300 |
| June 6-16, 2002 | Pore Water | 11800 | <i>C. dubia</i> | EDTA |

* *H. azteca* growth not significant different from control sediment.
 60% survival in SIR900, 68.3% survival in control.

In July 2001, it was determined that sediment pore water was not acutely toxic; however, pore water in sediment from stations 11793 and 11798 produced persistent and repeatable toxic effects on *C. dubia* reproduction. EDTA treatment reduced pore water toxicity at Station 11793, but not at 11798. A subsequent TIE was performed with several treatment media selective for metals which removed toxicity in August 2001. Each TIE treatment improved *C. dubia* reproduction relative to untreated pore waters. Because arsenic was suspected as a causative toxicant, total arsenic, arsenate and arsenite levels were quantified in each TIE treatment. Arsenic concentrations in pore waters were not sufficiently elevated to solely cause toxicity; total arsenic pore water concentration was 266 µg/L. Previous investigators found that arsenic treatment up to 1.46 mg/L did not significantly affect *C. dubia* reproduction.

In February 2002, another TIE was performed on station 11798 pore waters. Metal bioavailability and toxicity were reduced with increasing water hardness. SIR-900 and SIR-900 + SIR-300 treatments significantly increased *C. dubia* reproduction in undiluted and 50% dilution, respectively. SIR-300 increased neonate production at 50% dilution, although not significantly, by an average of four neonates/female relative to baseline pore waters. Baseline pore water copper concentration was 722 µg/L, a value two orders of magnitude higher than previously reported lowest observed effect concentrations for *C. dubia* reproduction in

laboratory water and substantially higher than the Texas Surface Water Quality Standards (TSWQS) of 48.6 µg/L.

Whole sediments from Station 11798 were amended with SIR-300, SIR-900 and SIR-300 + SIR-900 at a 1:4, volume to volume (V:V) ratio in March 2002. Following a 10-day exposure period, *H. azteca* growth was significantly improved, relative to reference sediment from the University of North Texas Water Research Field Station, by only SIR-900 treatments.

In June 2002, a TIE was performed with Station 11798 sediment pore waters. Because multiple metals were measured in pore waters, a toxic unit approach was taken to evaluate metal pore water toxicity. Concentrations of zinc, iron, lead and barium decreased 26%, 32%, 37% and 96%, respectively, with SIR-300 treatments.

A subsequent TIE study with SIR-300 was conducted to further remove metal contaminants from Station 11798 pore waters whereby reduced toxicity was more clearly assigned to potentially causative toxicants. In addition, contaminant addition procedures (Phase III TIE) were subsequently performed to recreate pore water toxicity and provide corroborating information. Phase III TIE procedures were conducted such that pore waters were first treated with SIR 300 followed by reintroduction of copper, lead or zinc at nominal concentrations. Results indicated copper and zinc as the primary factors affecting aquatic life. A toxic units approach suggests copper to be of greater concern than zinc, however, 100% mortality was observed in zinc treated pore waters. This information also indicates metals similar to copper and zinc (examples) are of concern in both pore water and whole sediment. The exception to this is arsenic which does not appear to be a problem in pore waters. The containment addition test results suggest there is more toxicity effect from zinc than copper.

Parsons' recommends periodic monitoring of the sediment toxicity and development of a legacy TMDL for copper and zinc.

Table ES.5
Finfeather Lake Segment 1209B
Whole Sediment Chemistry and Toxic Units

| | Station ID 11799 | Station ID 11798 | Station ID 11799 | Station ID 11798 | Station ID 11800 | | |
|-----------|----------------------------------|----------------------------------|----------------------------------|---------------------------------|---------------------------------|--------------------------------|------------------|
| PARAMETER | Sample Collected 5/21/2001 | Sample Collected 7/18/2001 | Sample Collected 7/18/2001 | Sample Collected 5/9/2002 | Sample Collected 5/9/2002 | Lowest Screening Values* | UNITS |
| Toxicity | Toxic ¹ | NA | NA | NA | NA | | |
| Arsenic | 58.5 (8.08) | 196 (27.1) | 28.8 (4.0) | 79.2 (10.9) | 160 (22.1) | 7.24 | mg/Kg- dry wt |
| Copper | 65.4 (3.5) | 575 (30.7) | 44.5 (2.37) | 171 (9.14) | 113 (6.04) | 18.7 | mg/Kg- dry wt |
| Lead | 17.5 | 56.9 (1.88) | 12.6 | 33.3 (1.10) | 51.8 (1.71) | 30.24 | mg/Kg- dry wt |
| Zinc | 241 (1.94) | 1280 (10.3) | 151 (1.22) | 447 (3.61) | 466 (3.76) | 124 | mg/Kg- dry wt |

Notes:

* Criteria is from *Equilibrium and Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices* tables. The value is the lowest value from the Indices as stated in the Appendix.

mg/kg-dry = milligrams per kilogram dry weight

¹ No significant difference from control for survival and growth of *C. tentans* in 10 day sediment exposures; significant difference in survival of *H. azteca* in 10 day sediment exposure.

() = Toxic Units; calculated by dividing detected metal concentration by the lowest screening level.

Bold and highlighted results indicate TU is above 1.0

Table ES.6
Bryan Municipal Lake 1209A
Whole Sediment Chemistry and Toxic Units

| | Station ID 11793 | Station ID 11793 | Station ID 11792 | Station ID 11793 | Station ID 11794 | | |
|-----------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|--------------------------------|------------------|
| PARAMETER | Sample Collected 5/21/2001 | Sample Collected 7/18/2001 | Sample Collected 7/12/2002 | Sample Collected 7/12/2002 | Sample Collected 7/12/2002 | Lowest Screening Values* | UNITS |
| Toxicity | Toxic ¹ | NA | Not Toxic ² | Toxic ³ | Not Toxic ² | | |
| Arsenic | 57.6 (7.96) | 95.8 (13.2) | 17.8 (2.46) | 90.2 (12.5) | 141 (19.5) | 7.24 | mg/Kg- dry wt |
| Copper | 52.5 (2.8) | 40.6 (2.17) | 13.9 | 44 (2.35) | 178 (9.5) | 18.7 | mg/Kg- dry wt |
| Lead | 36.4 (1.21) | 37.1 (1.23) | 21.8 | 42.3 (1.40) | 99.7 (3.30) | 30.24 | mg/Kg- dry wt |
| Zinc | 227 (1.83) | 183 (1.5) | 67.5 | 215 (1.73) | 799 (6.44) | 124 | mg/Kg- dry wt |

Notes:

* Criteria is from *Equilibrium and Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices* tables. The value is the lowest value from the Indices as stated in the Appendix.

mg/kg-dry = milligrams per kilogram dry weight

¹ No significant difference from control for survival of *C. tentans* and *H. azteca* in 10 day sediment exposures; significant difference in growth for sublethal effects of *H. azteca*.

² No significant difference from control for survival and growth of *H. azteca* in 10 day sediment exposures.

³ No significant difference from control for survival of *H. azteca* in 10 day sediment exposures; although significant difference in growth for sublethal effects of *H. azteca*.

() = Toxic Units; calculated by dividing detected metal concentration by the lowest screening level.

Bold and highlighted results indicate TU is above 1.0

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LIST OF ACRONYMS

| | |
|-------|---|
| BML | Bryan Municipal Lake |
| Cfs | Cubic feet per second |
| CWA | Clean Water Act |
| DQO | Data quality objectives |
| FFL | Finfeather Lake |
| FM | Farm to market |
| Km | Kilometer |
| LCS | Laboratory control standards |
| m | Meter |
| µg/L | Microgram per liter |
| mg/L | Milligrams per liter |
| MS | Matrix Spike |
| MSD | Matrix Spike Duplicate |
| QAO | Quality assurance officer |
| QAPP | Quality assurance projected plan |
| QC | Quality control |
| SSI | Screening site inspection |
| SWQM | Surface water quality manual |
| TAC | Texas Administrative Code |
| TIE | Toxicity identification evaluation |
| TMDL | Total maximum daily load |
| TCEQ | Texas Commission on Environmental Quality |
| TNRCC | Texas Natural Resources Conservation Commission |
| TSWQS | Texas Surface Water Quality Standards |
| USEPA | United States Environmental Protection Agency |
| USGS | United States Geologic Survey |

SECTION 1 INTRODUCTION

The Texas Commission on Environmental Quality (TCEQ) is responsible for administering provisions of the constitution and laws of the State of Texas to promote judicious use and the protection of the quality of waters in the State. A major aspect of this responsibility is continuous monitoring and assessment of water quality to evaluate compliance with the state water quality standards established within Texas Water Code, §26.023 and Title 30 Texas Administrative Code (TAC) §§307.1-307.10. Texas Surface Water Quality Standards 30 TAC 370.4(d) specify that surface waters will not be toxic to aquatic life. Pursuant to the federal Clean Water Act (CWA) §303(d), states must establish total maximum daily loads (TMDLs) for pollutants contributing violations of water quality standards. The purpose of this work is to document the assessment of the presence and causes of ambient sediment toxicity in Finfeather and Bryan Municipal Lakes, two Texas water bodies on the 1998 and Draft TCEQ 2000, Federal CWA §303(d) lists.

Ambient sediment toxicity testing complements routine chemical monitoring to identify water bodies with aquatic life impairment. Finfeather Lake and Bryan Municipal Lake are shown to be contaminated with arsenic due to long-term releases of arsenic compounds into Finfeather Lake from an adjacent pesticide formulating facility. The two lakes are located on an unnamed tributary within the city of Bryan. Special studies first conducted by the Texas Water Quality Board in 1973 revealed high levels of arsenic in the lakes, and the unnamed tributary, with adverse impacts on the biological community. This led to a long-term remediation of the problem, which continues today.

The U.S. Environmental Protection Agency's (USEPA) Region 6 laboratory in Houston performed the toxicity testing by standard protocols. Based on this toxicity testing data, the Finfeather Lake and Bryan Municipal Lake were identified on the 1998 and Draft TCEQ 2000, CWA §303(d) list as impaired due to potential acute or chronic toxicity of ambient sediments. However, chemical toxicants or stressors responsible for the observed toxic effects in the laboratory have not yet been identified, although arsenic is suspected. Thus, Finfeather Lake and Bryan Municipal Lake are candidates for a more intensive assessment to confirm the occurrence of toxic conditions or nonsupport of aquatic life uses, and to determine the causes and sources of toxicity. Based on results of this assessment, the TCEQ may elect to remove one or both of the water bodies from the §303(d) list for sediment toxicity, or to develop a TMDL(s) for identified toxicants or stressors.

1.1 BACKGROUND INFORMATION

Finfeather Lake was formed in the 1930s by the construction of a railroad track across the stream. Finfeather Lake is fed by an unnamed stream, and the lake and watershed lie in an industrial area of Bryan, Texas. The lake has a surface area of 18.5 acres and an average depth of 5-7 feet. The Bryan Municipal Power Station has been the main discharger into the lake in recent times, but has reduced the discharge into the lake. Discharges from Finfeather Lake flow into an unnamed stream, this stream flows through a residential area, then into

Williamson Park and into Bryan Municipal Lake. Bryan Municipal Lake is a shallow lake, adjacent to the Bryan Municipal Golf Course. Bryan Municipal Lake has a surface area of approximately 14 acres, and an average depth of 2-3 feet. Discharges from Bryan Municipal Lake flow into Burton Creek, then to Charters Creek, and into the Navasota River, segment 1209 of the Brazos River Basin. The primary potential toxicant of concern for these bodies of water is elevated concentrations of arsenic in sediment. See Figure 1 for an overhead view of the two lakes. Numerous studies have been conducted on the lakes including one by the current owner of the site, Elf-Atochem (Parametric 1994).

1.2 DESCRIPTION OF THE SAMPLING STATIONS AT FINFEATHER LAKE

The TCEQ established three sampling stations in Finfeather Lake (Figure 2). The sampling station descriptions are as follows:

- 11798: Finfeather Lake near Dam Spillway
- 11799: Finfeather Lake Main Body
- 11800: Finfeather Lake Headwater

1.3 DESCRIPTION OF THE SAMPLING STATIONS AT BRYAN MUNICIPAL LAKE

The TCEQ established three sampling stations in Bryan Municipal Lake (Figure 2). The sampling station descriptions are as follows:

- 11792: Bryan Municipal Lake Near Dam Spillway
- 11793: Bryan Municipal Lake Main Body
- 11794: Bryan Municipal Lake Headwater

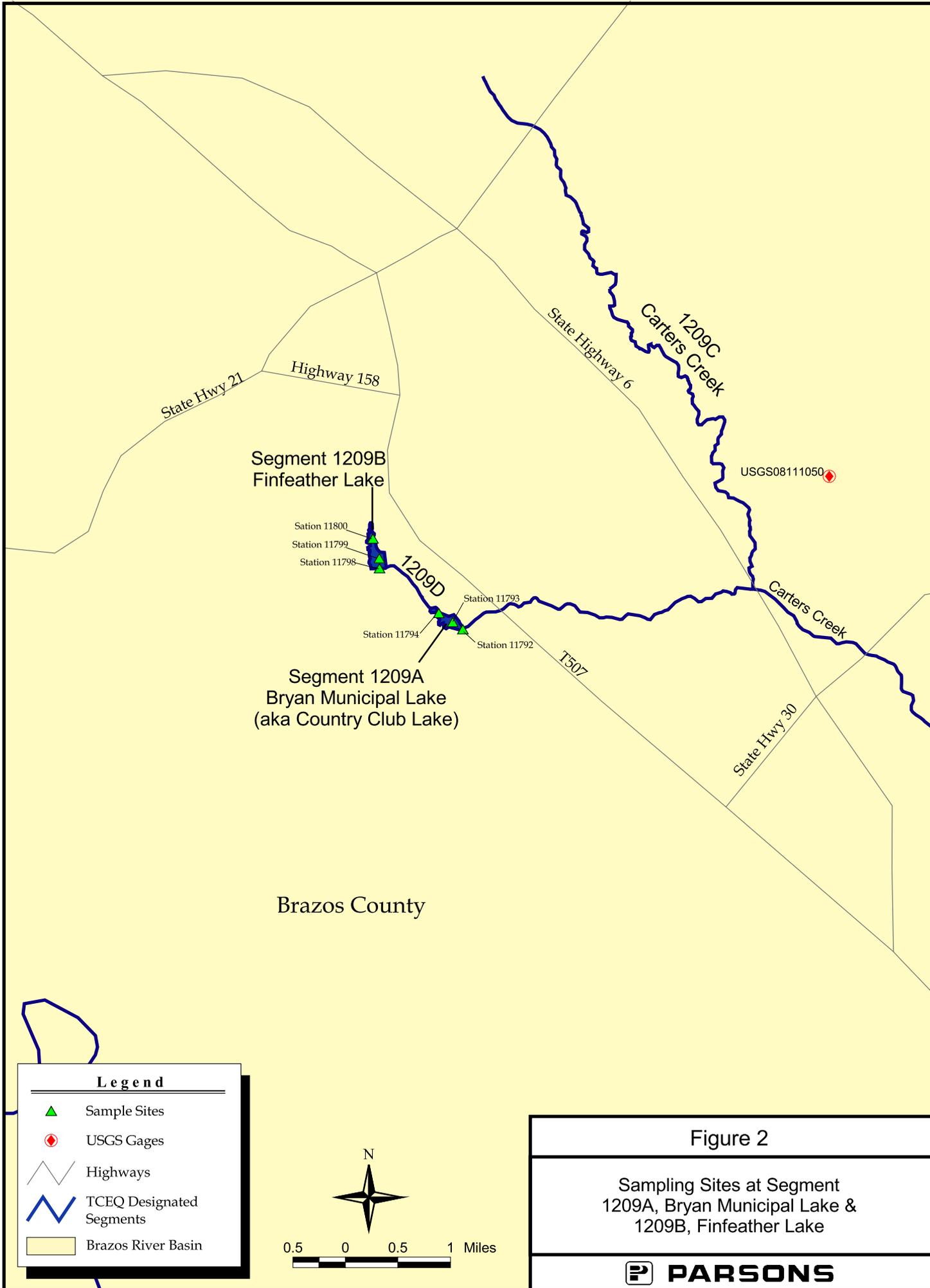


1000 0 1000 Feet

Figure 1

GPS Locations at
Bryan Municipal Lake and
Finfeather Lake





Legend

-  Sample Sites
-  USGS Gages
-  Highways
-  TCEQ Designated Segments
-  Brazos River Basin

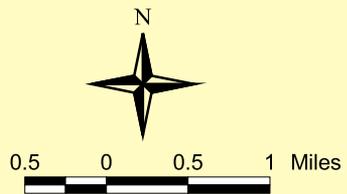


Figure 2

Sampling Sites at Segment 1209A, Bryan Municipal Lake & 1209B, Finfeather Lake

 **PARSONS**

SECTION 2 PROBLEM DEFINITION

Ambient sediment toxicity monitoring showed significant lethality in three out of five samples to *Ceriodaphnia dubia* (*C. dubia*) and one out of five samples to Fathead minnows obtained from Finfeather Lake from 1992 to 1997. In addition, Bryan Municipal Lake exhibited significant lethality to one out of six samples to *C. dubia* and sublethal effects to *C. dubia* in two out of six samples. See Tables 2.1 and 2.2 - Historical Toxicity Tests Results Justifying 303(d) listing for Finfeather Lake and Bryan Municipal Lake, respectively, for a breakdown of the water and sediment toxicity data. These test results required the TCEQ to list both Finfeather and Bryan Municipal lakes on the state's 303(d) list. It should be noted that toxicity was observed in 5 of the 16 (31%) ambient water test, which should have resulted in the TCEQ listing FFL as not meeting aquatic life uses due to ambient water toxicity.

The TCEQ's 303(b) report for 1999 and the draft 303(d) report for 2000 document significant sediment toxicity in both lakes. Toxicity monitoring was conducted from February 1990 to August 1997 for Finfeather Lake and May 1990 to August 1997 for Bryan Municipal Lake.

From 1995 through 2000, sediment samples collected from Finfeather Lake were significantly lethal to *C. dubia*, one out of three times. Bryan Municipal Lake did not show lethality effects to *C. dubia* or Fathead minnows during any of the four samples, from 1995 through 2000, and only once showed sublethal effects to *C. dubia* (Tables 2.3 and 2.4).

The historical sediment toxicity tests were performed by the USEPA Region 6 laboratory using the sediment elutriate test. This test requires mixing the sediment in lab water for a specified period of time then letting the sediment settle. The toxicity test is performed on the supernatant. It is understood that this test maximizes the amount of potentially toxic dissolved compound in the supernatant and may overstate the actual whole sediment toxicity to endemic benthic organisms. In addition, measured water column concentrations may also be overstated due to the elutriate procedure.

Guidance developed by TCEQ for Texas Surface and Drinking Water Quality Data, requires that data used to evaluate waterbodies for 303(d) listing and TMDL development not be more than five years old. Therefore, tasks within this assessment include collection of additional water and sediment samples to confirm the toxicity. Then determine the cause and the source of the toxicity. The results of the analysis will determine whether to proceed with TMDL development or establish the basis for removing the bayou from the 303(d) list.

Tables 2.5 and 2.6 contain a summary of the historical sediment chemistry detections measured over the past 5 years. Table 2.5 presents the data collected from Station 11798 for Finfeather Lake, while Table 2.6 data is for Bryan Municipal Lake. For Finfeather Lake, Arsenic, Cadmium, Chromium, Copper, lead, Mercury, Nickel, and Zinc all exceeded the screening values listed in at least one sample. Of the exceedances, Arsenic has the highest percent exceedance of the screening values.

**Table 2.1
Historical Toxicity Tests Results Justifying 303(d) Listing for Finfeather Lake***

| Species | Number of Tests | Exhibits Primary Toxicity | Exhibits Secondary Toxicity | Total Exhibiting Toxicity | Total % Toxic |
|----------------------------|------------------------|----------------------------------|------------------------------------|----------------------------------|----------------------|
| <u>Ceriodaphnia dubia</u> | | | | | |
| Water Toxicity | 16 | 4 | 1 | 5 | 31 |
| Sediment Toxicity | 5 | 3 | 2 | 5 | 100 |
| <u>Pimephales promelas</u> | | | | | |
| Water Toxicity | 15 | 1 | NP | 1 | 7 |
| Sediment Toxicity | 5 | 1 | NP | 1 | 20 |
| Total | 41 | 9 | 3 | 12 | |

NP = Not Performed

* Samples were collected from 18 sampling events that occurred between February 1990 and August 1997

Table 2.2
Historical Toxicity Tests Results Justifying 303(d) Listing for Bryan Municipal Lake*

| Species | Number of Tests | Exhibits Primary Toxicity | Exhibits Secondary Toxicity | Total Exhibiting Toxicity | Total % Toxic |
|----------------------------|------------------------|----------------------------------|------------------------------------|----------------------------------|----------------------|
| <u>Ceriodaphnia dubia</u> | | | | | |
| Water Toxicity | 15 | 0 | 3 | 3 | 20 |
| Sediment Toxicity | 6 | 1 | 2 | 3 | 50 |
| <u>Pimephales promelas</u> | | | | | |
| Water Toxicity | 14 | 0 | NP | 0 | 0 |
| Sediment Toxicity | 6 | 0 | NP | 0 | 0 |
| Total | 41 | 1 | 5 | 6 | |

NP = Not Performed

* Samples were collected from 17 sampling events that occurred between May 1990 and August 1997

**Table 2.3
Historical Sediment Toxicity Results
Finfeather Lake**

| Finfeather Lake 1209A | | % Survival | | Sub-Lethal Effect | |
|-----------------------|---------|---------------------|--------------------|----------------------------|-------------------------------|
| | | Pimephales Promelas | Ceriodaphnia dubia | Growth Pimephales Promelas | # Neonates Ceriodaphnia dubia |
| August 19, 1997 | Control | 100 | 100 | | 18.5 |
| | 11798 | 93 | 100 | | 13.0 |
| August 13, 1996 | Control | 97 | 100 | | 19.9 |
| | 11798 | 97 | 100 | | 11.1 |
| August 7, 1995** | Control | 93 | 100 | | |
| | 11798 | 17 | 0 | | |
| June 27, 1994 | Control | 97 | 100 | | |
| | 11798 | 93 | 0 | | |
| August 2, 1993 | Control | 93 | 100 | | |
| | 11798 | 83 | 0 | | |

Bold - denotes significant toxicity

** Test only lasted one day while the rest were 7 days

Table 2.4
Historical Sediment Toxicity Results
Bryan Municipal Lake

| Bryan Municipal Lake 1209B | | % Survival | | Sub-Lethal Effect | |
|----------------------------|---------|---------------------|--------------------|----------------------------|-------------------------------|
| | | Pimephales Promelas | Ceriodaphnia dubia | Growth Pimephales Promelas | # Neonates Ceriodaphnia dubia |
| August 19 1997 | Control | 100 | 100 | | 18.5 |
| | 11792 | 97 | 100 | | 17.0 |
| August 13, 1996 | Control | 97 | 100 | | 19.9 |
| | 11792 | 93 | 100 | | 12.4 |
| February 20, 1996 | Control | 97 | 100 | | 17.9 |
| | 11792 | 93 | 100 | | 16.1 |
| August 7, 1995 | Control | 93 | 100 | | 13.2 |
| | 11792 | 93 | 100 | | 14.7 |
| June 27, 1994 | Control | 97 | 100 | | 17.4 |
| | 11792 | 100 | 100 | | 16.7 |
| August 2, 1993 | Control | 93 | 100 | | 19.0 |
| | 11792 | 87 | 80 | | 8.0 |

Bold - denotes significant difference from the control

Table 2.5
Finfeather Lake
Historical Sediment Chemistry Detections

| PARAMETER | Historical Average* | Historical Minimum* | Historical Maximum* | Lowest Screening Value** | UNITS |
|---|---------------------|---------------------|---------------------|--------------------------|-------|
| 1,2-Dibromoethan Sediment, Dry Weight (µg/KG) | 650 | ND | 1300 | | µg/KG |
| Aluminum in Bottom Deposits (mg/KG as AL Dry Wgt) | 20567 | 13800 | 36800 | | mg/KG |
| Arsenic in Bottom Deposits (mg/KG as AS Dry Wgt) | 222 | 91.2 | 441 | 7.24 | mg/KG |
| Barium in Bottom Deposits (mg/KG as BA Dry wgt) | 269 | 182 | 429 | | mg/KG |
| Cadmium, Total in Bottom Deposits (mg/Kg, Dry Wgt) | 0.5 | 0.2 | 0.936 | 0.676 | mg/KG |
| Chlordane (Tech Mix&Metabs) Sed, Dry Wgt, µg/KG | 19.5 | ND | 39.0 | | µg/KG |
| Chromium, Total in Bottom Deposits (mg/KG, Dry Wgt) | 78.6 | 26.4 | 144 | 52.3 | mg/KG |
| Copper in Bottom Deposits (mg/KG as CU Dry Wgt) | 160 | 32.6 | 276 | 18.7 | mg/KG |
| Lead in Bottom Deposits (mg/KG as PB Dry Wgt) | 47.8 | 24.0 | 73.2 | 30.24 | mg/KG |
| Manganese in Bottom Deposits (mg/KG as MN Dry Wgt) | 253.2 | 134.0 | 394 | | mg/KG |
| Nickel, Total in Bottom Deposits (mg/KG, Dry Wgt) | 16.5 | 10.0 | 21.6 | 15.9 | mg/KG |
| Nitrogen Kjeldahl Total Bottom Dep. Dry Wt mg/KG | 4769 | 3950 | 5256 | | mg/KG |
| Phosphorus, Total, Bottom Deposit (mg/KG Dry Wgt) | 1118 | 974 | 1360 | | mg/KG |
| Sediment Prctl. Size Class, 0.0039 Clay % Dry Wt | 29 | 7 | 63 | | % |
| Sediment Prctl. Size, Sand .0625-2mm % Dry Wt | 14 | 0 | 46 | | % |
| Sediment Prctl. Size Class.0039.0625 Silt % Dry Wt | 57 | 18 | 93 | | % |
| Selenium in Bottom Deposits (mg/KG as SE Dry Wt) | 0.7 | ND | 2.4 | | mg/KG |
| Silver in Bottom Deposits (mg/KG as AG Dry Wgt) | 0.1 | ND | 0.7 | | mg/KG |
| Solids in Sediment, Percent by Weight (Dry) | 24.8 | 20.1 | 32.8 | | % |
| Total Organic Carbon in Sediment Dry Wgt (mg/KG) | 48500 | 23300 | 125200 | | mg/KG |
| Zinc in Bottom Deposits (mg/KG as ZN Dry Wgt) | 415 | ND | 966 | 124 | mg/KG |

Notes:

* TCEQ database information for Station 11798 of Finfeather Lake for the period of March 1995 to August 1997.

** Criteria is from *Equilibrium and Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices* tables. The value is the lowest value of Tier 1 indices based on an aquatic chronic toxicity data set and Tier 2 indices based on draft EPA secondary chronic values (Appendix).

Table 2.6
Bryan Municipal Lake
Historical Sediment Chemistry Detections

| PARAMETER | Historical Average* | Historical Minimum* | Historical Maximum* | Lowest Screening Value** |
|---|---------------------|---------------------|---------------------|--------------------------|
| Aluminum in Bottom Deposits (mg/KG as AL Dry Wgt) | 21310 | 9060 | 38000 | |
| Arsenic in Bottom Deposits (mg/KG as AS Dry Wgt) | 129 | 37 | 395 | 7.24 |
| Barium in Bottom Deposits (mg/KG as BA Dry wgt) | 219 | 106 | 374 | |
| Cadmium, Total in Bottom Deposits (mg/Kg, Dry Wgt) | 0.5 | 0.2 | 0.8 | 0.68 |
| Chlordane (Tech Mix&Metabs) Sed, Dry Wgt, µg/KG | 290 | 290 | 290 | |
| Chromium, Total in Bottom Deposits (mg/KG, Dry Wgt) | 43 | 10 | 132 | 52.3 |
| Copper in Bottom Deposits (mg/KG as CU Dry Wgt) | 71.6 | 10.7 | 267 | 18.7 |
| Lead in Bottom Deposits (mg/KG as PB Dry Wgt) | 52 | 22 | 65 | 30.24 |
| Manganese in Bottom Deposits (mg/KG as MN Dry Wgt) | 200 | 135 | 391 | |
| Nickel, Total in Bottom Deposits (mg/KG, Dry Wgt) | 11.4 | 5.2 | 20.7 | 15.9 |
| Nitrogen Kjeldahl Total Bottom Dep. Dry Wt mg/KG | 3443 | 2980 | 3700 | |
| Phosphorus, Total, Bottom Deposit (mg/KG Dry Wgt) | 822 | 668 | 1070 | |
| Pyrene Dry wgtbotµg/KG | 3873 | ND | 7710 | |
| Sediment Prctl. Size Class, 0.0039 Clay % Dry Wt | 22 | 6 | 54 | |
| Sediment Prctl. Size, Sand .0625-2mm % Dry Wt | 26 | 7 | 77 | |
| Sediment Prctl. Size Class.0039.0625 Silt % Dry Wt | 52 | 10 | 87 | |
| Selenium in Bottom Deposits (mg/KG as SE Dry Wt) | 0.8 | ND | 2.8 | |
| Silver in Bottom Deposits (mg/KG as AG Dry Wgt) | 0.2 | ND | 0.9 | |
| Solids in Sediment, Percent by Weight (Dry) | 31 | 25 | 48 | |
| Total Organic Carbon in Sediment Dry Wgt (mg/KG) | 40958 | 19500 | 88500 | |
| Zinc in Bottom Deposits (mg/KG as ZN Dry Wgt) | 107 | ND | 223 | 124 |

Notes:

* TCEQ database information for Station 11792 of Finfeather Lake for the period of February 1995 to July 1999.

** Criteria is from *Equilibrium and Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices* tables.

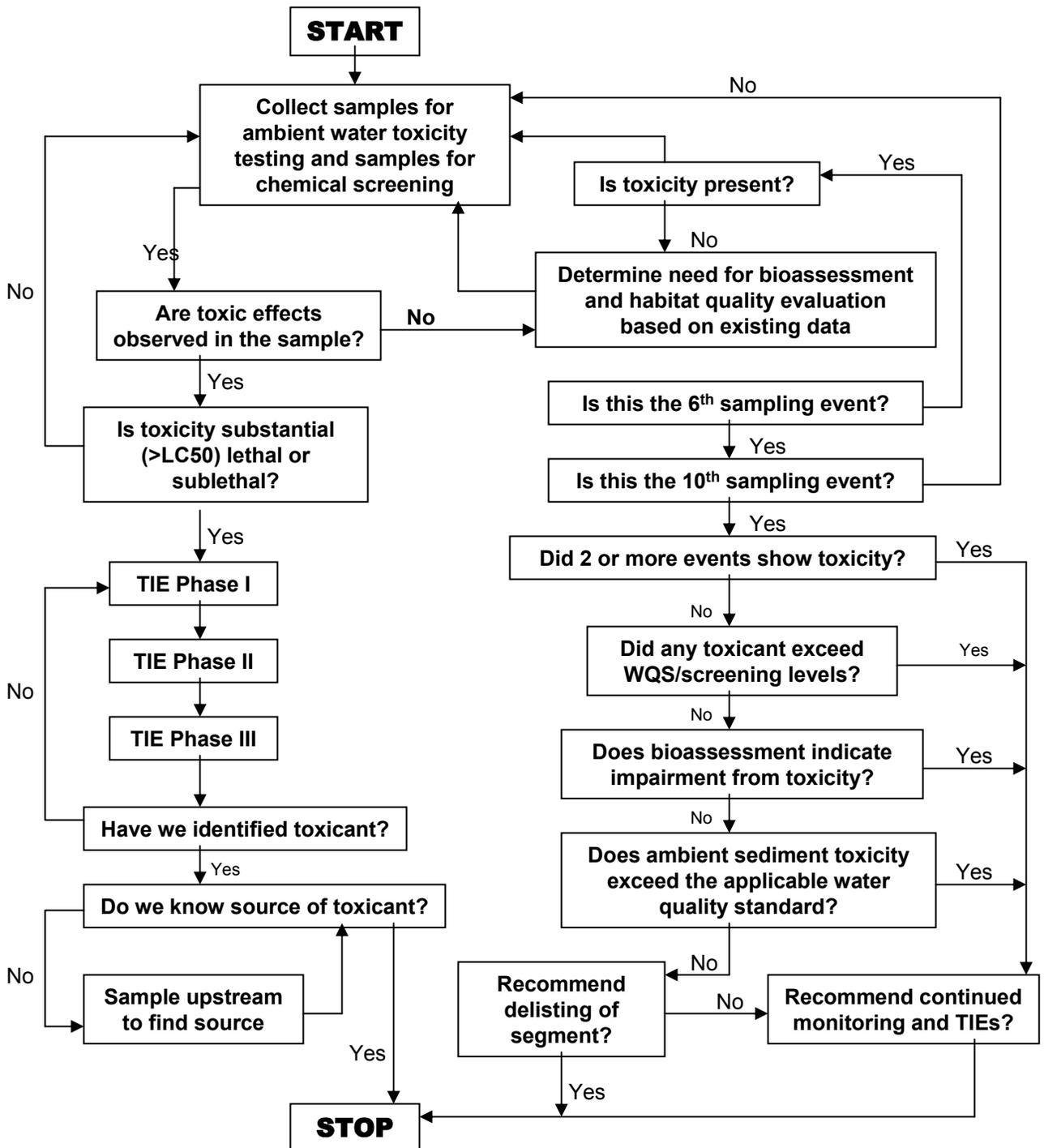
The value is the lowest value of Tier 1 indices based on an aquatic chronic toxicity data set and Tier 2 indices based on draft EPA secondary chronic values (Appendix).

In Bryan Municipal Lake, arsenic, cadmium, chromium, copper, lead, mercury, nickel, and zinc also have exceedances of the screening values. Again, arsenic has the highest percent exceedance of any screening value. Appendix A contains the complete database of all the historical chemical data.

SECTION 3 ASSESSMENT STRATEGY AND OBJECTIVES

The objective of this assessment is one part of the larger objective of establishing fully supported designated uses for the waterbody. The assessment seeks to determine the presence and causes of ambient sediment toxicity. Figure 3 provides a conceptual toxicity strategy flow diagram for this Assessment Study.

Figure 3 Conceptual Toxicity Strategy Flow Diagram



SECTION 4 ASSESSMENT METHODS

4.1 STUDY DESIGN

The general approach used in this assessment is a two-step investigative process. The first step involves determining if an impairment of the designated uses continues. Delisting of the waterbody from the 303(d) list would be pursued if monitoring results demonstrate that the waterbody is no longer impaired. Second, if toxicity is found to be present, a Toxicity Identification Evaluation (TIE) will be performed to identify the toxicant or toxicants causing the impairment. Based on results of the TIE, attempts will be made to identify the source(s) of the toxicity.

4.2 SAMPLING METHODS

Field measurements and sediment samples were collected from three stations in Finfeather Lake and Bryan Municipal Lake during seven sampling events starting in April 2001 and ending in August 2002. Tables 4.1 and 4.2 identify the stations on the two lakes that were sampled, sampling frequencies, toxicity tests conducted and chemical parameters analyzed.

Field staff of Parsons were instructed to follow the field sampling procedures for field, habitat, toxicity, conventional, and chemical parameters documented in the TCEQ *Surface Water Quality Monitoring Procedures Manual* (TCEQ, 1999a) and the TCEQ *Receiving Water Assessment Procedures Manual* (TCEQ, 1999b). Additional procedures for field sampling outlined in this section reflect specific requirements for sampling under this TMDL Project and/or provide additional clarifications in accordance with the approved QAPP.

4.3 GENERAL WATER CHEMISTRY

Four general water chemistry parameters were routinely analyzed during sample collections. Temperature, pH, dissolved oxygen, and specific conductivity were measured with a YSI 600 XL Multi-Parameter Probe. These parameters were measured when sediment samples were collected from a sample location.

4.4 SUMMARY OF FIELD NOTES FOR EACH SAMPLING EVENT

4.4.1 Sampling on April 19, 2001

The crew arrived at Bryan Municipal Lake at 12:30 PM. They collected YSI field measurement and sediment samples at Stations 11792, 11793, and 11794, in that order. They then moved to Finfeather Lake.

The crew arrived at Finfeather Lake at 2:50 PM. They collected YSI field measurement and sediment samples at Stations 11798, 11799, and 11800, in that order.

Table 4.1

Summary of Water and Sediment Sampling Events in Finfeather Lake, Segment 1209B

| ANALYSES | April 19, 2001 | | May 21, 2001 | | June 5, 2001 | | July 18, 2001 | | August 7, 2001 | | Sub-Total |
|----------|----------------|----------|--------------|----------|--------------|----------|---------------|----------|----------------|-------|-----------|
| | Stations | Stations | Stations | Stations | Stations | Stations | Stations | Stations | Stations | | |
| | 11798 | 11799 | 11798 | 11799 | 11798 | 11799 | 11798 | 11799 | 11798 | 11799 | 11800 |

| | | | | | | | | | | | |
|--|---|---|---|---|---|---|---|---|---|---|----|
| Field-measured parameters | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 12 |
| Temperature, DO, pH, conductivity | | | | | | | | | | | |
| SEDIMENT TOXICITY EVALUATION | | | | | | | | | | | |
| Chronic toxicity bioassays | | | | | | | | | | | |
| <i>C. tentans</i> | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 10 |
| <i>H. azteca</i> | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 10 |
| Total metals | | | | | | | | | | | |
| As, Cd, Cr, Cu, Pb, Hg, Ni, Se, Ag, Zn | | | | | 1 | | 1 | | 1 | | 3 |
| VOCs | | | | | | | | | | | |
| Includes priority pollutant list | | | | | 1 | | 1 | | 1 | | 3 |
| SVOCs | | | | | | | | | | | |
| Includes priority pollutant list | | | | | 1 | | 1 | | 1 | | 3 |
| PCBs | | | | | 1 | | 1 | | 1 | | 3 |
| Pesticides/Herbicides including modern compounds | | | | | 1 | | 1 | | 1 | | 3 |
| Bioavailability evaluation | | | | | | | | | | | |
| TOC, AVS, SEM | | | | | 1 | | 1 | | 1 | | 3 |
| Grain-size evaluation | | | | | 1 | | 1 | | 1 | | 3 |
| Percent sand, silt, clay | | | | | | | | | | | |

4.4.2 Sampling on May 21, 2001

The crew arrived at Bryan Municipal Lake, Station 11792 at 11:00 AM. It was partly cloudy, with an air temperature of 72 °F, and winds 10 to 15 mph. They collected YSI readings and sediment samples for both toxicity and chemistry at Stations 11792, 11793, and 11794, in that order. They reached the shore and packed samples at 1:30 PM.

The crew arrived at Finfeather Lake, Station 11798 at 2:20 PM. They collected YSI readings and sediment samples at Stations 11798, 11799, and 11800, in that order. They reached the shore and packed samples at 4:15 PM.

4.4.3 Sampling on June 5, 2001

The crew first arrived at Bryan Municipal Lake, Station 11792 at 10:40 AM. They collected YSI readings and GPS coordinates at Stations 11792, 11793, and 11794, in that order. Sediment samples were only collected at Station 11793. They reached the shore and packed samples at 12:05 PM.

The crew arrived at Finfeather Lake, Station 11798 at 2:20 PM. They collected YSI readings, and GPS coordinates at Stations 11798, 11799, and 11800, in that order. Sediment samples were only collected at Station 11798. The weather was partly cloudy, winds 10 to 15 mph, and 85°F.

Sediment samples from both Finfeather and Bryan Municipal Lakes were also sent to USEPA. USEPA performed an elutriate toxicity test on each sample.

4.4.4 Sampling on July 18, 2001

The crew arrived a Bryan Municipal Lake, Station 11793 at 12:15 PM. They collected YSI readings and sediment samples. A composite sediment sample was created for both toxicity and chemical analysis.

The crew arrived at Finfeather Lake, Station 11798 at 2:30 PM. They collected sediment. The crew recorded the YSI measurement. They noted the water in Finfeather Lake was teeming with blue-green algae and the water's color was very green. The crew then proceeded to Station 11799. YSI readings were recorded. Sediment samples for toxicity and chemistry were collected.

4.4.5 Sampling on August 7, 2001

The crew arrived at Bryan Municipal Lake, Station 11793 at 11:15 AM. YSI readings were recorded. The air temperature was 80°F with a slight wind from the northwest, and was cloudy with a 40 percent chance of rain. YSI measurements were recorded. Sediment samples for both toxicity and chemistry were collected and a composite sample created. The water color was brown to brown-green with small white and green particles in the water column. The sediment was brown with an odor and contained some cattail pieces.

The crew arrived at Finfeather Lake, Station 11798 at 12:45 PM. The air temperature was 94°F with a southeast wind of 3 mph and partly cloudy. The water color was green-brown, turbid, with dark gray particles floating. YSI measurements were recorded and sediment samples collected. The sediment appeared to be brown-black clay with detritus gray particles floating in the sediment bucket.

Sediment samples from both Finfeather and Bryan Municipal Lakes were also sent to USEPA. USEPA performed an elutriate toxicity test on each sample.

4.4.6 Sampling on October 30, 2001

The sampling crew arrived at Bryan Municipal Lake, station 11793 at 1008. After calibrating YSI in Austin at 0940, readings were taken at 11793. One 3.5 gallon bucket of sediment sample 11793-7 was collected for toxicity analysis. Depart Bryan Municipal Lake and arrive at Finfeather Lake station 11798 at 1030. YSI readings were taken and recorded. Sediment sample 11798-7 was collected and for toxicity analysis at 1050. These samples were packaged and shipped to UNT Lab for analysis at 1420.

4.4.7 Sampling on February 7, 2002

The sampling crew arrived at station 11798 on Finfeather Lake at 1120. After calibration, water quality measurements were taken and recorded with the YSI. Sediment sample 11798-8, 4 gallons sediment were collected at 1130. This sample was packed on ice and shipped to UNT via Fed Ex for toxicity analysis.

4.4.8 Sampling on May 9, 2002

The sampling crew arrived at Finfeather Lake at 1227. After YSI calibration, water quality measurements were taken and recorded with the YSI. Sediment sample 11798 was collected at 12:49. This sample was packed on ice and shipped to UNT via Fed Ex for toxicity analysis.

4.4.9 Sampling on July 12, 2002

The sampling crew arrived at Bryan Municipal Lake at 12:35. After YSI was calibrated, water quality measurements were taken and recorded. Sediment samples were collected at station 11793. The crew arrived at station 11794 at 14:50. Sediment sample 11794 was collected. This sample was packed on ice and shipped to UNT via Fed Ex for toxicity analysis.

4.4.10 Sampling on August 5, 2002

The sampling crew arrived at Finfeather Lake at 10:45. After calibrating YSI meter, the water quality measurements were taken. Sediment samples at station 11798 were collected. This sample was packed on ice and shipped to UNT via Fed Ex for toxicity analysis.

4.4.11 Sampling on October 8, 2002

The sampling crew arrived at FFL Station 11798 at 11:40. Sediment samples at Station 11798 were collected at 12:00. A total of two 3.5 gallon buckets were collect. The sample was packed on ice and shipped to UNT via Fed Ex at 15:00 for toxicity analysis.

4.4.12 Sampling on October 30, 2002

The sampling crew arrived at FFL at 13:00. Samples were collected approximately 20 yards towards the outfall (east) in order to collect “fresh” sediment. Sediment was collected at 15:30. Crew shipped sediment sample to UNT via Fed Ex on the morning of October 31, 2002.

4.4.13 Toxicity Testing Method

The toxicity of sediment was assessed by the following methods using the freshwater species *Chironomus tentans* (*C. tentans*) and *Hyallela azteca* (*H. azteca*). *Methods for Measuring the Toxicity and Bioaccumulation of Sediment-associated Contaminants with Freshwater Invertebrates*. Second Edition. USEPA-600-R-99-064, March 2000

For toxicity testing, freshwater midge, *C. tentans* and scud, *H. azteca* were exposed for 10-days to sediment collected from the three previously described stations. Mortality at the end of the 10-day exposure period was statistically compared to mortality found in control exposures where the organisms were exposed to clean sediments supplied by the testing laboratory.

Whereas USEPA approved methods have been developed to identify causes of toxicity in effluents and ambient water, approved methods are not yet available for performing TIEs on sediments. In recent years, considerable progress has been made by USEPA and other research entities to develop TIE methods for sediments. The sediment TIE methods used in this investigation were developed through the coordinated efforts of scientists at USEPA’s laboratory in Duluth, Minnesota, scientist at TRAC Laboratories, scientist at North Texas State University, and Parsons using the most recent scientific advances in the subject area.

Field measurements and sediment samples were collected from Stations 11798, 11799 and 11800 in Finfeather Lake (Segment 1209B) and Stations 11792, 11793, and 11794 in Bryan Municipal Lake during nine sampling events starting in April 2001 and ending in August, 2002. Tables 4-1 and 4.2 identifies the stations that were sampled, sampling frequencies, toxicity tests conducted and chemical parameters analyzed.

Field staff of Parsons was instructed to follow the field sampling procedures for field, habitat, toxicity, conventional, and chemical parameters documented in the TCEQ *Surface Water Quality Monitoring Procedures Manual* (TCEQ, 1999a) and the TCEQ *Receiving Water Assessment Procedures Manual* (TCEQ, 1999b).

4.5 ANALYTICAL METHODS

Appendix F lists a combination of the analytical methods used and potential methods for potential toxicant identification. The analyses listed in Appendix F are USEPA approved methods as cited in TCEQ TMDL guidance document, CRP or SWQM Program Guidance and in 40 Code of Federal Regulations, Section 136, Part B. Exceptions to this include analyses and sample matrices for which no regulated methods exist, or where USEPA has not approved any method with adequate sensitivity for TMDL data requirements.

4.6 QUALITY CONTROL REQUIREMENTS

Refer to the Assessment of the Presence and Causes of Ambient Toxicity Quality Assurance Project Plan (QAPP), Revision 4, FY 2002-03.

4.6.1 Sampling Quality Control Requirements and Acceptability Criteria

The minimum field quality control (QC) requirements followed by Parsons are outlined in the TCEQ *Surface Water Quality Monitoring Procedures Manual* and in Section B5 of the project QAPP. Sampling QC involved use of field duplicates and matrix spikes and matrix spike duplicates.

4.6.2 Laboratory Measurement Quality Control Requirements and Acceptability Criteria

These requirements and criteria are applicable to all laboratories used for analysis of various required parameters. Detailed laboratory QC requirements are contained within each individual method and Laboratory Quality Assurance Manuals. As described in Section B5 of the project QAPP, the minimum requirements followed by analytical laboratories included: 1) laboratory duplicates; 2) laboratory control standards (LCS); 3) matrix spikes (MS) and matrix spike duplicates (MSD); 4) method blanks; and 5) additional QC samples such as surrogates, internal standards, continuing calibration samples, and interference check samples. Laboratory QC sample results are reported with the data report (see Section C2 of the project QAPP).

4.6.3 Failures in Quality Control Requirements

As described in Section B5 of the project QAPP, sampling QC excursions were evaluated by the Parsons Project Manager, in consultation with the Parsons QAO. Differences in field duplicate sample results are used to assess the entire sampling process, including environmental variability. The arbitrary rejection of results based on pre-determined limits was not practical, therefore, the professional judgment of the Parsons Project Manager and QAO was relied upon when evaluating results. Rejecting sample results based on wide variability was a possibility. Corrective action included identification of the cause of the failure where possible. Response actions typically included re-analysis of questionable samples. In some cases, a site was re-sampled to achieve project goals. The disposition of such failures and conveyance to the TCEQ are discussed in Section B4 of the

project QAPP under Failures or Deviations in Analytical Methods Requirements and Corrective Actions.

Refer to Appendix F for the summarization of QA/QC findings, data acceptability and qualifiers to deviations.

4.7 DATA MANAGEMENT

Data Management Protocols are addressed in the Data Management Plan which is Appendix E of the project QAPP.

4.8 STREAM HABITAT CHARACTERIZATION

Stream habitat characterization utilizing TCEQ procedures was performed during the April 2001 sampling event by completing copies of the TCEQ's receiving water assessment forms (Stream Physical Characteristics Worksheets) for each location. The detailed Habitat forms are located in Appendix H.

SECTION 5 RESULTS OF AMBIENT SEDIMENT ANALYSIS

5.1 FIELD MEASUREMENTS RESULTS

Tables 5.1 and 5.2 presents the results for Finfeather Lake and Bryan Municipal Lake, respectively. The dissolved oxygen measurement of 0.34 mg/l and the total residual chlorine measurement of 0.25 mg/l at Station 11798 on August 7, 2001 were unexpected. The chlorine could be associated with the power plant discharge. A total residual chlorine measurement of 0.3 mg/l was also measured in Bryan Municipal Lake on the same day. A low chlorine residual should not affect sediment toxicity.

5.2 AMBIENT SEDIMENT TOXICITY RESULTS

Sediment toxicity was evaluated by a 10-day sediment exposure test with the fresh water species *C. tentans* and *H. azteca* using methods specified in Section 4.5 of this report. Criteria for determining whether significant sediment toxicity has occurred to *C. tentans* and *H. azteca* are specified in the Technical Memorandum in Appendix G to this report. The following conditions must each be met for a sediment to be considered toxic:

- There is a statistically significant reduction in survival, at alpha equal to 0.05,
- Survival in the sample is at least 20 percentage points less than the survival in the control, and
- Survival in the control must be greater than 70% for *C. tentans* and 80% for *H. azteca* for the test to be valid.

Similar conditions to these have been utilized by the TCEQ previously in the TPDES permit requirements for conditions that trigger a TIE/TRE. These conditions assure that a sample is ecologically significant and some quantifiable amount of increased survival may be observed in conducting a TIE.

For Bryan Municipal Lake and Finfeather Lake, nine sampling events for each waterbody were scheduled for sediment toxicity testing at the three identified stations utilizing *C. tentans* and *H. azteca*. Tables 5.3 and 5.4 present a summary of the test results. Section 6 provides toxicity test results obtained during the TIE.

Toxicity tests performed on sediment samples from the first event, April 19, 2001, demonstrated significant toxicity due to lethality of *C. tentans* at Stations 11798 and 11800 in Finfeather Lake and Station 11793 in Bryan Municipal Lake. Statistically significant sublethal effects were also observed at Stations 11792 and 11794 for Bryan Municipal Lake. Significant toxicity due to lethality of *Hyalella* occurred at Station 11798 in Finfeather Lake. Statistically significant sublethal effects were observed at Station 11800 in Finfeather Lake and Stations 11792 and 11793 in Bryan Municipal Lake. Due to the toxicity of the sediments, a TIE was initiated for both Finfeather Lake and Bryan Municipal Lake. Phase I of the TIE has been initiated at Station 11798 in Finfeather Lake and Station 11793 in Bryan Municipal Lake.

**Table 5.1
Field Measurements
Finfeather Lake**

| Water Quality Measurements Finfeather Lake - Segment 1209B | | | | | |
|---|--|-------------------------------|-----------|-----------------------------|---------------------------|
| Station 11798 | | | | | |
| Date M/D/Y | Temp °C | DO Conc mg/L | pH | Cond uS/cm | TRC mg/l |
| 4/19/2001 | 21.6 | 6.43 | 7.7 | 720 | NR |
| 5/21/2001 | 28.1 | 6.23 | 6.56 | 680 | NR |
| 6/5/2001 | 28.7 | 5.32 | 6.65 | 645 | NR |
| 7/18/2001 | YSI Suspected to be out of Calibration | | | | NR |
| 8/7/2001 | 30.41 | 0.34 | 8.96 | 604 | 0.25 |
| 10/30/2001 | 19.97 | 8.11 | 8.04 | 369 | NR |
| 2/7/2002 | 10.8 | 8.96 | 8.13 | 436 | NR |
| 5/9/2002 | 26.8 | 9.12 | 8 | 368 | NR |
| 8/5/2002 | 31.65 | 7.17 | 8.86 | 349 | NR |

| Station 11799 | | | | | |
|-----------------------------|--|-------------------------------|-----------|-----------------------------|---------------------------|
| Date M/D/Y | Temp °C | DO Conc mg/L | pH | Cond uS/cm | TRC mg/l |
| 4/19/2001 | 21.5 | 6.76 | 7.57 | 720 | NR |
| 5/21/2001 | 27.8 | 6.99 | 6.8 | 650 | NR |
| 6/5/2001 | 28.8 | 5.76 | 7.75 | 533 | NR |
| 7/18/2001 | YSI Suspected to be out of Calibration | | | | NR |

| Station 11800 | | | | | |
|-----------------------------|--------------------------|-------------------------------|-----------|-----------------------------|---------------------------|
| Date M/D/Y | Temp °C | DO Conc mg/L | pH | Cond uS/cm | TRC mg/l |
| 4/19/2001 | 21.5 | 7.00 | 7.88 | 700 | NR |
| 5/21/2001 | 27.7 | 6.27 | 6.77 | 680 | NR |
| 6/5/2001 | 28.9 | 5.95 | 6.1 | 578 | NR |
| 5/9/2002 | 27.7 | 12.4 | 8.38 | 369 | NR |

NR - Not Reported

°C - degrees Celcius

mg/L - milligrams per liter

uS/cm - micro Siemens per centimeter

ft - feet

pH is in standard units

Cond - Conductivity

DO Conc - Dissolved oxygen concentration

**Table 5.2
Field Measurements
Bryan Municipal Lake**

| Station 11792 | | | | | |
|-----------------------------|--------------------------|-------------------------------|-----------|-----------------------------|---------------------------|
| Date M/D/Y | Temp °C | DO Conc mg/L | pH | Cond uS/cm | TRC mg/l |
| 4/19/2001 | 20.5 | 5.05 | 7.99 | 750 | NR |
| 5/21/2001 | 28 | 6.55 | 6.98 | 670 | NR |
| 6/5/2001 | 27.8 | 5.35 | 6.01 | 668 | NR |

| Station 11793 | | | | | |
|-----------------------------|--------------------------|-------------------------------|-----------|-----------------------------|---------------------------|
| Date M/D/Y | Temp °C | DO Conc mg/L | pH | Cond uS/cm | TRC mg/l |
| 4/19/2001 | 21.2 | 5.8 | 7.85 | 760 | NR |
| 5/21/2001 | 27.5 | 6.22 | 7.29 | 550 | NR |
| 6/5/2001 | 28 | 5.93 | 6.35 | 565 | NR |
| 7/18/2001 | 31.55 | 7.07 | 8.52 | 431 | NR |
| 8/7/2001 | 30.35 | 0.26 | 7.78 | 585 | 0.3 |
| 10/30/2001 | 18.56 | 9.18 | 7.79 | 296 | NR |
| 7/16/2002 | 33.42 | 7.49 | 8.02 | 749 | NR |

| Station 11794 | | | | | |
|-----------------------------|--------------------------|-------------------------------|-----------|-----------------------------|---------------------------|
| Date M/D/Y | Temp °C | DO Conc mg/L | pH | Cond uS/cm | TRC mg/l |
| 4/19/2001 | 21.3 | 5.84 | 7.75 | 860 | NR |
| 5/21/2001 | 27.9 | 6.33 | 6.91 | 690 | NR |
| 6/5/2001 | 28 | 5.55 | 6.55 | 629 | NR |
| 7/16/2002 | 32.31 | 5.18 | 8.08 | 546 | NR |

NR - Not Reported

°C - degrees Celcius

mg/L - milligrams per liter

uS/cm - micro Siemens per centimeter

ft - feet

pH is in standard units

Cond - Conductivity

DO Conc - Dissolved oxygen concentration

Table 5.3 Ambient Sediment Toxicity Results for Finfeather Lake

FinFeather Lake 1209B

10 day Sediment Survival and Growth Results Summary

April 19, 2001

| | | % Survival | | Sub-Lethal Effect Growth | |
|-----------------|---------------------|------------------|----------------|--------------------------|----------------|
| | | Chironmus tetans | Hyaella azteca | Chironmus tetans | Hyaella azteca |
| Finfeather Lake | Control | 74 | 91 | 0.082 | 0.112 |
| | 4/19/2001 11798 | 33 | 54 | NA | NA |
| | 4/19/2001 11799 | 58 | 86 | 0.066 | 0.107 |
| | 4/19/2001 11800 | 46 | 89 | NA | 0.094 |
| | 4/19/2001 11800-Dup | 49 | 88 | NA | 0.071 |

May 21, 2001

| | | % Survival | | Sub-Lethal Effect Growth | |
|-----------------|-----------------|------------------|----------------|--------------------------|----------------|
| | | Chironmus tetans | Hyaella azteca | Chironmus tetans | Hyaella azteca |
| Finfeather Lake | Control | 75 | 99 | 0.269 | 0.167 |
| | 5/21/2001 11798 | 23 | 84 | NA | 0.091 |
| | 5/21/2001 11799 | 79 | 80 | 0.238 | 0.146 |
| | 5/21/2001 11800 | 69 | 56 | 0.303 | 0.112 |

Bold - denotes significant difference from the control

Table 5.4 Ambient Sediment Toxicity Results for Bryan Municipal Lake

Bryan Municipal Lake 1209A
10 day Sediment Survival and Growth Results Summary
April 19, 2001

| | | % Survival | | Sub-Lethal Effect Growth | |
|---|---------|------------------|-----------------|--------------------------|-----------------|
| | | Chironmus tetans | Hyalella azteca | Chironmus tetans | Hyalella azteca |
| Bryan Municipal Lake 4/19/2001 4/19/2001 4/19/2001 | Control | 81 | 91 | 0.706 | 0.112 |
| | 11792 | 69 | 79 | 0.455 | 0.091 |
| | 11793 | 31 | 84 | NA | 0.086 |
| | 11794 | 71 | 85 | 0.367 | 0.115 |

May 21, 2001

| | | % Survival | | Sub-Lethal Effect Growth | |
|---|---------|------------------|-----------------|--------------------------|-----------------|
| | | Chironmus tetans | Hyalella azteca | Chironmus tetans | Hyalella azteca |
| Bryan Municipal Lake 5/21/2001 5/21/2001 5/21/2001 | Control | 75 | 99 | 0.269 | 0.167 |
| | 11792 | 66 | 84 * | 0.378 | 0.128 |
| | 11793 | 70 | 92 | NA | 0.109 |
| | 11794 | 84 | 90 * | 0.388 | 0.145 |

Bold - denotes significant difference from the control

* Note that while statistically significant mortality effects were observed, H. azteca survival was 83.8% for 11792 test #2 and 90% for 11794 test #2.

Toxicity tests performed on sediment samples from the second event, May 21, 2001, demonstrated significant lethality to *C. tentans* at Station 11798 in Finfeather Lake. Statistically significant sublethal effects to *C. tentans* were also observed at Stations 11792 and 11794 for Bryan Municipal Lake. Significant toxicity due to lethality of *H. azteca* occurred at Station 11800 in Finfeather Lake.

Statistically significant sublethal effects on *H. azteca* were observed at Station 11798 in Finfeather Lake and Station 11793 in Bryan Municipal Lake.

Toxicity was similar between the two organisms tested, with *C. tentans* showing slightly more sensitivity, in general, to the toxicant(s) than *H. azteca*. In addition, the effects were slightly greater in Finfeather Lake than in Bryan Municipal Lake. Due to the organisms responses, it would appear that the same toxicant(s) is present in both lakes and that the concentrations are higher in Finfeather Lake than Bryan Municipal Lake. Note: Three out of four sublethal effects tests for *C. tentans* growth in the control were below the minimum control growth of 0.48 mg AFDW.

5.3 SEDIMENT CHEMICAL ANALYSIS RESULTS

Tables 5.5 and 5.6 presents only detected concentrations of parameters found in samples taken from Stations 11798, 11799 and 11800 for Finfeather Lake and Stations 11792, 11793 and 11794 in Bryan Municipal Lake in the three sampling events. Detections of arsenic, copper, and zinc at stations of Finfeather Lake were consistently above the corresponding lowest screening levels. Pesticides, consisting of DDD, DDE, and DDT were detected but were at too low a concentration to quantify. The lowest screening levels for these pesticides were below the minimum analytical level for USEPA method 8081.

Detections of arsenic, copper, lead, and zinc at stations of Bryan Municipal Lake were consistently above the corresponding lowest screening levels. The pesticide DDD was also detected above the lowest screening level, but was not quantifiable. Appendix E contains the results from all chemical analytes tested.

Tables 5.5 and 5.6 show calculated toxic units greater than 1.0 in whole sediments.

Table 5.5
Finfeather Lake Segment 1209B

| Finfeather Lake PARAMETER | Station ID 11799 | Station ID 11798 | Station ID 11799 | Station ID 11798 | Station ID 11800 | Station ID 11800 | Lowest Screening Values* | UNITS |
|------------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|------------------|-----------------------------|--------------|
| | 5/21/2001 | 7/18/2001 | 7/18/2001 | 5/9/2002 | 5/9/2002 | 5/9/2002 | | |
| Ions | Toxic ¹ | NA | NA | NA | NA | NA | | |
| | 35.3 | 64.5 | 33.7 | 57.5 | 142 | | | mg/Kg-dry wt |
| | 72.6 | 57.5 | 78.9 | 149 | 289 | | | mg/Kg-dry wt |
| Metals | 7870 | 12800 | 4680 | 6600 | 15000 | | | |
| Aluminum | 58.5 (8.08) | 196 (27.1) | 28.8 (4.0) | 79.2 (10.9) | 160 (22.1) | | 7.24 | mg/Kg-dry wt |
| Arsenic | 113 | 641 | 96.6 | 164 | 207 | | | mg/Kg-dry wt |
| Barium | 0.33 | 0.71 (1.05) | ND | 0.454 | 0.96 (1.42) | | 0.676 | mg/Kg-dry wt |
| Cadmium | 10100 | 19500 | 7750 | 8460 | 82500 | | | mg/Kg-dry wt |
| Calcium | 26.3 | 95.4 (1.82) | 16.9 | 34.7 | 46.8 | | 52.3 | mg/Kg-dry wt |
| Chromium | 65.4 (3.5) | 575 (30.7) | 44.5 (2.37) | 171 (9.14) | 113 (6.04) | | 18.7 | mg/Kg-dry wt |
| Copper | 6740 | 14700 | 4220 | 6670 | 13800 | | | mg/Kg-dry wt |
| Iron | 17.5 | 56.9 (1.88) | 12.6 | 33.3 (1.10) | 51.8 (1.71) | | 30.24 | mg/Kg-dry wt |
| Lead | 1340 | 2520 | 924 | 1130 | 4470 | | | mg/Kg-dry wt |
| Magnesium | 6.91 | 76.7 (4.82) | ND | 21.7 (1.36) | ND | | 15.9 | mg/Kg-dry wt |
| Nickel | 652 | 882 | 402 | 464 | 1370 | | | mg/Kg-dry wt |
| Potassium | 462 | 1040 | 333 | 366 | 1030 | | | mg/Kg-dry wt |
| Sodium | 241 (1.94) | 1280 (10.3) | 151 (1.22) | 447 (3.61) | 466 (3.76) | | 124 | mg/Kg-dry wt |
| Zinc | ND | 0.17 | ND | ND | ND | | | µg/Kg-dry wt |
| Semi-Vol | ND | 0.14 | ND | ND | ND | | 27372 | µg/Kg-dry wt |
| Benzo(b)fluoranthene | ND | 0.17 | ND | ND | ND | | 108 | µg/Kg-dry wt |
| Chrysene | ND | 0.16 | ND | ND | ND | | 113 | µg/Kg-dry wt |
| Fluoranthene | ND | 0.16 | ND | ND | ND | | 153 | µg/Kg-dry wt |
| Pyrene | ND | 0.16 | ND | ND | ND | | | µg/Kg-dry wt |
| PARAMETER | 5/21/2001 | 7/18/2001 | 7/18/2001 | 5/9/2002 | 5/9/2002 | | Lowest Screening Values* | UNITS |
| Pest/PCBs | | | | | | | | |
| 4,4'-DDD | 15.0 | J | ND | ND | ND | | 1.2 | µg/Kg-dry wt |
| 4,4'-DDE | 24.0 | J | ND | ND | ND | | 2.1 | µg/Kg-dry wt |
| 4,4'-DDT | 5.2 | J | 3.6 | J | ND | | 1 | µg/Kg-dry wt |
| g-BHC (Lindane) | ND | 3.3 | J | ND | ND | | | µg/Kg-dry wt |
| SEM | | | | | | | | |
| Cadmium | 0.29 | 0.28 | ND | ND | ND | | | µmol/dry g |
| Copper | 2.13 | ND | 1.2 | 4.5 | 3.9 | | | µmol/dry g |
| Lead | 20.68 | 22.00 | 6.7 | 0.11 | 0.34 | | | µmol/dry g |
| Mercury | 0.0006 | J | ND | ND | ND | | | µmol/dry g |
| Nickel | 1.71 | 4.2 | ND | 0.23 | 0.19 | | | µmol/dry g |
| Silver | 0.407 | J | ND | NA | NA | | | µmol/dry g |
| Zinc | 270.9 | 660 | 90 | 7.9 | 12 | | | µmol/dry g |
| Total Organic Carbon (TOC) | 23100 | 26430 | 14710 | 18000 | 51300 | | | mg/Kg |
| Acid Volatile Sulfide (AVS) | 1061 | 260 | 51 | 29.6 | 112 | | | µmol/dry g |

Notes:

* Criteria is from *Equilibrium and Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices* tables. The value is the lowest value from the Indices as stated in the Appendix.

J- result is estimated

ND- result was Not Detected

NA - Not Analyzed

mg/kg-dry = milligrams per kilogram dry weight

µg/kg-dry = microgram per kilogram dry weight

µmol/dry g = micromole per dry gram

¹ No significant difference from control for survival and growth of *C. tentans* in 10 day sediment exposures; significant difference in survival of *H. azteca* in 10 day sediment exposure.

() = Toxic Units; calculated by dividing detected metal concentration by the lowest screening level.

SECTION 6 TOXICITY IDENTIFICATION EVALUATION

Both Finfeather Lake (FFL) and Bryan Municipal Lake (BML) were sampled for a total of six events each during an initial 5-month period from April 2001 to August 2001. Once TIE studies began, routine whole sediment toxicity testing ceased. Detections of aluminum, arsenic, copper, and zinc at Station 11799 of FFL were above the corresponding screening levels. Pesticides, consisting of DDD, DDE, and DDT were also detected but were at concentrations too low to quantify. The lowest sediment screening levels for these pesticides were below the minimum analytical level for EPA method 8081. Detections of aluminum, arsenic, cadmium, chromium, copper, lead, nickel and zinc were above the corresponding screening levels at Station 11798. Pesticide, DDT, was detected but was at a concentration too low to quantify. Detections of aluminum, arsenic, cadmium, copper, lead, and zinc at Station 11793 of BML were also above the corresponding screening levels. The pesticide DDD was detected above the sediment screening level, but was not quantifiable.

Toxicity test results for sediment samples collected in April and May 2001 indicated the sediments were significantly toxic due to lethality at Stations 11798 and 11800 in FFL and Station 11793 in BML to *Chironmus tentans* and *Hyallela azteca* species using whole sediment toxicity test methods. Statistically significant sublethal effects were also observed in sediment collected from Station 11800 in FFL and Stations 11792, 11793, and 11794 in BML. Due to the toxicity of the sediments, a TIE was initiated for both FFL and BML. Phase I of the TIE was initiated at Station 11798 in FFL and Station 11793 in BML. Since it is very likely that the same contaminant is affecting both lakes, it was decided to focus on the most toxic site first (in FFL) for the TIE.

U.S. EPA has not finalized sediment pore water or whole sediment Toxicity Identification Evaluation (TIE) methodology. Draft sediment TIE guidelines are available for pore waters and elutriates (EPA 1991) and closely follow effluent TIE procedures. Some whole sediment procedures for reducing toxicity of specific toxicant classes have been reported in the literature; however, whole sediment TIE procedures are not published in guideline format (Ho et al. 2002). Therefore, a tiered approach based on pore water tests was employed in this project (Ankley and Schubauer-Berigan 1995). Additional whole sediment TIE procedures were performed on Fin Feather Lake. Generally, 40-60% of sediment volume was isolated as pore water. *C. dubia* was chosen for pore water testing because of test volume requirements. *Hyallela azteca* and *Chironomus tentans* were also used to test whole sediments. Table 6.1 provides the TIE toxicity test results.

Table 6.1
C. dubia, 7-Day Toxicity Tests using Sediment Pore Water

| Sample Date | Test Date | Station | Treatment | Results | | |
|-------------|-----------|---------|----------------|--------------|---------|------------|
| | | | | Reproduction | Std Dev | % Survival |
| 6/5/2001 | 7/31/2001 | 11798 | RHW (Control) | 26.4 | 2.19 | |
| | | | Baseline 100% | 15 | 2.55 | |
| 6/5/2001 | 7/31/2001 | 11793 | RHW (Control) | 26.4 | 2.19 | |
| | | | Baseline 100% | 15.2 | 3.11 | |
| 6/5/2001 | 8/25/2001 | 11798 | RHW (Control) | 24.2 | 2.28 | |
| | | | Baseline 100% | 14.6 | 1.82 | |
| 10/30/2001 | 2/2/2002 | 11798 | RMHW (Control) | 23 | 3.16 | 100 |
| | | | Baseline 100% | 11.4 | 0.894 | 100 |
| 10/30/2001 | 6/6/2002 | 11798 | RMHW (Control) | 25.8 | 3.03 | 100 |
| | | | Baseline 100% | 0 | 0 | 20 |
| 10/30/2001 | 6/26/2002 | 11798 | RMHW (Control) | 23.8 | 2.57 | 100 |
| | | | Baseline 100% | 10.6 | 4.9 | 100 |
| 10/30/2002 | 12/12/200 | 11798 | RMHW (Control) | 27.4 | 2.07 | 100 |
| | | | Baseline 100% | 2 | 0.82 | 80 |

All general pore water TIE procedures followed EPA (1991) draft guidelines. Whole sediment TIEs followed procedures previously reported in the peer-reviewed literature. In addition to draft EPA TIE procedures, we used three ion exchange media to remove organic or metal toxicants. The cation exchange resin SIR-300, a styrene and divinylbenzene copolymer with iminodiacetic functional group in the sodium form, was chosen for metal removal because of its ability to chelate heavy metal cations (ResinTech, New Berlin NJ). SIR-300 was previously suggested as an effective metal treatment in sediment TIE procedures (Burgess et al. 2000). SIR-300 affinity for metals is: Hg²⁺>Cu²⁺>V²⁺>Pb²⁺>Ni²⁺>Zn²⁺>Co²⁺>Cd²⁺>Fe²⁺>Be²⁺, Mn²⁺>Mg²⁺, Ca²⁺>Sr²⁺>Ba²⁺>Na²⁺.

Although SIR-300 is a parallel TIE treatment to EDTA for divalent metals, we used SIR-300 in addition to EDTA because metals reduced by SIR-300 may be measured following TIE treatment. Because conventional TIE treatments are not effective for arsenic contaminated media, SIR-900, a synthetic aluminum oxide absorbent media specific for

arsenic (arsenate and arsenite) and lead, was utilized in several TIE procedures for FFL sediment because of historic arsenic contamination (ResinTech, West Berlin NJ). C18 solid phase extraction columns, typically used in TIE procedures to remove organic contaminants, may also filter or remove other contaminants (e.g. metals) and complicate TIE interpretation. Amborsorb 563, a carbonaceous adsorbent, for organic removal was used because it has 5 to 10 times the capacity of granular activated carbon. Amborsorb 563 in addition to C18 treatment was used in several TIEs to selectively remove organics without filtration complications. Amborsorb has been used to treat contaminated groundwater (EPA 1995) and lake water (Guzzella et al. 2002) and to remove organic contaminants in sediment TIE procedures (West et al. 2001). A summary of all TIEs performed on this segment is provided in Table 6.2.

Table 6.2
Sediment Toxicity Identification Evaluation Procedures

| Test Date | Test Type | Station | Organism | Effective Treatment |
|-------------------------|----------------|---------|------------------|---------------------|
| July 13-23, 2001 | Pore Water | 11793 | <i>C. dubia</i> | EDTA |
| July 13-23, 2001 | Pore Water | 11798 | <i>C. dubia</i> | None |
| Aug. 25 – Sept. 4, 2001 | Pore Water | 11798 | <i>C. dubia</i> | SIR300, SIR900 |
| February 2-12, 2002 | Pore Water | 11798 | <i>C. dubia</i> | SIR300, SIR900 |
| March 10-20, 2002 | Whole Sediment | 11798 | <i>H. azteca</i> | SIR900* |
| June 6-16, 2002 | Pore Water | 11798 | <i>C. dubia</i> | EDTA, SIR300 |
| June 6-16, 2002 | Pore Water | 11800 | <i>C. dubia</i> | EDTA |

* *H. azteca* growth not significant different from control sediment.
 60% survival in SIR900, 68.3% survival in control.

In July 2001, it was determined that sediment pore water was not acutely toxic; however, pore water in sediment from stations 11793 and 11798 produced toxic effects on *C. dubia* reproduction. EDTA treatment reduced pore water toxicity at Station 11793, but not at 11798. A subsequent TIE was performed with several treatment media selective for metals which removed toxicity in August 2001. These treatment media included: SIR-900 media selective for arsenic; TXI Shale, previously demonstrated to remove arsenic from aqueous solutions in sorbtion isotherm studies (F. Saleh, UNT, pers. comm.); and SIR-300. Station 11798 pore water was chosen for TIE treatments because of higher ambient sediment metal concentrations than other stations. Each TIE treatment improved *C. dubia* reproduction relative to untreated pore waters. Because arsenic was suspected as a causative toxicant, total

arsenic, arsenate and arsenite levels were quantified in each TIE treatment. SIR-900 and TXI Shale treatments reduced total arsenic, arsenate and arsenite pore water concentrations. Arsenic concentrations in pore waters were not sufficiently elevated to solely cause toxicity; total arsenic pore water concentration was 266 µg/L. Previous investigators found that arsenic treatment up to 1.46 mg/L did not significantly affect *C. dubia* reproduction. Other metals may have been removed by SIR-900 and TXI shale treatments; however, metal concentrations were not measured following these initial treatments.

In February 2002, another TIE was performed on Station 11798 pore waters. Metal bioavailability and toxicity were reduced with increasing water hardness. Unlike previous TIEs, in which reconstituted hard water was used for pore water dilution, reconstituted moderately hard water was used for dilution in this TIE. Dilution water with lower hardness was chosen to maximize pore water metal bioavailability and toxicity, and potentially the effectiveness of TIE treatments. SIR-900 and SIR-900 + SIR-300 treatments significantly increased *C. dubia* reproduction in undiluted and 50% dilution, respectively. SIR-300 increased neonate production at 50% dilution, although not significantly, by an average of four neonates/female relative to baseline pore waters. Baseline pore water copper concentration was 722 µg/L, a value two orders of magnitude higher than previously reported lowest observed effect concentrations for *C. dubia* reproduction in laboratory water and substantially higher than the TSWQS of 48.6 µg/L. The lowest *C. dubia* fecundity was observed in undiluted SIR-300 TIE treatments. This sub-lethal response was attributed to reduction of the essential nutrients calcium and magnesium from pore waters in this test .

Whole sediments from Station 11798 were amended with SIR-300, SIR-900 and SIR-300 + SIR-900 at a 1:4, volume to volume (V:V) ratio in March 2002. Following a 10-day exposure period, *H. azteca* growth was significantly improved, relative to reference sediment from the University of North Texas Water Research Field Station, by only SIR-900 treatments.

TIE procedures conducted in August 2001 and February 2002 determined arsenic concentrations in sediment pore water were not high enough to affect *C. dubia* reproduction. However, copper was measured at greater than 700 µg/L during the February 2002 TIE. SIR-300, the ion exchange resin reported to possess high selectivity for copper, also dramatically reduced calcium and magnesium concentrations in the February study. It was hypothesized that this reduction in Ca and Mg led to lower *C. dubia* fecundity because both metals are essential nutrients. In June 2002, a TIE was performed with Station 11798 sediment pore waters. As with the February study, reconstituted moderately hard water served as dilution water. Following SIR-300 treatment, hardness was measured by titration, and Ca and Mg salts reintroduced to pore waters until hardness values returned to pre-SIR-300 levels. EDTA (3 mg/L) and SIR-300 treatments significantly improved survival and reproduction relative to baseline pore waters. Because multiple metals were measured in pore waters, a toxic unit approach was taken to evaluate metal pore water toxicity (Tables 6.3). Concentrations of zinc, iron, lead and barium decreased 26%, 32%, 37% and 96%, respectively, with SIR-300 treatments.

Bioavailability of these metals was clearly affected by compounds not accounted for in water hardness measures. An example is organic carbon binding to metals which affects bioavailability and toxicity. Total organic carbon in these pore waters was measured at elevated concentrations (baseline, 22.2 mg/L; SIR-300, 14.8 mg/L). An EDTA treatment of 3 mg/L also improved *C. dubia* survival and fecundity in Station 11798 pore waters. Average neonate production was 2x higher in EDTA treatments than in SIR-300 treatments. Although the manufacturer of SIR-300 indicated that this resin is more selective for zinc, iron, lead and copper than calcium and magnesium, our data suggests that this resin preferentially removed calcium and magnesium. If the binding capacity of SIR-300 was exhausted by preferential binding of calcium and magnesium ions, ligands that bound calcium and magnesium in pretreated pore waters would be available for complexing with other divalent metals, specifically those metals measured at high enough concentrations (e.g. copper, to adversely affect *C. dubia* in 'clean' laboratory water toxicity tests).

A subsequent TIE study with SIR-300 was conducted to further remove metal contaminants from Station 11798 pore waters (Tables 6.4 and 6.5). By increasing the V:V ratio of SIR-300 to pore water during TIE treatment, SIR-300 metal binding capacity was increased and total metal pore water concentrations were decreased (refer to Table 6.6). Following reintroduction of calcium and magnesium salts, effective in the June 2002 TIE with this pore water, a reduction in toxicity was more clearly assigned to potentially causative toxicants. In addition, contaminant addition procedures (Phase III TIE) were subsequently performed to recreate pore water toxicity and provide corroborating information. Phase III TIE procedures were conducted such that pore waters were first treated with SIR 300 followed by reintroduction of copper, lead or zinc at nominal concentrations (refer to Table 6.6). Results indicated copper and zinc as the primary concerns. These results, using a toxic units approach suggests copper as a greater concern than zinc; however, 100% mortality was observed in treated pore waters in which zinc had been reintroduced. In addition, Table 6.6 suggests zinc is more of a concern than copper. Refer to the UNT report in Appendix D for more details.

Table 6.3
Dissolved Metals Chemistry and Toxic Units of Finfeather Lake,
Station 11798, Sediment Pore Water Resin TIE
7-Day C. Dubia Test Initiated June 6, 2002

| Treatment | % Survival | Mean # Neonates | Aluminum | Arsenic | Barium ³ | Cadmium | Calcium |
|--|------------|-----------------|-----------------------|-------------|---------------------|---------------|---------|
| 11798 100% Baseline | 20 | 0 | 707 (0.713) | 212 (1.12) | 225 (0.225) | <1.0 (<0.8) | 43500 |
| 11798 100% + SIR 300 ¹ | 100 | 8.6 | 1190 (1.20) | 260 (1.37) | <10 (<0.01) | <1.0 (<0.8) | 1940 |
| RMHW ² + SIR 300 ¹ | 100 | 23.2 | <100 (0.100) | <10 (<0.05) | <10 (<0.01) | <1.0 (<1.45) | 1200 |
| RMHW ² Control | 100 | 25.8 | <100 (0.100) | <10 (<0.05) | <10 (<0.01) | <1.0 (<1.11) | 15400 |
| Treatment | % Survival | Mean # Neonates | Chromium ⁴ | Copper | Iron | Lead | |
| 11798 100% Baseline | 20 | 0 | 11.2 (0.011) | 251 (16.5) | 1410 (1.41) | 24.2 (7.01) | |
| 11798 100% + SIR 300 ¹ | 100 | 8.6 | 15.8 (0.016) | 290 (20.2) | 955 (0.96) | 15.3 (4.81) | |
| RMHW ² + SIR 300 ¹ | 100 | 23.2 | <10 (<0.01) | <10 (<1.26) | <100 (<0.10) | <3.0 (<2.27) | |
| RMHW ² Control | 100 | 25.8 | <10 (<0.01) | <10 (<0.95) | <100 (<0.10) | <3.0 (<1.48) | |
| Treatment | % Survival | Mean # Neonates | Magnesium | Nickel | Potassium | Selenium | |
| 11798 100% Baseline | 20 | 0 | 4910 | <10 (<0.05) | 4060 | <10 (<2.0) | |
| 11798 100% + SIR 300 ¹ | 100 | 8.6 | 1570 | <10 (<0.05) | 4670 | <10 (<2.0) | |
| RMHW ² + SIR 300 ¹ | 100 | 23.2 | 1120 | <10 (<0.10) | 1760 | <10 (<2.0) | |
| RMHW ² Control | 100 | 25.8 | 11000 | <10 (<0.07) | 1840 | <10 (<2.0) | |
| Treatment | % Survival | Mean # Neonates | Silver | Sodium | Zinc | Mercury | |
| 11798 100% Baseline | 20 | 0 | <2.0 (<0.32) | 87900 | 375 (2.91) | <0.20 (<0.15) | |
| 11798 100% + SIR 300 ¹ | 100 | 8.6 | <2.0 (<0.36) | 228000 | 276 (2.26) | <0.20 (<0.15) | |
| RMHW ² + SIR 300 ¹ | 100 | 23.2 | <2.0 (<1.18) | 101000 | <10 (<0.15) | <0.20 (<0.15) | |
| RMHW ² Control | 100 | 25.8 | <2.0 (<0.66) | 25100 | <10 (<0.11) | <0.20 (<0.15) | |

Footnotes for Table 6.3.

Metal concentrations ($\mu\text{g/L}$) are reported from one replicate.

Toxic units (in parentheses) are based on TCEQ or EPA chronic surface water quality criteria for aquatic life protection. Toxic units for 11798, 11798 + SIR 300, RMHW + SIR 300 and RMHW are based on, where appropriate, hardness values of 128, 120, 60 and 80 mg/L as CaCO_3 , respectively.

Highlighted results indicate metals which have toxic unit greater than 1.0.

¹SIR 300 = SIR-300 ion-exchange resin, Resin Tech Inc., Cherry Hill, New Jersey.

²RMHW = Reconstituted Moderately Hard Water.

³EPA lists 1000 $\mu\text{g/L}$ as the water quality criterion for the protection of human health. No water quality criteria for the protection of aquatic life are available. US Environmental Protection Agency. 1986. Quality Criteria for Water. EPA/440/5-86-001. US Environmental Protection Agency, Office of Water Regulations and Standards, Washington, DC.

⁴EPA lists 100 $\mu\text{g/L}$ as the aquatic life protection criterion for total recoverable chromium. This value is used here because chromium measurements were not differentiated between Cr(III) and Cr(VI). US Environmental Protection Agency. 1980. Ambient Water Quality Criteria for Chromium. EPA/440/5-80-035. US Environmental Protection Agency, Office of Water Regulations and Standards, Criteria and Standards Division, Washington DC.

Table 6.4
Finfeather Lake Station 11798
Porewater Resin TIE

| Treatment | Mean | Metals ($\mu\text{g/L}$) | | | | | | | | Metals (mg/L) | | | |
|-----------------------|------|----------------------------|------|------|------|------|-----|------|------|--------------------------|------|------|-----|
| | | Al | As | Ba | Cr | Cu | Fe | Pb | Zn | Na | Ca | Mg | K |
| RMHW (control) | 23.2 | ND | ND | ND | ND | ND | ND | ND | ND | 25.4 | 13.6 | 11.6 | 1.8 |
| 11798 Baseline 25% | 22.0 | 270 | 26.5 | 62.0 | 4.0 | 181 | 220 | 3.3 | 68.0 | 39.3 | 9.5 | 1.2 | 1.1 |
| 11798 Baseline 50% | 17.0 | 540 | 53.0 | 124 | 7.9 | 362 | 440 | 6.5 | 136 | 78.5 | 19.0 | 2.4 | 2.3 |
| 11798 Baseline 100% | 11.4 | 1080 | 106 | 248 | 15.8 | 723 | 880 | 13.0 | 272 | 157 | 38.0 | 4.8 | 4.5 |
| 11798 SIR900 25% | 22.0 | 1070 | 23.1 | 17.7 | 2.8 | 57.5 | 192 | 5.5 | 96.3 | 30.3 | 3.6 | 0.7 | 0.8 |
| 11798 SIR900 50% | 17.6 | 2140 | 46.2 | 35.3 | 5.6 | 115 | 384 | 11.0 | 193 | 60.5 | 7.1 | 1.3 | 1.7 |
| 11798 SIR900 100% | 15.2 | 4280 | 92.3 | 70.6 | 11.1 | 230 | 767 | 21.9 | 385 | 121 | 14.2 | 2.7 | 3.4 |
| 11798 SIR300 25% | 21.2 | 230 | 40.0 | 11.6 | 3.4 | 74.3 | 239 | 6.9 | 120 | 48.0 | 0.7 | 0.1 | 1.1 |
| 11798 SIR300 50% | 23.0 | 459 | 80.0 | 23.2 | 6.8 | 149 | 477 | 13.8 | 240 | 96.0 | 1.5 | 0.2 | 2.1 |
| 11798 SIR300 100% | 2.2 | 918 | 160 | 46.3 | 13.6 | 297 | 954 | 27.6 | 480 | 192 | 3 | 0.4 | 4.3 |
| 11798 SIR300+900 25% | 23.0 | 1100 | 7.5 | ND | ND | 46.8 | 127 | 3.8 | 72.8 | 19.8 | 0.1 | 0.1 | 0.2 |
| 11798 SIR300+900 50% | 22.6 | 2200 | 15.0 | ND | ND | 93.5 | 255 | 7.7 | 146 | 39.5 | 0.2 | 0.2 | 0.5 |
| 11798 SIR300+900 100% | 15.0 | 4400 | 29.9 | ND | ND | 187 | 509 | 15.3 | 291 | 79.0 | 0.4 | 0.4 | 1.0 |

7-day Resin test initiated 2/2/02.

Porewater dilutions with Reconstituted Moderately Hard Water (RMHW)

Samples collected on 10/30/01

**Italicized values are derived from 100% baseline values and assume 50% dilution with RMHW.*

Nickel, selenium, silver, cadmium and mercury were all non-detected.

Table 6.5
Finfeather Lake Station 11798
Pore Water Resin/EDTA TIE

| Treatment | Mean | Std Dev | % Survival |
|------------------------|------|----------|------------|
| RHW (control) | 23 | 3.162278 | 100 |
| RMHW (control) | 25.8 | 3.03315 | 100 |
| RMHW EDTA 3mg/l | 25 | 2.738613 | 100 |
| RMHW SIR300 25% | 25 | 5.43139 | 100 |
| RMHW SIR300 50% | 27.6 | 2.073644 | 100 |
| RMHW SIR300 100% | 23.2 | 8.348653 | 100 |
| 11798 Baseline 25% | 26 | 1.870829 | 100 |
| 11798 Baseline 50% | 15.4 | 8.619745 | 100 |
| 11798 Baseline 100% | 0 | 0 | 20 |
| 11798 SIR300 25% | 26.2 | 5.80517 | 100 |
| 11798 SIR300 50% | 21.2 | 5.761944 | 100 |
| 11798 SIR300 100% | 8.6 | 4.335897 | 100 |
| 11798 EDTA 3 mg/L 25% | 25 | 2.708013 | 100 |
| 11798 EDTA 3 mg/L 50% | 23.2 | 3.271085 | 100 |
| 11798 EDTA 3 mg/L 100% | 16 | 3.162278 | 100 |

7-day *C. dubia* test initiated 6/6/02.

Porewater dilutions with Reconstituted Moderately Hard Water (RMHW)

Table 6.6
Finfeather Lake, Station 11798, Additions Study
Samples Collected 10/30/02

| | 11798 Baseline | 11798 SIR 300 (20%) | 11798 SIR 300 (50%) |
|--------------|-------------------|------------------------|------------------------|
| % Survival | 80 | 100 | 100 |
| Mean | 2 | 4.4 | 16 |
| 1 Std Dev | 0.82 | 0.89 | 2.12 |
| Aluminum (T) | 2000 | 1640 | 1460 |
| Aluminum (D) | ND (<100) | ND (<100) | ND (<100) |
| Arsenic (T) | 208 | 179 | 164 |
| Arsenic (D) | 181 | 180 | 159 |
| Copper (T) | 280 | 241 | 220 |
| Copper (D) | 37 | 69.4 | 59 |
| Lead (T) | 23.3 | 14.8 | 9.09 |
| Lead (D) | 4.01 | 3.86 | ND (<3.0) |
| Zinc (T) | 296 | 183 | 149 |
| Zinc (D) | 115 | 62.3 | 19.7 |

Table 6.6 Con't

| Metals Additions Results | | | | | |
|---------------------------------|------|------------|------|------------|------|
| % Survival | 100 | % Survival | 0 | % Survival | 100 |
| Mean | 4.2* | Mean | 0* | Mean | 14* |
| 1 Std Dev | 1.3 | 1 SD | 0 | 1 SD | 2.35 |
| Copper (T) | 441 | Zinc (T) | 350 | Lead (T) | 31.6 |
| Copper (D) | 172 | Zinc (D) | 42.7 | Lead (D) | 5.03 |

| Total Metals Concentrations with Additions | |
|---|-------------------------|
| Copper | 220 + 256 = 476 ug/L |
| Zinc | 149 + 317 = 466 ug/L |
| Lead | 9.09 + 26.4 = 35.5 ug/L |

Footnotes for all three parts to Table 6.6

All metals units are µg/L

All 11798 plus metals treatments were treated with SIR 300 (50%) prior to metal additions

* Tests was significantly different from the control (27.4 neonates/female)

(T) = Total metals

(D) = Dissolved metals

SECTION 7 SOURCE ANALYSIS AND IDENTIFICATION

Source Analysis was not initiated due to lack of funding and recently concluding the TIE that copper and zinc are the likely cause in pore water toxicity. More details concerning the TIE methods and results are provided in UNT's report located in Appendix D.

SECTION 8 SUMMARY AND CONCLUSIONS

Finfeather Lake (FFL) and Bryan Municipal Lake (BML) were sampled for a total of six events each during an initial 5-month period from April 2001 to August 2001. Once TIE studies began, routine whole sediment toxicity testing ceased. Detections of aluminum, arsenic, copper, and zinc at Station 11799 of FFL were above the corresponding screening levels. Pesticides, consisting of DDD, DDE, and DDT were also detected but were at concentrations too low to quantify. The lowest sediment screening levels for these pesticides were below the minimum analytical level for EPA method 8081. Station 11798 indicated detections of aluminum, arsenic, cadmium, chromium, copper, lead, nickel and zinc were above the corresponding screening level. Pesticide, DDT, was detected but was at a concentration too low to quantify. Detections of aluminum, arsenic, cadmium, copper, lead, and zinc at Station 11793 of BML were also above the corresponding screening levels. The pesticide DDD was detected above the sediment screening level, but was not quantifiable.

Toxicity test results for sediment samples taken in April and May 2001 indicated the sediments were significantly toxic due to lethality at Stations 11798 and 11800 in FFL and Station 11793 in BML to *Chironmus tentans* and *Hyalloella azteca* species using whole sediment toxicity test methods. See Tables 8.1 and 8.2. Statistically significant sublethal effects were also observed in sediment taken from Station 11800 in FFL and Stations 11792, 11793, and 11794 in BML. Due to the toxicity of the sediments, a TIE was initiated for both FFL and BML. Phase I of the TIE was initiated at Station 11798 in FFL and Station 11793 in BML. Since it is very likely that the same contaminant is affecting both lakes, it was decided to focus on the most toxic site first (in FFL) for the TIE.

**Table 8.1
 Sediment Toxicity Test Results**

| Bryan Municipal Lake 1209A | | % Survival | | Sub-Lethal Effect | |
|----------------------------|---------|-------------------|-------------------|-------------------|-------------------|
| | | Growth | | Chironmus tentans | Hyalloella azteca |
| | | Chironmus tentans | Hyalloella azteca | | |
| April 19, 2001 | Control | 81 | 91 | 0.706 | 0.112 |
| | 11792 | 69 | 79 | 0.455 | 0.091 |
| | 11793 | 31 | 84 | NA | 0.086 |
| | 11794 | 71 | 85 | 0.367 | 0.115 |
| May 21, 2001 | Control | 86 | 99 | CW | 0.167 |
| | 11792 | 66 | 84 * | CW | 0.128 |
| | 11793 | 70 | 92 | NA | 0.109 |
| | 11794 | 84 | 90 * | CW | 0.145 |

| Finfeather Lake 1209B | | % Survival | | Sub-Lethal Effect Growth | |
|-----------------------|-----------|-------------------|----------------|--------------------------|----------------|
| | | Chironmus tentans | Hyaella azteca | Chironmus tentans | Hyaella azteca |
| April 19,2001 | Control | 74 | 91 | CW | 0.112 |
| | 11798 | 33 | 54 | NA | NA |
| | 11799 | 58 | 86 | CW | 0.107 |
| | 11800 | 46 | 89 | NA | 0.094 |
| | 11800-Dup | 49 | 88 | NA | 0.071 |
| May 21, 2001 | Control | 75 | 99 | CW | 0.167 |
| | 11798 | 23 | 84 | NA | 0.091 |
| | 11799 | 79 | 80 | CW | 0.146 |
| | 11800 | 69 | 56 | CW | 0.112 |

Bold/Shaded cell - denotes significant difference from the control; duplicate is for quality control purposes only
 * Note that while statistically significant mortality effects were observed, the results did not exceed recommended criteria.
 CW- Control weight below minimum of 0.48 mg AFDW
 NA = Not Analyzed

Table 8.2
Summary of Sediment Toxicity Test Results

| Station | Lethal <i>C. tentans</i> | Lethal <i>H. azteca</i> | Sublethal <i>C. tentans</i> | Sublethal <i>H. azteca</i> |
|---------|--------------------------|-------------------------|-----------------------------|----------------------------|
| 11792 | 1/2 | 0/2 | 1/1 | 1/2 |
| 11793 | 1/2 | 0/2 | 1/1 | 2/2 |
| 11794 | 0/2 | 0/2 | 1/1 | 0/2 |
| 11798 | 2/2 | ½ | 0/0 | 2/2 |
| 11799 | 1/2 | 0/2 | 0/0 | 0/2 |
| 11800 | 1/2 | ½ | 0/0 | 1/2 |

U.S. EPA has not finalized sediment pore water or whole sediment Toxicity Identification Evaluation (TIE) methodology. Draft sediment TIE guidelines are available for pore waters and elutriates (EPA 1991) and closely follow effluent TIE procedures. Some whole sediment procedures for reducing toxicity of specific toxicant classes have been reported in the literature; however, whole sediment TIE procedures are not published in guideline format (Ho et al. 2002). Therefore, a tiered approach based on pore water tests was employed in this project (Ankley and Schubauer-Berigan 1995). Additional whole sediment TIE procedures were performed on Fin Feather Lake. Generally, 40-60% of sediment volume was isolated as pore water. *C. dubia* was chosen for pore water testing because of test volume requirements. *Hyaella azteca* and *Chironomus tentans* were also used to test whole sediments. Table 8.3 provides the TIE toxicity test results.

Table 8.3
C. dubia, 7-Day Toxicity Tests using Sediment Pore Water

| Sample Date | Test Date | Station | Treatment | Results | | |
|-------------|------------|---------|----------------|--------------|---------|------------|
| | | | | Reproduction | Std Dev | % Survival |
| 6/5/2001 | 7/31/2001 | 11798 | RHW (Control) | 26.4 | 2.19 | |
| | | | Baseline 100% | 15 | 2.55 | |
| 6/5/2001 | 7/31/2001 | 11793 | RHW (Control) | 26.4 | 2.19 | |
| | | | Baseline 100% | 15.2 | 3.11 | |
| 6/5/2001 | 8/25/2001 | 11798 | RHW (Control) | 24.2 | 2.28 | |
| | | | Baseline 100% | 14.6 | 1.82 | |
| 10/30/2001 | 2/2/2002 | 11798 | RMHW (Control) | 23 | 3.16 | 100 |
| | | | Baseline 100% | 11.4 | 0.894 | 100 |
| 10/30/2001 | 6/6/2002 | 11798 | RMHW (Control) | 25.8 | 3.03 | 100 |
| | | | Baseline 100% | 0 | 0 | 20 |
| 10/30/2001 | 6/26/2002 | 11798 | RMHW (Control) | 23.8 | 2.57 | 100 |
| | | | Baseline 100% | 10.6 | 4.9 | 100 |
| 10/30/2002 | 12/12/2002 | 11798 | RMHW (Control) | 27.4 | 2.07 | 100 |
| | | | Baseline 100% | 2 | 0.82 | 80 |

All general pore water TIE procedures followed EPA (1991) draft guidelines. Whole sediment TIEs followed procedures previously reported in the peer-reviewed literature. In addition to draft EPA TIE procedures, we used three ion exchange media to remove organic or metal toxicants. The cation exchange resin SIR-300, a styrene and divinylbenzene copolymer with iminodiacetic functional group in the sodium form, was chosen for metal removal because of its ability to chelate heavy metal cations (ResinTech, New Berlin NJ). SIR-300 was previously suggested as an effective metal treatment in sediment TIE procedures (Burgess et al. 2000). SIR-300 affinity for metals is: Hg²⁺>Cu²⁺>V²⁺>Pb²⁺>Ni²⁺>Zn²⁺>Co²⁺>Cd²⁺>Fe²⁺>Be²⁺, Mn²⁺>Mg²⁺, Ca²⁺>Sr²⁺>Ba²⁺>Na²⁺.

Although SIR-300 is a parallel TIE treatment to EDTA for divalent metals, we used SIR-300 in addition to EDTA because metals reduced by SIR-300 may be measured following TIE treatment. Because conventional TIE treatments are not effective for arsenic contaminated media, SIR-900, a synthetic aluminum oxide absorbent media specific for arsenic (arsenate and arsenite) and lead, was utilized in several TIE procedures for Fin Feather

Lake sediment because of historic arsenic contamination (ResinTech, West Berlin NJ). C18 solid phase extraction columns, typically used in TIE procedures to remove organic contaminants, may also filter or remove other contaminants (e.g. metals) and complicate TIE interpretation. Ambersorb 563, a carbonaceous adsorbent, for organic removal was used because it has 5 to 10 times the capacity of granular activated carbon. Ambersorb 563 in addition to C18 treatment in several TIEs was used to selectively remove organics without filtration complications. Ambersorb has been used to treat contaminated groundwater (EPA 1995) and lake water (Guzzella et al. 2002) and to remove organic contaminants in sediment TIE procedures (West et al. 2001). A summary of all TIEs performed on this segment is provided in Table 8.4.

Table 8.4
Sediment Toxicity Identification Evaluation Procedures

| Test Date | Test Type | Station | Organism | Effective Treatment |
|-------------------------|----------------|---------|------------------|---------------------|
| July 13-23, 2001 | Pore Water | 11793 | <i>C. dubia</i> | EDTA |
| July 13-23, 2001 | Pore Water | 11798 | <i>C. dubia</i> | None |
| Aug. 25 – Sept. 4, 2001 | Pore Water | 11798 | <i>C. dubia</i> | SIR300, SIR900 |
| February 2-12, 2002 | Pore Water | 11798 | <i>C. dubia</i> | SIR300, SIR900 |
| March 10-20, 2002 | Whole Sediment | 11798 | <i>H. azteca</i> | SIR900* |
| June 6-16, 2002 | Pore Water | 11798 | <i>C. dubia</i> | EDTA, SIR300 |
| June 6-16, 2002 | Pore Water | 11800 | <i>C. dubia</i> | EDTA |

**H. azteca* growth not significant different from control sediment.
 60% survival in SIR900, 68.3% survival in control.

In July 2001, it was determined that sediment pore water was not acutely toxic; however, pore water in sediment from stations 11793 and 11798 produced toxic effects on *C. dubia* reproduction. EDTA treatment reduced pore water toxicity at Station 11793, but not at 11798. A subsequent TIE was performed with several treatment media selective for metals removed toxicity in August 2001. These treatment media included: SIR-900 media selective for arsenic; TXI Shale, previously demonstrated to remove arsenic from aqueous solutions in sorbtion isotherm studies (F. Saleh, UNT, pers. comm.); and SIR-300. Station 11798 pore water was chosen for TIE treatments because of higher ambient sediment metal concentrations than other stations. Each TIE treatment improved *C. dubia* reproduction relative to untreated pore waters. Because arsenic was suspected as a causative toxicant, total arsenic, arsenate and arsenite levels were quantified in each TIE treatment. SIR-900 and TXI Shale treatments reduced total arsenic, arsenate and arsenite pore water concentrations.

Arsenic concentrations in pore waters were not sufficiently elevated to solely cause toxicity; total arsenic pore water concentration was 266 µg/L. Previous investigators found that arsenic treatment up to 1.46 mg/L did not significantly affect *C. dubia* reproduction. Other metals may have been removed by SIR-900 and TXI shale treatments; however, metal concentrations were not measured following these initial treatments.

In February 2002, another TIE was performed on station 11798 pore waters. Metal bioavailability and toxicity were reduced with increasing water hardness. Unlike previous TIEs, in which reconstituted hard water was used for pore water dilution, reconstituted moderately hard water was used for dilution in this TIE. Dilution water with lower hardness was chosen to maximize pore water metal bioavailability and toxicity, and potentially the effectiveness of TIE treatments. SIR-900 and SIR-900 + SIR-300 treatments significantly increased *C. dubia* reproduction in undiluted and 50% dilution, respectively. SIR-300 increased neonate production at 50% dilution, although not significantly, by an average of four neonates/female relative to baseline pore waters. Baseline pore water copper concentration was 722 µg/L, a value two orders of magnitude higher than previously reported lowest observed effect concentrations for *C. dubia* reproduction in laboratory water and substantially higher than the TSWQS of 48.6 µg/L. The lowest *C. dubia* fecundity was observed in undiluted SIR-300 TIE treatments. This sub-lethal response was attributed to reduction of the essential nutrients calcium and magnesium from pore waters in this test.

Whole sediments from Station 11798 were amended with SIR-300, SIR-900 and SIR-300 + SIR-900 at a 1:4, volume to volume (V:V) ratio in March 2002. Following a 10-day exposure period, *H. azteca* growth was significantly improved, relative to reference sediment from the University of North Texas Water Research Field Station, by only SIR-900 treatments.

In June 2002, a TIE was performed with Station 11798 sediment pore waters. As with the February study, reconstituted moderately hard water served as dilution water. Following SIR-300 treatment, hardness was measured by titration, and Ca and Mg salts reintroduced to pore waters until hardness values returned to pre-SIR-300 levels. EDTA (3 mg/L) and SIR-300 treatments significantly improved survival and reproduction relative to baseline pore waters. Because multiple metals were measured in pore waters, a toxic unit approach was taken to evaluate metal pore water toxicity (Table 8.5). Concentrations of zinc, iron, lead and barium decreased 26%, 32%, 37% and 96%, respectively, with SIR-300 treatments.

Bioavailability of these metals was clearly affected by compounds not accounted for in water hardness measures. An example is organic carbon binding to metals which affects bioavailability and toxicity. Total organic carbon in these pore waters was measured at elevated concentrations (baseline, 22.2 mg/L; SIR-300, 14.8 mg/L). An EDTA treatment of 3 mg/L also improved *C. dubia* survival and fecundity in Station 11798 pore waters. Average neonate production was 2x higher in EDTA treatments than in SIR-300 treatments. Although the manufacturer of SIR-300 indicated that this resin is more selective for zinc, iron, lead and copper than calcium and magnesium, our data suggests that this resin preferentially removed calcium and magnesium. If the binding capacity of SIR-300 was exhausted by preferential binding of calcium and magnesium ions, ligands that bound calcium and magnesium in

pretreated pore waters would be available for complexing with other divalent metals, specifically those metals measured at high enough concentrations (e.g. copper, to adversely affect *C. dubia* in 'clean' laboratory water toxicity tests).

A subsequent TIE study with SIR-300 was conducted to further remove metal contaminants from Station 11798 pore waters. By increasing the V:V ratio of SIR-300 to pore water during TIE treatment, SIR-300 metal binding capacity was increased and total metal pore water concentration were decreased. Following reintroduction of calcium and magnesium salts, effective in the June 2002 TIE with this pore water, reduced toxicity was more clearly assigned to potentially causative toxicants. Contaminant addition procedures (Phase III TIE) were subsequently performed to recreate pore water toxicity and provide corroborating information. Phase III TIE procedures were conducted such that pore waters were first treated with SIR 300 followed by reintroduction of copper, lead or zinc at nominal concentrations. Results indicated copper and zinc as the primary factors affecting aquatic life. A toxic units approach suggests copper to be of greater concern than zinc, however, 100% mortality was observed in zinc treated pore waters. This information also indicates metals similar to copper and zinc (examples) are of concern in both pore water and whole sediment. The exception to this is arsenic which does not appear to be a problem in pore waters. The containment addition test results suggest there is more toxicity effect from zinc than copper.

Parsons' recommends periodic monitoring of the sediment toxicity and development of a legacy TMDL for copper and zinc.

Table 8.5
Dissolved Metals Chemistry and Toxic Units of Finfeather Lake,
Station 11798, Sediment Pore Water Resin TIE
7-Day C. Dubia Test Initiated June 6, 2002*

| Treatment | % Survival | Mean # Neonates | Aluminum | Arsenic | Barium ³ | Cadmium | Calcium |
|--|------------|-----------------|-----------------------|-------------|---------------------|---------------|---------|
| 11798 100% Baseline | 20 | 0 | 707 (0.713) | 212 (1.12) | 225 (0.225) | <1.0 (<0.8) | 43500 |
| 11798 100% + SIR 300 ¹ | 100 | 8.6 | 1190 (1.20) | 260 (1.37) | <10 (<0.01) | <1.0 (<0.8) | 1940 |
| RMHW ² + SIR 300 ¹ | 100 | 23.2 | <100 (0.100) | <10 (<0.05) | <10 (<0.01) | <1.0 (<1.45) | 1200 |
| RMHW ² Control | 100 | 25.8 | <100 (0.100) | <10 (<0.05) | <10 (<0.01) | <1.0 (<1.11) | 15400 |
| Treatment | % Survival | Mean # Neonates | Chromium ⁴ | Copper | Iron | Lead | |
| 11798 100% Baseline | 20 | 0 | 11.2 (0.011) | 251 (16.5) | 1410 (1.41) | 24.2 (7.01) | |
| 11798 100% + SIR 300 ¹ | 100 | 8.6 | 15.8 (0.016) | 290 (20.2) | 955 (0.96) | 15.3 (4.81) | |
| RMHW ² + SIR 300 ¹ | 100 | 23.2 | <10 (<0.01) | <10 (<1.26) | <100 (<0.10) | <3.0 (<2.27) | |
| RMHW ² Control | 100 | 25.8 | <10 (<0.01) | <10 (<0.95) | <100 (<0.10) | <3.0 (<1.48) | |
| Treatment | % Survival | Mean # Neonates | Magnesium | Nickel | Potassium | Selenium | |
| 11798 100% Baseline | 20 | 0 | 4910 | <10 (<0.05) | 4060 | <10 (<2.0) | |
| 11798 100% + SIR 300 ¹ | 100 | 8.6 | 1570 | <10 (<0.05) | 4670 | <10 (<2.0) | |
| RMHW ² + SIR 300 ¹ | 100 | 23.2 | 1120 | <10 (<0.10) | 1760 | <10 (<2.0) | |
| RMHW ² Control | 100 | 25.8 | 11000 | <10 (<0.07) | 1840 | <10 (<2.0) | |
| Treatment | % Survival | Mean # Neonates | Silver | Sodium | Zinc | Mercury | |
| 11798 100% Baseline | 20 | 0 | <2.0 (<0.32) | 87900 | 375 (2.91) | <0.20 (<0.15) | |
| 11798 100% + SIR 300 ¹ | 100 | 8.6 | <2.0 (<0.36) | 228000 | 276 (2.26) | <0.20 (<0.15) | |
| RMHW ² + SIR 300 ¹ | 100 | 23.2 | <2.0 (<1.18) | 101000 | <10 (<0.15) | <0.20 (<0.15) | |
| RMHW ² Control | 100 | 25.8 | <2.0 (<0.66) | 25100 | <10 (<0.11) | <0.20 (<0.15) | |

Footnotes for Table 8.5

* Sample collected on June 5, 2002.

Metal concentrations (µg/L) are reported from one replicate.

Toxic units (in parentheses) are based on TCEQ or EPA chronic surface water quality criteria for aquatic life protection. Toxic units for 11798, 11798 + SIR 300, RMHW + SIR 300 and RMHW are based on, where appropriate, hardness values of 128, 120, 60 and 80 mg/L as CaCO₃, respectively.

¹SIR 300 = SIR-300 ion-exchange resin, Resin Tech Inc., Cherry Hill, New Jersey.

²RMHW = Reconstituted Moderately Hard Water.

³EPA lists 1000 µg/L as the water quality criterion for the protection of human health. No water quality criteria for the protection of aquatic life are available. US Environmental Protection Agency. 1986. Quality Criteria for Water. EPA/440/5-86-001. US Environmental Protection Agency, Office of Water Regulations and Standards, Washington, DC.

⁴EPA lists 100 µg/L as the aquatic life protection criterion for total recoverable chromium. This value is used here because chromium measurements were not differentiated between Cr(III) and Cr(VI). US Environmental Protection Agency. 1980. Ambient Water Quality Criteria for Chromium. EPA/440/5-80-035. US Environmental Protection Agency, Office of Water Regulations and Standards, Criteria and Standards Division, Washington DC.

Table 8.6
Finfeather Lake Segment 1209B
Whole Sediment Chemistry and Toxic Units

| | Station ID 11799 | Station ID 11798 | Station ID 11799 | Station ID 11798 | Station ID 11800 | | |
|-----------|----------------------------------|----------------------------------|----------------------------------|---------------------------------|---------------------------------|--------------------------------|------------------|
| PARAMETER | Sample Collected 5/21/2001 | Sample Collected 7/18/2001 | Sample Collected 7/18/2001 | Sample Collected 5/9/2002 | Sample Collected 5/9/2002 | Lowest Screening Values* | UNITS |
| Toxicity | Toxic ¹ | NA | NA | NA | NA | | |
| Arsenic | 58.5 (8.08) | 196 (27.1) | 28.8 (4.0) | 79.2 (10.9) | 160 (22.1) | 7.24 | mg/Kg -dry wt |
| Copper | 65.4 (3.5) | 575 (30.7) | 44.5 (2.37) | 171 (9.14) | 113 (6.04) | 18.7 | mg/Kg -dry wt |
| Lead | 17.5 | 56.9 (1.88) | 12.6 | 33.3 (1.10) | 51.8 (1.71) | 30.24 | mg/Kg -dry wt |
| Zinc | 241 (1.94) | 1280 (10.3) | 151 (1.22) | 447 (3.61) | 466 (3.76) | 124 | mg/Kg -dry wt |

Notes:

* Criteria is from *Equilibrium and Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices* tables. The value is the lowest value from the Indices as stated in the Appendix.

mg/kg-dry = milligrams per kilogram dry weight

¹ No significant difference from control for survival and growth of *C. tentans* in 10 day sediment exposures; significant difference in survival of *H. azteca* in 10 day sediment exposure.

() = Toxic Units; calculated by dividing detected metal concentration by the lowest screening level.

Bold and highlighted results indicate TU is above 1.0

Table 8.7
Bryan Municipal Lake 1209A
Whole Sediment Chemistry and Toxic Units

| | Station ID 11793 | Station ID 11793 | Station ID 11792 | Station ID 11793 | Station ID 11794 | | |
|-----------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|--------------------------------|------------------|
| PARAMETER | Sample Collected 5/21/2001 | Sample Collected 7/18/2001 | Sample Collected 7/12/2002 | Sample Collected 7/12/2002 | Sample Collected 7/12/2002 | Lowest Screening Values* | UNITS |
| Toxicity | Toxic ¹ | NA | Not Toxic ² | Toxic ³ | Not Toxic ² | | |
| Arsenic | 57.6 (7.96) | 95.8 (13.2) | 17.8 (2.46) | 90.2 (12.5) | 141 (19.5) | 7.24 | mg/Kg- dry wt |
| Copper | 52.5 (2.8) | 40.6 (2.17) | 13.9 | 44 (2.35) | 178 (9.5) | 18.7 | mg/Kg- dry wt |
| Lead | 36.4 (1.21) | 37.1 (1.23) | 21.8 | 42.3 (1.40) | 99.7 (3.30) | 30.24 | mg/Kg- dry wt |
| Zinc | 227 (1.83) | 183 (1.5) | 67.5 | 215 (1.73) | 799 (6.44) | 124 | mg/Kg- dry wt |

Notes:

* Criteria is from *Equilibrium and Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices* tables. The value is the lowest value from the Indices as stated in the Appendix.

mg/kg-dry = milligrams per kilogram dry weight

¹ No significant difference from control for survival of *C. tentans* and *H. azteca* in 10 day sediment exposures; significant difference in growth for sublethal effects of *H. azteca*.

² No significant difference from control for survival and growth of *H. azteca* in 10 day sediment exposures.

³ No significant difference from control for survival of *H. azteca* in 10 day sediment exposures; although significant difference in growth for sublethal effects of *H. azteca*.

() = Toxic Units; calculated by dividing detected metal concentration by the lowest screening level.

Bold and highlighted results indicate TU is above 1.0

Table 8.8
Finfeather Lake, Station 11798, Additions Study
Samples Collected 10/30/02

| | 11798 Baseline | 11798 SIR 300 (20%) | 11798 SIR 300 (50%) |
|--------------|-------------------|------------------------|------------------------|
| % Survival | 80 | 100 | 100 |
| Mean | 2 | 4.4 | 16 |
| 1 Std Dev | 0.82 | 0.89 | 2.12 |
| Aluminum (T) | 2000 | 1640 | 1460 |
| Aluminum (D) | ND (<100) | ND (<100) | ND (<100) |
| Arsenic (T) | 208 | 179 | 164 |
| Arsenic (D) | 181 | 180 | 159 |
| Copper (T) | 280 | 241 | 220 |
| Copper (D) | 37 | 69.4 | 59 |
| Lead (T) | 23.3 | 14.8 | 9.09 |
| Lead (D) | 4.01 | 3.86 | ND (<3.0) |
| Zinc (T) | 296 | 183 | 149 |
| Zinc (D) | 115 | 62.3 | 19.7 |

| Metals Additions Results | | | | | |
|---------------------------------|------|------------|------|------------|------|
| % Survival | 100 | % Survival | 0 | % Survival | 100 |
| Mean | 4.2* | Mean | 0* | Mean | 14* |
| 1 Std Dev | 1.3 | 1 SD | 0 | 1 SD | 2.35 |
| Copper (T) | 441 | Zinc (T) | 350 | Lead (T) | 31.6 |
| Copper (D) | 172 | Zinc (D) | 42.7 | Lead (D) | 5.03 |

| Total Metals Concentrations with Additions | |
|---|-------------------------|
| Copper | 220 + 256 = 476 ug/L |
| Zinc | 149 + 317 = 466 ug/L |
| Lead | 9.09 + 26.4 = 35.5 ug/L |

Footnotes for all three parts to Table 8.8

All metals units are µg/L

All 11798 plus metals treatments were treated with SIR 300 (50%) prior to metal additions

* Tests was significantly different from the control (27.4 neonates/female)

(T) = Total metals

(D) = Dissolved metals

SECTION 9 REFERENCES

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**APPENDIX A
HISTORICAL DATA**

Appendix A
Bryan Municipal Lake Historical Sediment Chemical Analysis

| Station | Long Description | Data | Total |
|--|---|------------------|-------|
| 11792 | 1,1,1-TRICHLOROETHANE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | 1,1,2,2-TETRACHLOROETHANE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | 1,1,2-TRICHLOROETHANE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | 1,1-DICHLOROETHANE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | 1,1-DICHLOROETHYLENE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| 1,2,4,5-TETRACHLOROBENZENE SEDIMENT DRY WT (UG/K | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,2,4-TRICHLOROBENZENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,2,5,6-DIBENZANTHRACENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,2-DIBROMOETHANE SEDIMENT, DRY WEIGHT (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,2-DICHLOROBENZENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,2-DICHLOROETHANE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,2-DICHLOROPROPANE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,2-DIPHENYLHYDRAZINE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,3-DICHLOROBENZENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 1,4-DICHLOROBENZENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 2,4,5-TRICHLOROPHENOL IN SEDIMENT.DRY WT (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 2,4,6-TRICHLOROPHENOL DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 2,4-DICHLOROPHENOL DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 2,4-DIMETHYLPHENOL DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 2,4-DINITROPHENOL DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |

Appendix A
Bryan Municipal Lake Historical Sediment Chemical Analysis

| | | | |
|---|---|------------------|-----|
| 11792 | 2,4-DINITROTOLUENE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | 2,6-DINITROTOLUENE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | 2-CHLOROETHYL VINYL ETHER DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | 2-CHLORONAPHTHALENE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | 2-CHLOROPHENOL DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| 2-NITROPHENOL DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 3,3'-DICHLOROBENZIDINE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 4-BROMOPHENYL PHENYL ETHER DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 4-CHLOROPHENYL PHENYL ETHER DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| 4-NITROPHENOL DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| ACENAPHTHENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| ACENAPHTYLENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| ACRYLONITRILE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| ALUMINUM IN BOTTOM DEPOSITS (MG/KG AS AL DRY WGT) | Min of Value | 9060 | |
| | Max of Value | 38000 | |
| | Average of Value | 21310.0 | |
| | Count of Value | 6 | |
| ANTHRACENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| ARSENIC IN BOTTOM DEPOSITS (MG/KG AS AS DRY WGT) | Min of Value | 37.3 | |
| | Max of Value | 395 | |
| | Average of Value | 128.8 | |
| | Count of Value | 6 | |
| BARIUM IN BOTTOM DEPOSITS (MG/KG AS BA DRY WGT) | Min of Value | 106 | |
| | Max of Value | 374 | |
| | Average of Value | 218.8 | |
| | Count of Value | 6 | |
| B-BHC-BETA DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| BENZENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| BENZO(B)FLUORANTHENE,SEDIMENTS, DRY WGT,UG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| BENZO(K)FLOURANTHENE DRY WTBOT UG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |

Appendix A
Bryan Municipal Lake Historical Sediment Chemical Analysis

| | | | |
|---|--|------------------|-----|
| 11792 | BENZO-A-PYRENE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | BIS (2-CHLOROETHOXY) METHANE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | BIS (2-CHLOROETHYL) ETHER DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | BIS (2-CHLOROISOPROPYL) ETHER DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | BIS(2-ETHYLHEXYL) PHTHALATE SED, DRY WGT,UG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| BROMODICHLOROMETHANE DRY WEIGHT BOTTOM (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| BROMOFORM DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| BROMOMETHANE IN SEDIMENT, (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| CADMIUM,TOTAL IN BOTTOM DEPOSITS (MG/KG,DRY WGT) | Min of Value | 0.17 | |
| | Max of Value | 0.79 | |
| | Average of Value | 0.5 | |
| | Count of Value | 6 | |
| CARBON TETRACHLORIDE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| CHLORDANE(TECH MIX&METABS) SED,DRY WGT,UG/KG | Min of Value | 290 | |
| | Max of Value | 290 | |
| | Average of Value | 290.0 | |
| | Count of Value | 1 | |
| CHLOROBENZENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| CHLOROETHANE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| CHLOROFORM DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| CHLOROMETHANE SEDIMENT DRY WEIGHT (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| CHROMIUM,TOTAL IN BOTTOM DEPOSITS (MG/KG,DRY WGT) | Min of Value | 10.1 | |
| | Max of Value | 132 | |
| | Average of Value | 43.3 | |
| | Count of Value | 6 | |
| CHRYSENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| CIS-1,3-DICHLOROPROPENE SEDIMENT DRY WGT UG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| COPPER IN BOTTOM DEPOSITS (MG/KG AS CU DRY WGT) | Min of Value | 10.7 | |
| | Max of Value | 267 | |
| | Average of Value | 71.6 | |
| | Count of Value | 6 | |
| DELTA BENZENE HEXACHLORIDE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| DIBROMOCHLOROMETHANE DRY WEIGHT BOTTOM (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |

Appendix A
Bryan Municipal Lake Historical Sediment Chemical Analysis

| | | | |
|---|---|------------------|-----|
| 11792 | DIETHYL PHTHALATE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | DIMETHYL PHTHALATE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | DI-N-BUTYL PHTHALATE, SEDIMENTS, DRY WGT, UG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | DI-N-OCTYL PHTHALATE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | DURSBAN BOTTOM DEPOSITS DRY WGT (UG/KG) | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| ENDOSULFAN SULFATE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| ETHYLBENZENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| FLUORANTHENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| FLUORENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| GAMMA BHC (LINDANE), SEDIMENT, DRY WT (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| HEXACHLOROBUTADIENE BOT. DEPOS. (UG/KG DRY WGT) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| HEXACHLOROCYCLOPENTADIENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| HEXACHLOROETHANE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| INDENO (1,2,3-CD) PYRENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| ISOPHORONE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| LEAD IN BOTTOM DEPOSITS (MG/KG AS PB DRY WGT) | Min of Value | 22.4 | |
| | Max of Value | 65.1 | |
| | Average of Value | 51.7 | |
| | Count of Value | 6 | |
| MANGANESE IN BOTTOM DEPOSITS (MG/KG AS MN DRY WG) | Min of Value | 135 | |
| | Max of Value | 391 | |
| | Average of Value | 200.3 | |
| | Count of Value | 6 | |
| MERCURY, TOT. IN BOT. DEPOS. (MG/KG) AS HG DRY WG | Min of Value | 0 | |
| | Max of Value | 0.553 | |
| | Average of Value | 0.2 | |
| | Count of Value | 6 | |
| METHYLENE CHLORIDE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| MIREX SEDIMENT, DRY WT (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| NAPHTHALENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |

Appendix A
Bryan Municipal Lake Historical Sediment Chemical Analysis

| | | | |
|--|--|------------------|--------|
| 11792 | N-BUTYL BENZYL PHTHALATE, SEDIMENTS, DRY WGT, UG/K | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | NICKEL, TOTAL IN BOTTOM DEPOSITS (MG/KG, DRY WGT) | Min of Value | 5.16 |
| | | Max of Value | 20.7 |
| | | Average of Value | 11.4 |
| | | Count of Value | 6 |
| | NITROBENZENE DRY WGT BOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| | NITROGEN KJELDAHL TOTAL BOTTOM DEP DRY WT MG/KG | Min of Value | 2980 |
| | | Max of Value | 3700 |
| | | Average of Value | 3443.3 |
| | | Count of Value | 3 |
| | N-NITROSODIETHYLAMINE, SED DRY WT (UG/KG) | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| N-NITROSODIMETHYLAMINE DRY WGT BOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| N-NITROSO-DI-N-BUTYLAMINE, DRY WT, SEDIMENT (UG/K) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| N-NITROSO-DI-N-PROPYLAMINE DRY WGT BOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| N-NITROSODIPHENYLAMINE DRY WGT BOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| PARACHLOROMETA CRESOL DRY WGT BOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| PCB-1016 IN BOTTOM SEDIMENTS DRY WT (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| PHENANTHRENE DRY WGT BOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| PHENOL(C6H5OH)-SINGLE COMPOUND DRY WGTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| PHOSPHORUS, TOTAL, BOTTOM DEPOSIT (MG/KG DRY WGT) | Min of Value | 668 | |
| | Max of Value | 1070 | |
| | Average of Value | 821.7 | |
| | Count of Value | 3 | |
| PYRENE DRY WGT BOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| PYRIDINE SEDIMENT DRY WEIGHT (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| SEDIMENT PRCTL.SIZE CLASS 0.0039 CLAY %DRY WT | Min of Value | 6 | |
| | Max of Value | 53.6 | |
| | Average of Value | 21.6 | |
| | Count of Value | 6 | |
| SEDIMENT PRCTL.SIZE CLASS, SAND .0625-2MM %DRY W | Min of Value | 7 | |
| | Max of Value | 77.3 | |
| | Average of Value | 26.3 | |
| | Count of Value | 6 | |
| SEDIMENT PRCTL.SIZE CLASS >2.0MM GRAVEL %DRY WT | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 6 | |
| SEDIMENT PRCTL.SIZE CLASS.0039-.0625 SILT %DRY W | Min of Value | 10 | |
| | Max of Value | 87 | |
| | Average of Value | 52.1 | |
| | Count of Value | 6 | |
| SELENIUM IN BOTTOM DEPOSITS (MG/KG AS SE DRY WT) | Min of Value | 0 | |
| | Max of Value | 2.82 | |
| | Average of Value | 0.8 | |
| | Count of Value | 6 | |

Appendix A
Bryan Municipal Lake Historical Sediment Chemical Analysis

| | | | |
|--|---|------------------|--------|
| 11792 | SILVER IN BOTTOM DEPOSITS (MG/KG AS AG DRY WGT) | Min of Value | 0 |
| | | Max of Value | 0.91 |
| | | Average of Value | 0.2 |
| | | Count of Value | 6 |
| | SOLIDS IN SEDIMENT, PERCENT BY WEIGHT (DRY) | Min of Value | 24.91 |
| | | Max of Value | 48.2 |
| | | Average of Value | 31.2 |
| | | Count of Value | 6 |
| | TETRACHLOROETHYLENE DRY WGTBOTUG/KG | Min of Value | 0 |
| | | Max of Value | 0 |
| | | Average of Value | 0.0 |
| | | Count of Value | 1 |
| TOLUENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| TOTAL ORGANIC CARBON IN SEDIMENT DRY WGT (MG/KG) | Min of Value | 19500 | |
| | Max of Value | 88500 | |
| | Average of Value | 40958.3 | |
| | Count of Value | 6 | |
| TRANS-1,2-DICHLOROETHENE, IN SED. DRY WT. UG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| TRANS-1,3-DICHLOROPROPENE SEDIMENT DRY WGT UG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| TRICHLOROETHYLENE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| VINYL CHLORIDE DRY WGTBOTUG/KG | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| XYLENE SEDIMENT, DRY WGT (UG/KG) | Min of Value | 0 | |
| | Max of Value | 0 | |
| | Average of Value | 0.0 | |
| | Count of Value | 1 | |
| ZINC IN BOTTOM DEPOSITS (MG/KG AS ZN DRY WGT) | Min of Value | 0 | |
| | Max of Value | 223 | |
| | Average of Value | 106.9 | |
| | Count of Value | 6 | |
| 11792 Min of Value | | | 0 |
| 11792 Max of Value | | | 88500 |
| 11792 Average of Value | | | 1834.1 |
| 11792 Count of Value | | | 214 |
| Total Min of Value | | | 0 |
| Total Max of Value | | | 88500 |
| Total Average of Value | | | 1834.1 |
| Total Count of Value | | | 214 |

Appendix A
Finfeather Lake
Historical Sediment Chemical Analysis

| Station | Long Description | Data | Total |
|---|--|-----------------------|-------|
| 11798 | 1,1,1-TRICHLOROETHANE DRY WGTBOTUG/KG | Min of Value w ND | 0 |
| | | Max of Value w ND | 0 |
| | | Average of Value w ND | 0.0 |
| | | Count of Value w ND | 2 |
| | 1,1,2,2-TETRACHLOROETHANE DRY WGTBOTUG/KG | Min of Value w ND | 0 |
| | | Max of Value w ND | 0 |
| | | Average of Value w ND | 0.0 |
| | | Count of Value w ND | 2 |
| | 1,1,2-TRICHLOROETHANE DRY WGTBOTUG/KG | Min of Value w ND | 0 |
| | | Max of Value w ND | 0 |
| | | Average of Value w ND | 0.0 |
| | | Count of Value w ND | 2 |
| | 1,1-DICHLOROETHANE DRY WGTBOTUG/KG | Min of Value w ND | 0 |
| | | Max of Value w ND | 0 |
| | | Average of Value w ND | 0.0 |
| | | Count of Value w ND | 2 |
| | 1,1-DICHLOROETHYLENE DRY WGTBOTUG/KG | Min of Value w ND | 0 |
| | | Max of Value w ND | 0 |
| | | Average of Value w ND | 0.0 |
| | | Count of Value w ND | 2 |
| | 1,2,4,5-TETRACHLOROBENZENE SEDIMENT DRY WT (UG/K | Min of Value w ND | 0 |
| | | Max of Value w ND | 0 |
| | | Average of Value w ND | 0.0 |
| | | Count of Value w ND | 2 |
| | 1,2,4-TRICHLOROBENZENE DRY WGTBOTUG/KG | Min of Value w ND | 0 |
| | | Max of Value w ND | 0 |
| | | Average of Value w ND | 0.0 |
| | | Count of Value w ND | 2 |
| | 1,2,5,6-DIBENZANTHRACENE DRY WGTBOTUG/KG | Min of Value w ND | 0 |
| | | Max of Value w ND | 0 |
| Average of Value w ND | | 0.0 | |
| Count of Value w ND | | 2 | |
| 1,2-DIBROMOETHANE SEDIMENT, DRY WEIGHT (UG/KG) | Min of Value w ND | 0 | |
| | Max of Value w ND | 1300 | |
| | Average of Value w ND | 650.0 | |
| | Count of Value w ND | 2 | |
| 1,2-DICHLOROBENZENE DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 1,2-DICHLOROETHANE DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 1,2-DICHLOROPROPANE DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 1,2-DIPHENYLHYDRAZINE DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 1,3-DICHLOROBENZENE DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 1,4-DICHLOROBENZENE DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 2,4,5-TRICHLOROPHENOL IN SEDIMENT, DRY WT (UG/KG) | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 2,4,6-TRICHLOROPHENOL DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 2,4-DICHLOROPHENOL DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 2,4-DIMETHYLPHENOL DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 2,4-DINITROPHENOL DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |
| | Count of Value w ND | 2 | |
| 2,4-DINITROTOLUENE DRY WGTBOTUG/KG | Min of Value w ND | 0 | |
| | Max of Value w ND | 0 | |
| | Average of Value w ND | 0.0 | |

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Finfeather Lake

| 11798 | 2,4-DINITROTOLUENE DRY WGT/BOTOM/KG | Historical Sediment Chemical Analysis | Count of Value w ND | 2 |
|-------|-------------------------------------|---------------------------------------|---------------------|---|
|-------|-------------------------------------|---------------------------------------|---------------------|---|

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Finfeather Lake

Historical Sediment Chemical Analysis

| | | | |
|-------|---|--|--------------------------------|
| 11798 | 2,6-DINITROTOLUENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | 2-CHLOROETHYL VINYL ETHER DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | 2-CHLORONAPHTHALENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | 2-CHLOROPHENOL DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | 2-NITROPHENOL DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | 3,3'-DICHLOROBENZIDINE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | 4-BROMOPHENYL PHENYL ETHER DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | 4-CHLOROPHENYL PHENYL ETHER DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | 4-NITROPHENOL DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | ACENAPHTHENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | ACENAPHTYLENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | ACRYLONITRILE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | ALUMINUM IN BOTTOM DEPOSITS (MG/KG AS AL DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 13800 36800 20566.7 6 |
| | ANTHRACENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | ARSENIC IN BOTTOM DEPOSITS (MG/KG AS AS DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 91.2 441 222.0 6 |
| | BARIUM IN BOTTOM DEPOSITS (MG/KG AS BA DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 182 429 269.3 6 |
| | B-BHC-BETA DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BENZENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BENZO(B)FLUORANTHENE, SEDIMENTS, DRY WGT,UG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BENZO(K)FLOURANTHENE DRY WTBOT UG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BENZO-A-PYRENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BIS (2-CHLOROETHOXY) METHANE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND | 0 0 |

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| Historical Sediment Chemical Analysis | | | |
|---------------------------------------|---|--|----------------------------|
| 11798 | BIS (2-CHLOROETHOXY) METHANE DRY WGTBOTUG/KG | Average of Value w ND Count of Value w ND | 0.0 2 |
| | BIS (2-CHLOROETHYL) ETHER DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BIS (2-CHLOROISOPROPYL) ETHER DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BIS(2-ETHYLHEXYL) PHTHALATE SED, DRY WGT,UG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BROMODICHLOROMETHANE DRY WEIGHT BOTTOM (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BROMOFORM DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | BROMOMETHANE IN SEDIMENT, (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | CADMIUM, TOTAL IN BOTTOM DEPOSITS (MG/KG, DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0.195 0.936 0.5 6 |
| | CARBON TETRACHLORIDE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | CHLORDANE (TECH MIX&METABS) SED, DRY WGT, UG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 39 19.5 2 |
| | CHLOROBENZENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | CHLOROETHANE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | CHLOROFORM DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | CHLOROMETHANE SEDIMENT DRY WEIGHT (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | CHROMIUM, TOTAL IN BOTTOM DEPOSITS (MG/KG, DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 26.4 144 78.6 6 |
| | CHRYSENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | CIS-1,3-DICHLOROPROPENE SEDIMENT DRY WGT UG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | COPPER IN BOTTOM DEPOSITS (MG/KG AS CU DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 32.6 276 160.1 6 |
| | DELTA BENZENE HEXACHLORIDE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | DIBROMOCHLOROMETHANE DRY WEIGHT BOTTOM (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | DIETHYL PHTHALATE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | DIMETHYL PHTHALATE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |

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Historical Sediment Chemical Analysis

| | | | |
|-------|--|--|-----------------------------|
| 11798 | DI-N-BUTYL PHTHALATE, SEDIMENTS, DRY WGT, UG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | DI-N-OCTYL PHTHALATE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | DNOC (4,6-DINITRO-ORTHO-CRESOL) DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 1 |
| | DURSBAN BOTTOM DEPOSITS DRY WGT (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | ENDOSULFAN SULFATE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | ETHYLBENZENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | FLUORANTHENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | FLUORENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | GAMMA BHC (LINDANE), SEDIMENT, DRY WT (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | HEXACHLOROBUTADIENE BOT. DEPOS. (UG/KG DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | HEXACHLOROCYCLOPENTADIENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | HEXACHLOROETHANE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | INDENO (1,2,3-CD) PYRENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | ISOPHORONE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | LEAD IN BOTTOM DEPOSITS (MG/KG AS PB DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 24 73.2 47.8 6 |
| | MANGANESE IN BOTTOM DEPOSITS (MG/KG AS MN DRY WG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 134 394 253.2 6 |
| | MERCURY, TOT. IN BOT. DEPOS. (MG/KG) AS HG DRY WG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0.0689 0.392 0.2 6 |
| | METHYLENE CHLORIDE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | MIREX SEDIMENT, DRY WT (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | NAPHTHALENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | N-BUTYL BENZYL PHTHALATE, SEDIMENTS, DRY WGT, UG/K | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | NICKEL, TOTAL IN BOTTOM DEPOSITS (MG/KG, DRY WGT) | Min of Value w ND Max of Value w ND | 9.99 21.6 |

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| Historical Sediment Chemical Analysis | | | |
|---------------------------------------|--|--|-----------------------------|
| 11798 | NICKEL, TOTAL IN BOTTOM DEPOSITS (MG/KG DRY WGT) | Average of Value w ND Count of Value w ND | 16.5 6 |
| | NITROBENZENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | NITROGEN KJELDAHL TOTAL BOTTOM DEP DRY WT MG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 3950 5256 4768.7 3 |
| | N-NITROSODIETHYLAMINE, SED DRY WT (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | N-NITROSODIMETHYLAMINE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | N-NITROSO-DI-N-BUTYLAMINE, DRY WT, SEDIMENT (UG/K) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | N-NITROSODI-N-PROPYLAMINE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | N-NITROSODIPHENYLAMINE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | PARACHLOROMETA CRESOL DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | PCB-1016 IN BOTTOM SEDIMENTS DRY WT (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | PHENANTHRENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | PHENOL(C6H5OH)-SINGLE COMPOUND DRY WGTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | PHOSPHORUS, TOTAL, BOTTOM DEPOSIT (MG/KG DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 974 1360 1118.0 3 |
| | PYRENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | PYRIDINE SEDIMENT DRY WEIGHT (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | SEDIMENT PRCTL.SIZE CLASS 0.0039 CLAY %DRY WT | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 7 62.51 28.6 6 |
| | SEDIMENT PRCTL.SIZE CLASS,SAND .0625-2MM %DRY W | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 45.9 13.9 6 |
| | SEDIMENT PRCTL.SIZE CLASS >2.0MM GRAVEL %DRY WT | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 6 |
| | SEDIMENT PRCTL.SIZE CLASS .0039-.0625 SILT %DRY W | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 18 93 57.3 6 |
| | SELENIUM IN BOTTOM DEPOSITS (MG/KG AS SE DRY WT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 2.35 0.7 6 |
| | SILVER IN BOTTOM DEPOSITS (MG/KG AS AG DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0.731 0.1 6 |
| | SOLIDS IN SEDIMENT, PERCENT BY WEIGHT (DRY) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 20.13 32.79 24.8 6 |

Appendix A
Finfeather Lake

Historical Sediment Chemical Analysis

| | | | |
|-----------------------------|--|--|---------------------------------|
| 11798 | TETRACHLOROETHYLENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | TOLUENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | TOTAL ORGANIC CARBON IN SEDIMENT DRY WGT (MG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 23300 125200 48500.0 5 |
| | TRANS-1,2-DICHLOROETHENE, IN SED. DRY WT. UG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | TRANS-1,3-DICHLOROPROPENE SEDIMENT DRY WGT UG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | TRICHLOROETHYLENE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | VINYL CHLORIDE DRY WGTBOTUG/KG | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 2 |
| | XYLENE SEDIMENT, DRY WGT (UG/KG) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 0 0.0 1 |
| | ZINC IN BOTTOM DEPOSITS (MG/KG AS ZN DRY WGT) | Min of Value w ND Max of Value w ND Average of Value w ND Count of Value w ND | 0 966 415.3 6 |
| | 11798 Min of Value w ND | | |
| 11798 Max of Value w ND | | | 125200 |
| 11798 Average of Value w ND | | | 1284.8 |
| 11798 Count of Value w ND | | | 307 |
| Total Min of Value w ND | | | 0 |
| Total Max of Value w ND | | | 125200 |
| Total Average of Value w ND | | | 1284.8 |
| Total Count of Value w ND | | | 307 |

**APPENDIX B
HISTORICAL REPORT BY TCEQ ON FINEATHER AND
BRYAN MUNICIPAL LAKES**

**APPENDIX C
PHOTO LOG**

FINFEATHER LAKE



Segment 1209B, Finfeather Lake, looking downstream towards dam and Atkins Power Plant (2001).



Method -1

FINFEATHER LAKE



Station 5



Segment 12029B, Station 11799, Finfeather Lake mainbody, white pole identifies sample location (2001).

FINFEATHER LAKE



Station 6



Segment 1209B, Station 11800, Finfeather Lake headwaters (2001).

BRYAN MUNICIPAL LAKE



Method 1



Method 2

BRYAN MUNICIPAL LAKE



Method 3



Segment 1209A, Station 11792, Bryan Municipal Lake near dam spillway, white pole identifies sample location (2001).

BRYAN MUNICIPAL LAKE



Station 1



Segment 1209A, Station 11793, Bryan Municipal Lake mainbody, white pole identifies sample location (2001).

BRYAN MUNICIPAL LAKE



Segment 1209A, Station 11794, Bryan Municipal Lake headwaters, white pole identifies sample location (2001).

**APPENDIX D
TOXICITY TESTS LAB REPORTS AND DATA SUMMARY**

**Assessment of the Presence and Causes of Ambient Toxicity in Texas Waterbodies
on the 1999 Clean Water Act 303(d) List to Support the Development
of Total Maximum Daily Loads**

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TNRCC Work Order No. 582-2-44844

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Introduction

Problem Definition

The Texas Natural Resource Conservation Commission (TNRCC) is responsible for administering provisions of the constitution and laws of the State of Texas to promote judicious use and the protection of the quality of waters in the State. A major aspect of this responsibility is the continuous monitoring and assessment of water quality to evaluate compliance with state water quality standards which are established within Texas Water Code, •26.023 and Title 30 Texas Administrative Code, ••307.1-307.10. Texas Surface Water Quality Standards 30 TAC 370.4(d) specify that surface waters will not be toxic to aquatic life. Pursuant to the federal Clean Water Act •303(d), states must establish total maximum daily loads (TMDLs) for pollutants contributing to violations of water quality standards. The purpose of this contract is to support the assessment of the presence and causes of ambient toxicity in seven Texas waterbodies on the 2000 Federal Clean Water Act •303(d) List in an effort to comply with Texas law.

Ambient toxicity testing complements routine chemical monitoring to identify waterbodies with aquatic life impairment. Since 1989, the TNRCC has collected approximately 600 ambient water samples and 330 sediment samples to test for toxicity to sensitive aquatic organisms that serve as surrogates for indigenous species. The U.S. Environmental Protection Agency Houston Laboratory has performed the toxicity testing by standard protocols. Based on this toxicity testing data, eight Texas waterbodies are identified on the 2000 CWA •303(d) list as impaired due to potential acute or chronic toxicity of ambient water and/or sediments. However, toxic effects to indigenous species in the natural systems have not been confirmed. Also, chemical toxicants or stressors responsible for the observed toxic effects in the laboratory have not yet been identified. Thus, the TNRCC needs a more thorough and intensive assessment of the existence of toxicity and identification of likely toxicants in several waterbodies. Based on the results of this assessment, the TNRCC may elect to remove a waterbody from the 303(d) list for toxicity, if evidence supports a conclusion that no toxicity is occurring in the waterbody, or to develop total maximum daily loads for identified toxicants or stressors.

UNT had responsibility to test water and/or sediments from the following five waterbodies of concern (Note that Vince Bayou and Arroyo Colorado Tidal testing were conducted by a separate laboratory and that Patrick Bayou was part of a different project):

1. Alligator Bayou (Segment 0702A) in Jefferson County (toxicity in water and sediment)
2. Bryan Municipal Lake (Segment 1209A) in Brazos County (toxicity in sediment)
3. Finfeather Lake (Segment 1209B) in Brazos County (toxicity in sediment)
4. Rio Grande (Segment 2304) in Kinney, Maverick, and Webb Counties (toxicity in water)
5. Rio Grande (Segment 2306) in Presidio County (toxicity in water).

Water and Sediment Testing on the Segments of Concern

Sediment and water samples were received from Parsons personnel and tested at the UNT/IAS Aquatic Toxicology Laboratory, Denton, TX, to determine acute and sublethal effects of exposure on four species of freshwater organisms. The criterion for effect was survival, although growth and reproduction were monitored, as appropriate. All raw data related to this study are stored at UNT. Data are presented as hard copy data files and also were supplied to Parsons ES in Excel worksheet format.

Materials and Methods

1. Aqueous and Sediment Testing.

Test Conditions

All standardized sediment and water bioassays followed USEPA guidelines for effluents (USEPA 1992). *Ceriodaphnia dubia* and *Pimephales promelas* 7-day tests were conducted at 25°C with 16:8 hour light: dark cycles at the Institute of Applied Sciences, Aquatic Toxicology Laboratory, University of North Texas. Temperature, dissolved oxygen, conductivity and pH were measured in each aqueous sample prior to daily renewals using YSI meters.

Ceriodaphnia dubia and *Pimephales promelas* were selected as test organisms for aqueous testing. Standardized whole sediment bioassays using *Chironomus tentans* and *Hyalella azteca* were selected for this study. *Ceriodaphnia dubia*, *Pimephales promelas*, and *Chironomus tentans* and *Hyalella azteca* are widely used in ambient and research testing of waterborne and sediment contaminants, respectively. In addition, an expansive literature exists for the relative sensitivities of each selected organism to numerous contaminants with different modes of toxicological action.

Statistical Analyses

ANOVA and Dunnett's multiple range tests were used to identify samples in which survival was statistically lower from the negative controls. The survival proportions were transformed using Arcsine transformation ($\sqrt{p_i}$), where p_i = proportion surviving in replicates. The data were then examined for homogeneity of variance and departure from normality using Bartlett's and Shapiro-Wilks tests, respectively. If the data were normally distributed and the variances homogenous, the transformed data were analyzed with a one-way ANOVA. If the F test of the ANOVA was significant ($p \leq 0.05$), differences between the mean of each sample were compared with the control using Dunnett's test. Dunnett's test is specifically intended to compare treatment means with a control. If the F test in the ANOVA is not significant, no further analysis is performed, and the sample means are then statistically similar to the control. When the assumptions of normality and variance homogeneity cannot be verified, Steel's Many One Rank Test is used to examine differences between the control and each mean. Steel's Test is specifically intended to examine differences between treatments and a control when assumptions of normality and variance homogeneity cannot be verified.

Test Material 1.: Aqueous Samples.

Water samples were obtained from Parsons ES. All samples were shipped in 48 quart coolers on ice. A chain of custody form was initiated at the time samples were obtained. Sample label information was recorded in the receiving log as was date received at UNT. Sample coolers were visually checked at arrival to UNT; all samples were on ice upon arrival. Samples were maintained at 4°C in a walk-in refrigerator prior to testing. Sample identification, date of receipt, date of testing, and holding time are summarized in Table 2.

Control Water

Reconstituted hard water (RHW) served as control water for all water toxicity tests. RHW was prepared in 50-L batches following procedures outlined by Knight & Waller (1987) with the following exceptions: 1) initial water used to prepare RHW was reverse-osmosis deionized water, 2) glass columns were packed with granular activated carbon obtained from Culligan Water Conditioning, and 3) the final solution was not bubbled with CO₂ but vigorously aerated for at least 24 h.

Test Organisms

To feed the invertebrates, *Selenastrum capricornutum* (Printz) was cultured in 50-ml glass screw-cap culture tubes, 2-L Erlenmeyer flasks, and 20-L polycarbonate carboys. Solid-media slant cultures were obtained from UTEX Culture Collection of Algae (University of Texas at Austin).

Algal cells were resuspended, and 1 ml was transferred aseptically to 3 or 4 50-ml culture tubes containing 15 ml sterile Gorham's medium [ATCC 1974] (Gorham's tubes) and capped with foam plugs. Gorham's tubes were placed on a wrist-arm shaker and allowed to incubate at 22° C for 4 to 7 days. A 24-h light source was provided by cool-white fluorescent bulbs such that the light intensity was approximately 1500 lux.

After incubation, 1 ml from each tube was used to inoculate an additional 3 or 4 Gorham's tubes. These were allowed to incubate for 7 days. This second set of Gorham's tubes were used to inoculate additional tubes and 2-L flasks. After inoculation of new tubes, the remaining algal suspension was poured aseptically into 2-L foam plugged flasks containing 1 L sterile AAP medium (ATCC 1984), and a stir bar. Flasks were placed on magnetic stir plates and incubated for 7 days. Incubation conditions were the same as for the Gorham's tubes. At the end of the incubation period, the contents of the flasks were poured into 20-L carboys containing 5 to 6 L sterile AAP medium. Carboys were incubated under the same conditions as described above. In addition, vigorous aeration was provided throughout incubation. An additional 6 L sterile AAP medium was added to each carboy at 2 and 4 d after inoculation. 25 ml vitamin suspension was also added to each carboy on the sixth day of incubation. The vitamin suspension was prepared by crushing one Centrum Silver multivitamin with a mortar and pestle and mixing the resulting powder in 100 ml distilled water. On the seventh day, carboys were capped and stored in the dark at 4EC until needed.

Ceriodaphnia dubia and *Pimephales promelas* used for standardized testing were obtained from permanent cultures at the Institute of Applied Sciences, Aquatic Toxicology Laboratory, University of North Texas. All *P. promelas* culture and testing procedures followed U.S. Environmental Protection Agency (USEPA 1994) recommendations. *Ceriodaphnia dubia* were cultured in standard synthetic RHW (USEPA 1991) without the addition of sodium selenate. *C. dubia* were mass cultured as described by Knight & Waller (1992) with the following modifications: 1) 500-ml culture jars contained 300 ml RHW, 2) mass cultures were fed 10 ml algae-Cerophyl suspension for the first 4 d, 3) mass cultures were initiated with less than 12-h-old neonates but not necessarily within 4 h of each other, and 4) fluorescent lights were not covered with dark plastic, hence light intensity in the test chamber was approximately 125 lux (Hemming, et al. 2002).

C. dubia received the same feeding suspension in both mass culture and during 7-d toxicity tests. Algal cells were retrieved from 20-L carboys by centrifugation. The supernatant (AAP medium) was discarded, and the remaining algal pellets were rinsed with RHW. Algal cells were finally resuspended in 500 to 600 ml RHW and counted using a hemocytometer. This algae concentrate was stored in the dark at 4°C until needed. The final feeding suspension consisted of a mixture of algae and Cerophyl and was prepared following procedures described by Knight and Waller (1992).

Seven day toxicity tests with *Ceriodaphnia dubia* were conducted following general procedures recommended by the U.S. Environmental Protection Agency (1994) except the yeast-cerophyl-trout chow feeding suspension was replaced by that described above (Hemming et al. 2002). Toxicity tests were initiated within 4 d of receiving samples. 15 ml water from each segment or RHW was poured into each of ten 30-ml polystyrene cups. 0.5 ml algae-Cerophyl feeding suspension was added and one < 24-h-old neonate was then placed in each cup. Following a random block design, neonates were transferred from cultures to exposure cups using an eyedropper. Cups were covered with glass plates to prevent evaporation.

Test Material 2 : Sediment Samples.

Sediment samples were collected by Parsons ES personnel and delivered to UNT by Federal Express couriers. A chain of custody form was initiated at the time samples were obtained. Sample label information was recorded in a chain of custody receiving log when received at UNT. Sample coolers were visually checked at arrival to UNT; all samples were on ice. All samples were contained in 3.5 gallon buckets. Samples were maintained at 4°C in a walk-in refrigerator prior to testing. Sample identification, date of receipt, date of testing, and holding time are summarized in Table 2.

Control Water

Dechlorinated tap water was used as overlying water for *Hyalella azteca* and *Chironomus tentans* cultures and whole sediment tests (USEPA 2000).

Test Organisms

Hyalella azteca and *Chironomus tentans* used for standardized testing were obtained from permanent cultures at the Institute of Applied Sciences, Aquatic Toxicology Laboratory, University of North Texas. UNT *H. azteca* were originally obtained from US Army Corps of Engineers Waterways Experiment Station, Vicksburg, MS. UNT *C. tentans* were originally obtained from Environmental Consulting and Testing, Superior, WI.

Test Conditions

All standardized sediment bioassays followed USEPA guidelines for whole sediments (USEPA 2000). *H. azteca* and *C. tentans* tests were conducted at 23°C with 16:8 hour light: dark cycles at the Institute of Applied Sciences, Aquatic Toxicology Laboratory, University of North Texas.

Sediment Preparation

Following USEPA recommendations (EPA 2000), sediments were not sieved to remove indigenous organisms before addition to beakers, however, large indigenous organisms and large debris were removed with forceps. On Day 1, sediment samples were homogenized using a stainless steel or Teflon spoon for five minutes. Once homogenized, 100 ml aliquots of sediment were placed in each 300 ml high-form lipless beaker. Eight replicate exposure chambers for each treatment were randomly assigned to a Zumwalt dilution box. After addition of sediment, 175 ml of dechlorinated tap water.

Addition of Organisms

Sediments samples were tested separately with *H. azteca* and *C. tentans*. On Day 0, 10 second-instar (about 10 days old) *C. tentans* larvae and 7 -14 day old *H. azteca* (1 - 2 day age range) organisms were introduced to replicate units under the air-water interface (EPA 2000).

Feeding

On Test Days 0 - 9, *H. azteca* and *C. tentans* were fed 1.0 ml of YCT (“Yeast-Cerophyll-Tetrafin” mix) and 1.5 ml of an aqueous solution of Tetrafin fish food, respectively (EPA 2000).

Renewal of Overlying Water

Approximately 1.5 volume additions per day of dechlorinated tap water were supplied to each beaker by a Mount-Brungs diluter and a Zumwalt delivery system (EPA 2000). Using YSI meters, temperature and dissolved oxygen were measure daily during testing for a randomly selected experimental unit.

Test Termination

Sediment tests were terminated following a 10-d exposure period. Experimental units were removed from Zumwalt boxes and test organisms recovered with sieves. *H. azteca* from each unit were rinsed with deionized water and placed on tared aluminum pans then dried at 60°C for 24 hours. Following 24 hours, dry weights were determined. *C. tentans* from each unit were

rinsed with deionized water and placed on tared aluminum pans then dried at 60°C for 24 hours. Following 24 hours, dry weights were determined. Dried *C. tentans* were subsequently oxidized at 550°C for 1 hour using a muffle furnace. Ashed aluminum pans were then re-weighed to determine somatic growth.

Reference Sediment (Negative Control)

All sediment tests were accompanied by a negative control reference sediment (control sediments). Negative control reference sediment was obtained by UNT personnel from the University of North Texas Water Research Field Station, Denton, TX. The principal reason for selecting this site as a suitable reference sediment is our knowledge of little previous anthropogenic activity, supported by analytical chemistry data from previous studies (e.g. Suedell et al. 1993). Additional chemical analysis indicated that these sediments were not contaminated.

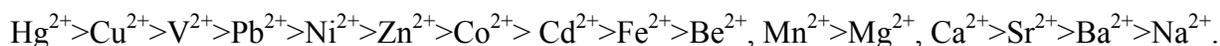
Reference Toxicant (Positive Control)

A positive control reference toxicant 48-hour test was conducted for each organism. Cadmium was selected as the reference toxicant because of extensive literature LC₅₀ values for each organism used in this study. *P. promelas* and *C. dubia* tests were conducted according to EPA guidelines (1992). *H. azteca* tests were conducted according to Steevens and Benson. LC₅₀s (95% conf. limits) for *H. azteca*, *P. promelas*, *C. dubia* were 18.8 ug/L (15.2, 22.0), 34.5 ug/L (29.4, 40.7), 36.7 ug/L (31.1, 43.1), respectively.

2. Sediment TIE.

U.S. EPA has not finalized sediment porewater or whole sediment Toxicity Identification Evaluation (TIE) methodology. Draft sediment TIE guidelines are available for porewaters and elutriates (EPA 1991) and closely follow effluent TIE procedures. Some whole sediment procedures for reducing toxicity of specific toxicant classes have been reported in the literature; however, whole sediment TIE procedures are not published in guideline format (Ho et al. 2002). Therefore, a tiered approach based on porewater tests was employed in this project (Ankley and Schubauer-Berigan 1995). Additional whole sediment TIE procedures were performed on Alligator Bayou and Fin Feather Lake sediments. Generally, 40-60% of sediment volume was isolated as pore water. *Ceriodaphnia dubia* was chosen for pore water testing because of test volume requirements. We also used *Hyalella azteca* and *Chironomus tentans* to test whole sediments.

All general porewater TIE procedures followed EPA (1991) draft guidelines. Whole sediment TIEs followed procedures previously reported in the peer-reviewed literature. In addition to draft EPA TIE procedures, we used three ion exchange media to remove organic or metal toxicants. The cation exchange resin SIR-300, a styrene and divinylbenzene copolymer with iminodiacetic functional group in the sodium form, was chosen for metal removal because of its ability to chelate heavy metal cations (ResinTech, New Berlin NJ). SIR-300 was previously suggested as an effective metal treatment in sediment TIE procedures (Burgess et al. 2000). SIR-300 affinity for metals is:



Although SIR-300 is a parallel TIE treatment to EDTA for divalent metals, we used SIR-300 in addition to EDTA because metals reduced by SIR-300 may be measured following TIE treatment. Because conventional TIE treatments are not effective for arsenic contaminated media, SIR-900, a synthetic aluminum oxide absorbent media specific for arsenic (arsenate and arsenite) and lead, was utilized in several TIE procedures for Fin Feather Lake sediment because of historic arsenic contamination (ResinTech, West Berlin NJ). C18 solid phase extraction columns, typically used in TIE procedures to remove organic contaminants, may also filter or remove other contaminants (e.g. metals) and complicate TIE interpretation. We chose Ambersorb 563, a carbonaceous adsorbent, for organic removal because it has 5 to 10 times the capacity of granular activated carbon. We used Ambersorb 563 in addition to C18 treatment in several TIEs to selectively remove organics without filtration complications. Ambersorb has been used to treat contaminated groundwater (EPA 1995) and lake water (Guzzella et al. 2002) and to remove organic contaminants in sediment TIE procedures (West et al. 2001). Appendix I provides a summary of tiered procedures we developed and followed for porewater and sediment TIEs.

Table 1. Assessment of Presence and Causes of Ambient Toxicity in Texas Waterbodies. University of North Texas, Institute of Applied Sciences. Water and sediment toxicity data summarized by station and test organisms. Mean and standard deviation statistics identify *Pimephales promelas*, *Chironomus tentans* and *Hyalella azteca* mortality (proportion surviving) and growth weights (mg), and *Ceriodaphnia dubia* mortality (percent survival) and reproduction (total number of neonates) endpoints. Statistical significant differences from control water or sediment were determined at $\alpha = 0.05$ and are identified by either Yes for a significant difference or No for a non-significant difference.

Table 1B. Segment 1209A: Bryan Municipal Lake; Segment 1209B: Fin Feather Lake, Brazos County, Texas.

| Segment | Event | Station | Matrix | Organism | Endpoint | Mean | S. D. | Sig. Effect (p=0.05) |
|---------|-------|---------|----------|-------------------|-----------|-------|-------|----------------------|
| 1209A | 1 | 11792 | Sediment | <i>C. tentans</i> | Growth | 0.455 | 0.102 | Yes |
| 1209A | 1 | 11793 | Sediment | <i>C. tentans</i> | Mortality | 0.313 | 0.270 | Yes |
| 1209A | 1 | 11794 | Sediment | <i>C. tentans</i> | Growth | 0.367 | 0.051 | Yes |
| 1209A | 1 | 11792 | Sediment | <i>H. azteca</i> | Growth | 0.091 | 0.018 | Yes |
| 1209A | 1 | 11793 | Sediment | <i>H. azteca</i> | Growth | 0.086 | 0.016 | Yes |
| 1209A | 1 | 11794 | Sediment | <i>H. azteca</i> | Growth | 0.115 | 0.016 | No |
| 1209A | 2 | 11792 | Sediment | <i>C. tentans</i> | Mortality | 0.663 | 0.169 | Yes |
| 1209A | 2 | 11793 | Sediment | <i>C. tentans</i> | Growth | 0.353 | 0.057 | No |
| 1209A | 2 | 11794 | Sediment | <i>C. tentans</i> | Growth | 0.389 | 0.110 | No |
| 1209A | 2 | 11792 | Sediment | <i>H. azteca</i> | Mortality | 0.838 | 0.106 | Yes* |
| 1209A | 2 | 11793 | Sediment | <i>H. azteca</i> | Growth | 0.109 | 0.009 | Yes |
| 1209A | 2 | 11794 | Sediment | <i>H. azteca</i> | Mortality | 0.900 | 0.053 | Yes* |
| 1209B | 1 | 11798 | Sediment | <i>C. tentans</i> | Mortality | 0.325 | 0.212 | Yes |
| 1209B | 1 | 11799 | Sediment | <i>C. tentans</i> | Growth | 1.122 | 0.327 | Yes |
| 1209B | 1 | 11800 | Sediment | <i>C. tentans</i> | Mortality | 0.463 | 0.245 | Yes |
| 1209B | 1 | 1209QA | Sediment | <i>C. tentans</i> | Mortality | 0.488 | 0.189 | Yes |
| 1209B | 1 | 11798 | Sediment | <i>H. azteca</i> | Mortality | 0.538 | 0.220 | Yes |
| 1209B | 1 | 11799 | Sediment | <i>H. azteca</i> | Growth | 0.107 | 0.011 | No |
| 1209B | 1 | 11800 | Sediment | <i>H. azteca</i> | Growth | 0.094 | 0.008 | Yes |
| 1209B | 1 | 1209QA | Sediment | <i>H. azteca</i> | Mortality | 0.627 | 0.127 | Yes |
| 1209B | 2 | 11798 | Sediment | <i>C. tentans</i> | Mortality | 0.225 | 0.237 | Yes |
| 1209B | 2 | 11799 | Sediment | <i>C. tentans</i> | Growth | 0.238 | 0.077 | No |
| 1209B | 2 | 11800 | Sediment | <i>C. tentans</i> | Growth | 0.303 | 0.144 | No |

| | | | | | | | | |
|-------|---|-------|----------|------------------|-----------|-------|-------|-----|
| 1209B | 2 | 11798 | Sediment | <i>H. azteca</i> | Growth | 0.091 | 0.021 | Yes |
| 1209B | 2 | 11799 | Sediment | <i>H. azteca</i> | Mortality | 0.800 | 0.107 | Yes |
| 1209B | 2 | 11800 | Sediment | <i>H. azteca</i> | Mortality | 0.563 | 0.200 | Yes |

*Although significant mortality effects were observed, *H. azteca* survival was 83.8% for 11792 test #2, 90% for 11794 test #2. QA, here and in the following tables, implies duplicate analysis for quality assurance on methods.

11792: Bryan Municipal Lake near Dam Spillway.

11793: Bryan Municipal Lake Mainbody.

11794: Bryan Municipal Lake Headwater.

11798: Finfeather Lake near Dam Spillway.

11799: Finfeather Lake Mainbody.

11800: Finfeather Lake Headwater.

Table 2. Chain of Custody Record. Assessment of Presence and Causes of Ambient Toxicity in Texas Waterbodies. University of North Texas, Institute of Applied Sciences.

| Segment | Event | Station | Matrix | Collect Date | Test Initiated | Hold Time Met |
|---------|-------|---------|----------|--------------|-------------------|---------------|
| 1209A | 1 | 11792 | Sediment | 04/19/2001 | 05/03/2001 | YES |
| 1209A | 1 | 11793 | Sediment | 04/19/2001 | 05/03/2001 | YES |
| 1209A | 1 | 11794 | Sediment | 04/19/2001 | 05/03/2001 | YES |
| 1209A | 2 | 11792 | Sediment | 05/21/2001 | 05/26, 06/13/2001 | YES |
| 1209A | 2 | 11793 | Sediment | 05/21/2001 | 05/26, 06/13/2001 | YES |
| 1209A | 2 | 11794 | Sediment | 05/21/2001 | 05/26, 06/13/2001 | YES |
| 1209B | 1 | 11798 | Sediment | 04/19/2001 | 05/03/2001 | YES |
| 1209B | 1 | 11799 | Sediment | 04/19/2001 | 05/03/2001 | YES |
| 1209B | 1 | 11800 | Sediment | 04/19/2001 | 05/03/2001 | YES |
| 1209B | 2 | 11798 | Sediment | 05/21/2001 | 05/26, 06/13/2001 | YES |
| 1209B | 2 | 11799 | Sediment | 05/21/2001 | 05/26, 06/13/2001 | YES |
| 1209B | 2 | 11800 | Sediment | 05/21/2001 | 05/26, 06/13/2001 | YES |

¹ Two dates correspond to initiation of *C. dubia* and *P. promelas* tests, respectively. Only *C. dubia* tests were performed following events 7 through 9.

11792: Bryan Municipal Lake near Dam Spillway.

11793: Bryan Municipal Lake Mainbody.

11794: Bryan Municipal Lake Headwater.

11798: Finfeather Lake near Dam Spillway.

11799: Finfeather Lake Mainbody.

11800: Finfeather Lake Headwater.

Results and Discussion

Ambient toxicity test results for the segments assessed during this project are detailed in Table 1. Table 1 provides summary data for each ambient toxicity test conducted on the segment, the matrix used (water or sediment), the organism tested, and the endpoint measured (mortality, growth, or reproduction). Each endpoint has an associated response, reported as the mean response, plus the standard deviation. For *Pimephales promelas*, *Chironomus tentans* and

Hyalella azteca, mortality was measured as proportion surviving. For *Ceriodaphnia dubia*, survivorship is measured as percentage survival. Growth for *Pimephales promelas*, *Chironomus tentans* and *Hyalella azteca* was measured as mean body weight (mg). Reproduction for *Ceriodaphnia dubia* was measured as total number of neonates produced per adult female during the 7-d test.

Survival data were used to calculate percent survival for each replicate. Mean and standard deviation were calculated for each sample. Statistical analyses were performed as defined above, with the exception of the *Ceriodaphnia* results, which were analyzed using Fishers Exact test (USEPA 1994).

Table 3. Sediment Toxicity Identification Evaluation Procedures.

Segment 1209, Fin Feather and Bryan Municipal Lakes

| Test Date | Test Type | Station | Organism | Effective TIE Treatment |
|-------------------------|-----------|---------|------------------|-------------------------|
| 13-23 July 2001 | Porewater | 11793 | <i>C. dubia</i> | EDTA |
| 13-23 July 2001 | Porewater | 11798 | <i>C. dubia</i> | None |
| 25 Aug. – 04 Sept. 2001 | Porewater | 11798 | <i>C. dubia</i> | S300, S900 |
| 02-12 February 2002 | Porewater | 11798 | <i>C. dubia</i> | S300, S900 |
| 10-20 March 2002 | Sediment | 11798 | <i>H. azteca</i> | S900* |
| 06-16 June 2002 | Porewater | 11798 | <i>C. dubia</i> | EDTA, S300 |
| 06-16 June 2002 | Porewater | 11800 | <i>C. dubia</i> | EDTA |

**H. azteca* growth not significantly different from control sediment. 60% survival in S900, 68.3% survival in control.

Table 1B; Segment 1209 A & 1209 B: Bryan Municipal Lake and Finfeather Lake.

Because sediments from stations within these two segments were consistently toxic to *Hyalella* and *Chironomus*, TIE procedures were initiated in July 2001 on station 11793 in segment 1209A and station 11798 in segment 11798. A summary of all TIEs conducted on this segment is provided in Table 3. In July 2001, sediment porewater was not acutely toxic; however, both station 11793 and 11798 porewaters reduced *C. dubia* reproduction. EDTA reduced toxicity of station 11793, but not 11798, porewaters (t-test, $p < 0.05$). A subsequent TIE was performed with several media selective for metals in August 2001 because of historic arsenic and metal contamination in segment 1209 A & B waterbodies. These media included: SIR-900 media selective for arsenic; TXI Shale, previously demonstrated to remove arsenic from aqueous solutions in sorption isotherm studies (F. Saleh, UNT, pers. comm.); and SIR-300. Station 11798 porewater was chosen for TIE treatments because of higher ambient sediment metal concentrations than other stations. Each TIE treatment improved *C. dubia* reproduction relative to untreated porewaters (t-test, $p < 0.05$). Because arsenic was suspected as a causative toxicant, total arsenic, arsenate and arsenite levels were measured in each TIE treatment. SIR-900 and

TXI Shale treatments reduced total arsenic, arsenate and arsenite porewater levels. Arsenic levels in porewaters were not sufficiently high enough to solely cause toxicity; total arsenic porewater concentration was 266 ug/L. Naddy et al. (1995) found that arsenic treatment up to 1.46 mg/L did not significantly affect *C. dubia* reproduction. Other metals may have been removed by SIR-900 and TXI Shale treatments; however, metal concentrations were not measured following this TIE.

In February 2002, another TIE was performed on station 11798 porewaters. Metal bioavailability and toxicity is reduced with increasing water hardness. Unlike previous TIEs where reconstituted hard water was used for porewater dilution, we used reconstituted moderately hard water for dilution in this TIE (APHA et al. 1995). Dilution water with lower hardness was chosen to maximize porewater metal bioavailability and toxicity, and potentially the effectiveness of TIE treatments. SIR-900 and SIR-900 + SIR-300 treatments significantly increased *C. dubia* reproduction at 100%, and 50% and 100% dilutions, respectively (t-test, $p < 0.05$). SIR-300 increased neonate production at 50% dilution, although not significantly, by an average of four offspring relative to baseline porewaters. Baseline porewater copper concentration was 722 ug/L, a value two orders of magnitude higher than previously reported lowest observed effect concentrations for *C. dubia* reproduction in laboratory water (Oris et al. 1991, Suedel et al. 1996). Lowest *C. dubia* fecundity was observed in 100% SIR-300 TIE treatments. Such a sub-lethal response was attributed to reduction of calcium and magnesium, essential nutrients, from porewaters.

Whole sediments from station 11798 were amended with SIR-300, SIR-900 and SIR-300 + SIR-900 at a 1:4, V:V ratio in March 2002 (Burgess et al. 2000). Following a 10-day exposure period, *H. azteca* growth was significantly improved, relative to reference sediment from the University of North Texas Water Research Field Station, by only SIR-900 treatments (t-test, $p < 0.05$).

Previous TIE procedures conducted in August 2001 and February 2002 identified that arsenic sediment porewater concentrations were not high enough to affect *C. dubia* reproduction. However, copper was measured at greater than 700 ug/L during the February 2002 TIE. SIR-300, the ion exchange resin reported to possess high selectivity for copper, dramatically reduced calcium and magnesium concentrations in the February study. We hypothesized that this reduction in Ca and Mg led to lower *C. dubia* fecundity because both metals are essential nutrients. In June 2002, a TIE was performed with station 11798 sediment porewaters. As with the February study, reconstituted moderately hard water served as dilution water. Following SIR-300 treatment, hardness was measured by titration (APHA et al. 1995), and Ca and Mg salts reintroduced to porewaters until hardness values returned to pre-SIR-300 levels. EDTA (3 mg/L) and SIR-300 treatments significantly improved survival and reproduction relative to baseline porewaters (t-test, $p < 0.05$). Because multiple metals were measured in porewaters, a toxic unit approach was taken to evaluate metal porewater toxicity (Tables 6 and 7). Interestingly, copper concentration (and toxic units) increased slightly from 11798 baseline porewaters and 11798 treated with SIR-300 (Table 6). However, concentrations of zinc, iron, lead and barium were decreased 26%, 32%, 37% and 96%, respectively, by SIR-300 treatments. Such a decrease in these metal concentrations and associated toxic units likely resulted in higher *C. dubia* survival and fecundity in SIR-300 treatments.

Bioavailability of these metals was clearly affected by compounds not accounted for in water hardness measures. For example, organic carbon binding to metals (Tipping and Hurley 1992) affects bioavailability and toxicity (Playle et al. 1993). Total organic carbon in these porewaters was measured at very high concentrations (baseline, 22.2 mg/L; SIR-300, 14.8 mg/L). EDTA treatment of 3 mg/L also improved *C. dubia* survival and fecundity in 11798 porewaters. Average neonate production was 2x higher in EDTA treatments than in SIR-300 treatments. Although the manufacturer of SIR-300 indicated that this resin is more selective for zinc, iron, lead and copper than calcium and magnesium, our data suggests that this resin preferentially removed calcium and magnesium. If the binding capacity of SIR-300 was exhausted by preferential binding of calcium and magnesium ions, ligands that bound calcium and magnesium in pretreated porewaters would be available for complexation with other divalent metals, specifically those metals measured at high enough levels (e.g. copper, to adversely affect *C. dubia* in 'clean' laboratory water toxicity tests. We are currently conducting a TIE study with SIR-300 to further remove metal contaminants from station 11798 porewaters. By increasing the V:V ratio of SIR-300 to porewater during TIE treatment, SIR300 metal binding capacity will be increased and should decrease total metal porewater concentrations. Following reintroduction of calcium and magnesium salts, effective in the June 2002 TIE with this porewater, reduced toxicity may be more clearly assigned to potentially causative toxicants. Contaminant addition procedures (Phase III TIE) will subsequently be performed to recreate porewater toxicity and provide confirmational information.

Table 6. Metal Chemistry and Toxic Units (in parentheses) of Finfeather Lake, Station 11798, Sediment Potewater Resin Toxicity Identification Evaluation. 7-Day *Ceriodaphnia dubia* test initiated June 6, 2002.

| Treatment | % Survival | Mean # Neonates | Aluminum | Arsenic | Barium | Cadmium | Calcium | |
|----------------------|------------|-----------------|-------------|--------------|--------------|--------------|--------------------|-----------|
| <u>Chromium</u> | | | | | | | | |
| 11798 100% Baseline | 20 | 0 | 707 (na) | 212 (1.12) | 225 (0.225) | <1.0 (<0.8) | 43500 11.2 (0.011) | |
| 11798 100% + SIR 300 | 100 | 8.6 | 1190 (na) | 260 (1.37) | <10 (<0.01) | <1.0 (<0.8) | 1940 15.8 (0.016) | |
| RMHW + SIR 300 | 100 | 23.2 | <100 (na) | <10 (<0.05) | <10 (<0.01) | <1.0 (<1.45) | 1200 <10 (<0.01) | |
| RMHW Control | 100 | 25.8 | <100 (na) | <10 (<0.05) | <10 (<0.01) | <1.0 (<1.11) | 15400 <10 (<0.01) | |
| ----- | | | | | | | | |
| Treatment | % Survival | Mean # Neonates | Copper | Iron | Lead | Magnesium | Nickel | Potassium |
| 11798 100% Baseline | 20 | 0 | 251 (16.5) | 1410 (1.41) | 24.2 (7.01) | 4910 | <10 (<0.05) | 4060 |
| 11798 100% + SIR 300 | 100 | 8.6 | 290 (20.2) | 955 (0.96) | 15.3 (4.81) | 1570 | <10 (<0.05) | 4670 |
| RMHW + SIR 300 | 100 | 23.2 | <10 (<1.26) | <100 (<0.10) | <3.0 (<2.27) | 1120 | <10 (<0.10) | 1760 |
| RMHW Control | 100 | 25.8 | <10 (<0.95) | <100 (<0.10) | <3.0 (<1.48) | 11000 | <10 (<0.07) | 1840 |
| ----- | | | | | | | | |
| Treatment | % Survival | Mean # Neonates | Selenium | Silver | Sodium | Zinc | Mercury | |
| 11798 100% Baseline | 20 | 0 | <10 (<2.0) | <2.0 (<0.32) | 87900 | 375 (2.91) | <0.20 (<0.15) | |
| 11798 100% + SIR 300 | 100 | 8.6 | <10 (<2.0) | <2.0 (<0.36) | 228000 | 276 (2.26) | <0.20 (<0.15) | |
| RMHW + SIR 300 | 100 | 23.2 | <10 (<2.0) | <2.0 (<1.18) | 101000 | <10 (<0.15) | <0.20 (<0.15) | |
| RMHW Control | 100 | 25.8 | <10 (<2.0) | <2.0 (<0.66) | 25100 | <10 (<0.11) | <0.20 (<0.15) | |

Metal concentrations (µg/L) are reported from one replicate.

Toxic units (in parentheses) are based on TNRCC or EPA chronic surface water quality criteria for aquatic life protection. Toxic units for 11798, 11798 + SIR 300, RMHW + SIR 300 and RMHW are based on, where appropriate, hardness values of 128, 120, 60 and 80 mg/L as CaCO₃, respectively.

¹SIR 300 = SIR-300 ion-exchange resin, ResinTech Inc., Cherry Hill, New Jersey.

²RMHW = Reconstituted Moderately Hard Water.

³EPA lists 1000 µg/L as the water quality criterion for the protection of human health. No water quality criteria for the protection of aquatic life are available. US Environmental Protection Agency. 1986. Quality Criteria for Water. EPA/440/5-86-001. US Environmental Protection Agency, Office of Water Regulations and Standards, Washington, DC.

⁴There is no chronic water quality criterion for aluminum.

⁵EPA lists 100 µg/L as the aquatic life protection criterion for total recoverable chromium. This value is used here because chromium measurements were not differentiated between Cr(III) and Cr(VI). US Environmental Protection Agency. 1980. Ambient Water Quality Criteria for Chromium. EPA/440/5-80-035. US Environmental Protection Agency, Office of Water Regulations and Standards, Criteria and Standards Division, Washington DC.

Table 7. Water quality criteria used in Finfeather Lake, station 11798, porewater chronic toxic units determination.

| Metal | 60 mg/L ¹ | Hardness (mg/L as CaCO ₃) | | 128 mg/L ⁴ | Source |
|----------|----------------------|---------------------------------------|-----------------------|-----------------------|--------------------|
| | | 80 mg/L ² | 120 mg/L ³ | | |
| Aluminum | NL ⁵ | NL | NL | NL | ----- |
| Arsenic | 190 | 190 | 190 | 190 | TNRCC ⁶ |
| Barium | 1000 | 1000 | 1000 | 1000 | EPA ⁷ |
| Cadmium | 0.690 | 0.899 | 1.19 | 1.25 | TNRCC |
| Chromium | 100 | 100 | 100 | 100 | EPA ⁸ |
| Copper | 7.94 | 10.58 | 14.35 | 15.17 | TNRCC |
| Iron | 1000 | 1000 | 1000 | 1000 | EPA ⁷ |
| Lead | 1.32 | 2.02 | 3.18 | 3.45 | TNRCC |
| Mercury | 1.3 | 1.3 | 1.3 | 1.3 | EPA ⁷ |
| Nickel | 102 | 136 | 183 | 194 | TNRCC |
| Selenium | 5.0 | 5.0 | 5.0 | 5.0 | EPA ⁷ |
| Silver | 1.69 | 3.01 | 5.55 | 6.20 | EPA ⁷ |
| Zinc | 67.8 | 90.1 | 122 | 129 | TNRCC |

¹Reconstituted moderately hard water (RMHW) after treatment with SIR 300 and calcium and magnesium reintroduced.

²RMHW.

³Station 11798 porewater after treatment with SIR 300 and calcium and magnesium reintroduced.

⁴Station 11798 porewater.

⁵No Listing.

⁶Texas Natural Resources Conservation Commission. 2000. Chapter 307: Texas Surface Water Quality Standards.

⁷US Environmental Protection Agency. 1986. Quality Criteria for Water. EPA/440/5-86-001. US Environmental Protection Agency, Office of Water Regulations and Standards, Washington, DC. For Barium, EPA lists 1000 µg/L as the water quality criterion for the protection of human health. No water quality criteria for the protection of aquatic life are available.

⁸EPA lists 100 µg/L as the aquatic life protection criterion for total recoverable chromium. This value is used here because chromium measurements were not differentiated between Cr(III) and Cr(VI). US Environmental Protection Agency. 1980. Ambient Water Quality Criteria for Chromium. EPA/440/5-80-035. US Environmental Protection Agency, Office of Water Regulations and Standards, Criteria and Standards Division, Washington DC.

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Appendix I. Sediment porewater TIE tiered procedures.

A. Pore Water Testing

Sample preparation

Centrifuge @ 7,500 to 10,000 xG for 30 min under refrigeration (4° C); decant pore water; no filtration.

Tiered Phase 1

Tier I: Initial Test

Initial test to confirm and define toxicity of pore water

Treatment: 0, 6.25, 12.5, 25, 50, 100% sample

Organism: *C. dubia*

Duration: up to 7 days

Tier II:

Standard Procedures:

Baseline toxicity

Treatment w/ EDTA (2 concentration levels) to chelate metals

Treatment w/ sodium thiosulfate (2 concentration levels)

Filtration with glass fiber filter (GFF), and post treatment analysis.

C₁₈-Solid Phase Extraction following Filtration to remove organics, and post treatment analysis.

Tier III:

Additional Procedures:

SIR-300 cationic resin for cationic metal chelation and post-treatment metals analysis

SIR-900 resin for removal of arsenic; post-treatment chemical analysis

Ambersorb 563 for organic removal without metal filtration and post-treatment metals analysis

B. Whole Sediment Testing

Whole-sediment toxicity reduction procedures:

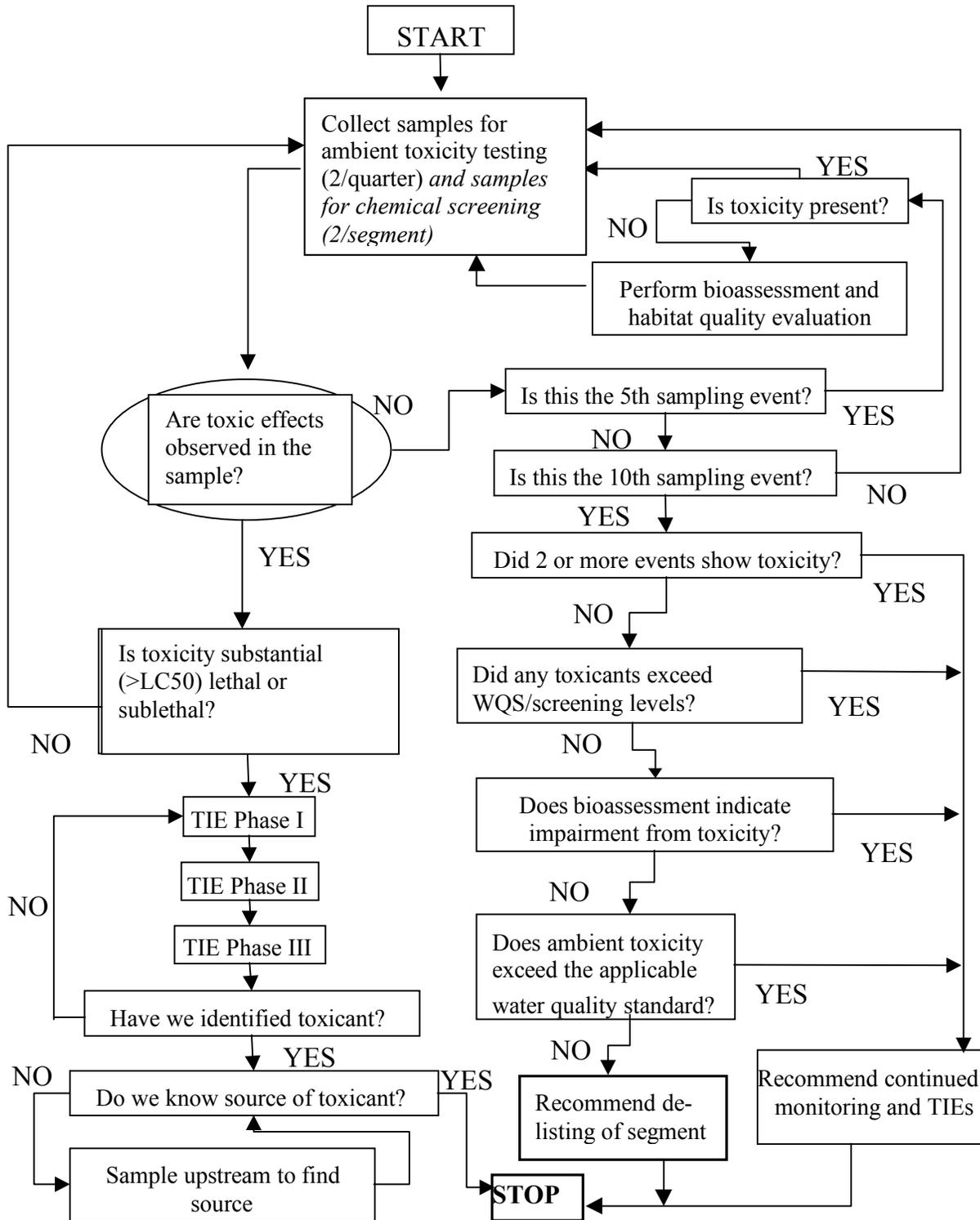
SIR-300 for cationic metal removal

SIR-900 for arsenic removal

Ambersorb 563 to remove organics

Coconut charcoal to absorb non-polar organics

Figure 1: Conceptual Toxicity Strategy flow diagram



APPENDIX II: PHASE III TIE FOR FINFEATHER LAKE

Finfeather Lake Tier 3 Testing

Date: 4/2/2003

PHASE III TIE REPORT

to:

Mr. Randy M. Palachek,
Parsons Engineering Services, Inc.
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Austin, TX 78754

University of North Texas

PI: T.W. La Point

**Project Title: Assessment of the Presence and Causes of Ambient Toxicity in Texas
Waterbodies on the 1999 Clean Water Act 303(d) List to Support the Development of
Total Maximum Daily Loads**

Accomplishments:

Our previous results for this site (Finfeather Lake, Station 11798) identified arsenic and other metals as probable toxins. Previous arsenic chemistry indicated that SIR900 resin treatment removed As; however, As concentrations were lower than would be expected to affect *Ceriodaphnia dubia* toxicity. To further investigate this relationship, we used treatments of SIR 900 (specific for arsenic), SIR 300 (specific for divalent metals), and a combination of both, as well as EDTA. Per previous discussions, we used reconstituted moderately hard water instead of hard water to try to 'increase' toxicity (e.g., increase the bioavailability of metals). Compared to the previous 1209-11798 porewater TIE (performed with hard water), toxicity increased slightly.

For the TIEs conducted in February and June, 2002, note that, relative to 50% baseline fecundity, neonate production was improved in SIR 300 and SIR 300+900 treatments, but not in 50% SIR 900 treatments. These treatment designations refer to 50% baseline, treated with a 20% (1:4) water volume ratio. This suggests that neither divalent metals nor arsenic in Finfeather Lake sediment porewaters are affecting *C. dubia* reproduction. This is supported by a decrease in copper

APPENDIX II: PHASE III TIE FOR FINFEATHER LAKE

Finfeather Lake Tier 3 Testing

Date: 4/2/2003

concentrations in SIR 300, SIR 900, and SIR 300+900 treatments. Low fecundity at 100% SIR300 treatments is likely due to the resin reducing Mg and Ca content (not surprising since SIR300 works on all divalents) and a relatively high pH of 8.9 and tht SIR 300 adds sodium ions to the water.

The February 2002 tests with porewater from Station 11798 (Table II-1, "Porewater Resin TIE," June 2002 repeat (Table II-2, "Porewater Phase III TIE"), the December 2002 test (Table II-3, "Porewater Addition Study") and the chemical analyses conducted on sediment porewaters (Table II-4, "Metal Concentrations") indicate that there is high variance in chemistry metals residues from separate sampling dates and a fairly large variance in *Ceriodaphnia dubia* reproductive responses. Our results indicate that organics are probably not a problem in these sediments and that arsenic and lead also do not present a problem. The question comes in as to whether zinc and copper are a problem. We think the evidence lies in this direction, probably more with Cu than with Zn. However, the distinction between Cu and Zn is based on a "toxic unit" approach, which may not yield precise information on which specific ion was "most" damaging to reproduction.

Whole sediment exposures demonstrated toxicity to *H. azteca* and *C. tentans*. Comparing measured whole sediment metal concentrations with sediment quality screening threshold effect levels (TELS) (Table II-5) indicated that toxicity is most likely due to metals. Measured concentrations of arsenic, cadmium, chromium, copper, lead, nickel, and zinc exceeded TELs in the whole sediment. Pore water toxicity testing was then conducted to confirm whole sediment results. Finfeather Lake porewaters were shown to be chronically toxic and occasionally acutely toxic to *C. dubia* in 7-day static renewal testing. A TIE approach using traditional methods (i.e. EDTA) as well as treatment with SIR 300 and SIR 900 resins was performed. Arsenic was eliminated as the primary contaminant of concern because despite a reduction in total As from 266 ug/L to 11.4 ug/L by SIR 900 (specific for arsenic), no improvement in *C. dubia* reproduction was observed (Table II-2). Subsequent metal analyses and toxic units determination indicated copper, lead and zinc as concerns (see Table 6).

APPENDIX II: PHASE III TIE FOR FINFEATHER LAKE

Finfeather Lake Tier 3 Testing

Date: 4/2/2003

Phase III TIE procedures were conducted such that porewaters were first treated with SIR 300 followed by reintroduction of copper, lead or zinc at nominal concentrations (previously determined measured concentrations; Table II-5). Results (Table II-6) indicated copper and zinc as the primary concerns. A toxic units approach suggests copper as a greater concern than zinc, however, 100% mortality was observed in treated porewaters in which zinc had been reintroduced.

Appended to this report are six Excel Tables with porewater results from Station 11798:

Table II-1, February 2002, "Porewater Resin TIE."

Table II-2, June 2002 repeat ("Porewater Resin/EDTA TIE")

Table II-3, December 2002 "Porewater Addition Study."

Table II-4, "Metal Concentrations," the chemical analyses conducted on sediment porewaters.

Table II-5. Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices, in $\mu\text{g}/\text{kg}$ sediment.

Table II-6. Mean number of neonates produced during a 7-day *Ceriodaphnia dubia* reproduction study with Finfeather Lake (11798) porewater and resins.

APPENDIX II: PHASE III TIE FOR FINFEATHER LAKE

Finfeather Lake Tier 3 Testing

Date: 4/2/2003

Table II-1, February 2002, "Porewater Resin TIE."

Finfeather Lake Station 11798 Porewater Resin TIE, 7-day Resin test initiated 2/2/02.

Porewater dilutions with Reconstituted Moderately Hard Water (RMHW), Samples collected on 10/30/01

C. *dubia* 7-day Reproduction

| Treatment | Neonates | | | | | Mean | Stdev |
|-----------------------|----------------------|------------------|------------------|----------------------|------------------|------|--------|
| | Rep1 #/fe male | Rep2 #/female | Rep3 #/female | Rep4 #/femal e | Rep5 #/female | | |
| RHW (control) | 21 | 27 | 19 | 23 | 25 | 23 | 3.1623 |
| RMHW (control) | 24 | 18 | 25 | 22 | 27 | 23.2 | 3.4205 |
| RMHW SIR300 | 21 | 26 | 23 | 25 | 27 | 24.4 | 2.4083 |
| RMHW SIR900 | 12 | 15 | 10 | 11 | 16 | 12.8 | 2.5884 |
| 11798 Baseline 25% | 22 | 21 | 20 | 24 | 23 | 22 | 1.5811 |
| 11798 Baseline 50% | 15 | 15 | 20 | 18 | . | 17 | 2.4495 |
| 11798 Baseline 100% | 12 | 10 | 11 | 12 | 12 | 11.4 | 0.8944 |
| 11798 SIR900 25% | 17 | 23 | 24 | 25 | 21 | 22 | 3.1623 |
| 11798 SIR900 50% | 16 | 18 | 17 | 17 | 20 | 17.6 | 1.5166 |
| 11798 SIR900 100% | 15 | 15 | 15 | 16 | 15 | 15.2 | 0.4472 |
| 11798 SIR300 25% | 19 | 25 | 20 | 23 | 19 | 21.2 | 2.6833 |
| 11798 SIR300 50% | 25 | 17 | 24 | 19 | 20 | 21 | 3.3912 |
| 11798 SIR300 100% | 2 | 0 | 6 | 3 | 0 | 2.2 | 2.49 |
| 11798 SIR300+900 25% | 22 | 25 | 22 | 20 | 26 | 23 | 2.4495 |
| 11798 SIR300+900 50% | 20 | 23 | 23 | 23 | 24 | 22.6 | 1.5166 |
| 11798 SIR300+900 100% | 16 | 20 | 12 | 13 | 14 | 15 | 3.1623 |

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| Metals (ug/L) | | | | | | | | | | Metals (mg/L) | | | | | | | | | |
|---------------|-------------|--------|-------|------------|--------|-------|------------|----|----|---------------|----|----|-------|-------|--------|--------|--|--|--|
| Al | As | Ba | Cr | Cu | Fe | Pb | Zn | Ni | Se | Ag | Cd | Hg | Na | Ca | Mg | K | | | |
| ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 25.4 | 13.6 | 11.6 | 1.84 | | | |
| 270 | 26.5 | 62 | 3.95 | 180.75 | 220 | 3.25 | 68 | | | | | | 39.25 | 9.5 | 1.1875 | 1.135 | | | |
| 540 | 53 | 124 | 7.9 | 361.5 | 440 | 6.5 | 136 | | | | | | 78.5 | 19 | 2.375 | 2.27 | | | |
| 1080 | 106 | 248 | 15.8 | 723 | 880 | 13 | 272 | ND | ND | ND | ND | ND | 157 | 38 | 4.75 | 4.54 | | | |
| 1070 | 23.1 | 17.65 | 2.775 | 57.5 | 191.75 | 5.475 | 96.25 | | | | | | 30.25 | 3.55 | 0.6625 | 0.845 | | | |
| 2140 | 46.15 | 35.3 | 5.55 | 115 | 383.5 | 10.95 | 192.5 | | | | | | 60.5 | 7.1 | 1.325 | 1.69 | | | |
| 4280 | 92.3 | 70.6 | 11.1 | 230 | 767 | 21.9 | 385 | ND | ND | ND | ND | ND | 121 | 14.2 | 2.65 | 3.38 | | | |
| 229.5 | 40 | 11.575 | 3.4 | 74.25 | 238.5 | 6.9 | 120 | | | | | | 48 | 0.75 | 0.1115 | 1.0625 | | | |
| 459 | 80 | 23.15 | 6.8 | 148.5 | 477 | 13.8 | 240 | | | | | | 96 | 1.49 | 0.223 | 2.125 | | | |
| 918 | 160 | 46.3 | 13.6 | 297 | 954 | 27.6 | 480 | ND | ND | ND | ND | ND | 192 | 2.98 | 0.446 | 4.25 | | | |
| 1100 | 7.475 | ND | ND | 46.75 | 127.25 | 3.825 | 72.75 | | | | | | 19.75 | 0.11 | 0.094 | 0.248 | | | |
| 2200 | 14.95 | ND | ND | 93.5 | 254.5 | 7.65 | 145.5 | | | | | | 39.5 | 0.22 | 0.188 | 0.496 | | | |
| 4400 | 29.9 | ND | ND | 187 | 509 | 15.3 | 291 | ND | ND | ND | ND | ND | 79 | 0.449 | 0.376 | 0.992 | | | |

*Italicized values are derived from 100% baseline values and assume 50% dilution with RMHW.

Table II-2, June 2002 repeat (“Porewater Resin/EDTA TIE”)

Finfeather Lake Station 11798 Porewater Resin/EDTA TIE

7-day *C. dubia* test initiated 6/6/02.

Porewater dilutions with Reconstituted Moderately Hard Water (RMHW)

| Treatment | Rep1 #/female | Rep2 #/female | Rep3 #/female | Rep4 #/female | Rep5 #/female | Mean | Stdev | % Survival |
|------------------------|------------------|------------------|------------------|------------------|------------------|------|----------|------------|
| RHW (control) | 21 | 24 | 22 | 20 | 28 | 23 | 3.162278 | 100 |
| RMHW (control) | 24 | 22 | 27 | 26 | 30 | 25.8 | 3.03315 | 100 |
| RMHW EDTA 3mg/l | 29 | 22 | 23 | 25 | 26 | 25 | 2.738613 | 100 |
| RMHW SIR300 25% | 30 | 30 | 17 | 23 | 25 | 25 | 5.43139 | 100 |
| RMHW SIR300 50% | 28 | 26 | 25 | 30 | 29 | 27.6 | 2.073644 | 100 |
| RMHW SIR300 100% | 26 | 9 | 30 | 28 | 23 | 23.2 | 8.348653 | 100 |
| 11798 Baseline 25% | 25 | 29 | 26 | 26 | 24 | 26 | 1.870829 | 100 |
| 11798 Baseline 50% | 19 | 20 | 19 | 0 | 19 | 15.4 | 8.619745 | 100 |
| 11798 Baseline 100% | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 20 |
| 11798 SIR300 25% | 29 | 31 | 24 | 17 | 30 | 26.2 | 5.80517 | 100 |
| 11798 SIR300 50% | 28 | 12 | 22 | 22 | 22 | 21.2 | 5.761944 | 100 |
| 11798 SIR300 100% | 6 | 9 | 6 | 6 | 16 | 8.6 | 4.335897 | 100 |
| 11798 EDTA 3 mg/L 25% | 27 | . | 26 | 26 | 21 | 25 | 2.708013 | 100 |
| 11798 EDTA 3 mg/L 50% | 27 | 26 | 22 | 22 | 19 | 23.2 | 3.271085 | 100 |
| 11798 EDTA 3 mg/L 100% | 15 | 21 | 17 | 14 | 13 | 16 | 3.162278 | 100 |

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Table II-3, December 2002 “Porewater Addition Study.”

Finfeather Lake, 11798, porewater Addition Study

7-d Cerio test initiated 12/12/02

Sediments collected 10/30/02

Porewater collected 12/05, 09 and 10/02

SIR treatment 12/11/02

| Treatment | Rep 1 | Rep 2 | Rep 3 | Rep 4 | Rep 5 | Mean | 1 SD |
|--------------------------------|-------|-------|-------|-------|-------|------|------|
| RMHW | 27 | 26 | 26 | 27 | 31 | 27.4 | 2.07 |
| RMHW + SIR (20%) | 23 | 19 | 20 | 19 | 18 | 19.8 | 1.92 |
| RMHW + SIR (50%) | 24 | 16 | 22 | 22 | 21 | 21 | 3 |
| 11798 Baseline | 2 | 3 | 2 | 1 | D/0 | 2 | 0.82 |
| 11798 baseline + EDTA (3 mg/L) | 16 | 13 | 14 | 18 | 15 | 15.2 | 1.92 |
| 11798 baseline + EDTA (8 mg/L) | 15 | 12 | 14 | 19 | 20 | 16 | 3.39 |
| 11798 + SIR 300 (20%) | 5 | 4 | 3 | 5 | 5 | 4.4 | 0.89 |
| 11798 + SIR 300 (50%) | 14 | 17 | 16 | 14 | 19 | 16 | 2.12 |
| 11798 + SIR 300 + Copper | 4 | 5 | 5 | 2 | 5 | 4.2 | 1.3 |
| 11798 + SIR 300 + Zinc | D/0 | D/0 | D/0 | D/0 | D/0 | 0 | |
| 11798 + SIR 300 + Lead | 15 | 15 | 16 | 14 | 10 | 14 | 2.35 |

All 11798 + Metal treatments were treated with SIR 300 (50%) prior to metal additions

Nominal copper concentration = 256 µg/L

Nominal zinc concentration = 317 µg/L

Nominal lead concentration = 26.4 µg/L

Table II-4, “Metal Concentrations,” the chemical analyses conducted on sediment porewaters.

Finfeather Lake, 11798, Addition Study 2; Samples Collected 10/30/02

Metal treatments were treated with SIR 300 (50%0 prior to metals additions

| Metals (ug/L) | Baseline Total Analyzed 12/10/02 | Baseline Dissolved Analyzed 12/10/02 | % Dissolved |
|---------------|--|--|-------------|
| Aluminum | 1330 | 101 | 7.59 |
| Arsenic | 176 | 143 | 81.3 |
| Copper | 256 | 67.8 | 26.5 |
| Lead | 26.4 | 5.96 | 22.6 |
| Zinc | 317 | 90.1 | 28.4 |
| Metals (ug/L) | UNTIAS01 Baseline Total Analyzed 12/23/02 | UNTIAS02 Baseline Dissolved Analyzed 12/23/02 | % Dissolved |
| Aluminum | 2000 | ND (<100) | 0.00 |
| Arsenic | 208 | 181 | 87.0 |
| Copper | 280 | 37 | 13.2 |
| Lead | 23.3 | 4.01 | 17.2 |
| Zinc | 296 | 115 | 38.9 |
| Metals (ug/L) | UNTIAS03 SIR 300 (20%) Total Analyzed 12/23/02 | UNTIAS04 SIR 300 (20%) Dissolved Analyzed 12/23/02 | % Dissolved |
| Aluminum | 1640 | ND (<100) | 0.00 |
| Arsenic | 179 | 180 | 100.6 |
| Copper | 241 | 69.4 | 28.8 |
| Lead | 14.8 | 3.86 | 26.1 |
| Zinc | 183 | 62.3 | 34.0 |

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| Metals (ug/L) | UNTIAS05 SIR 300 (50%) Total Analyzed 12/23/02 | UNTIAS06 SIR 300 (50%) Dissolved Analyzed 12/23/02 | % Dissolved |
|---------------|--|--|-------------|
| Aluminum | 1460 | ND (<100) | 0.00 |
| Arsenic | 164 | 159 | 97.0 |
| Copper | 220 | 59 | 26.8 |
| Lead | 9.09 | ND (<3.0) | 0.00 |
| Zinc | 149 | 19.7 | 13.2 |

| Metal (ug/L) | UNTIAS07 Total Cu Analyzed 12/23/02 | UNTIAS08 Dissolved Cu Analyzed 12/23/02 | % Dissolved |
|--------------|--|--|-------------|
| Copper | 411 | 172 | 41.8 |

| Metal (ug/L) | UNTIAS09 Total Zn Analyzed 12/23/02 | UNTIAS10 Dissolved Zn Analyzed 12/23/02 | % Dissolved |
|--------------|--|--|-------------|
| Zinc | 350 | 42.7 | 12.2 |

| Metal (ug/L) | UNTIAS11 Total Pb Analyzed 12/23/02 | UNTIAS12 Dissolved Pb Analyzed 12/23/02 | % Dissolved |
|--------------|--|--|-------------|
| Lead | 31.6 | 5.03 | 15.9 |

Total Concentrations with Additions

| | |
|--------|-------------------------|
| Copper | 220 + 256 = 476 ug/L |
| Zinc | 149 + 317 = 466 ug/L |
| Lead | 9.09 + 26.4 = 35.5 ug/L |

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Percent Decrease between 20% SIR 300 and 50% SIR 300

| Metals (ug/L) | UNTIAS03 SIR 300 (20%) Total | UNTIAS05 SIR 300 (50%) Total | % Decrease |
|---------------|---------------------------------|---------------------------------|------------|
| Aluminum | 1640 | 1460 | 11.0 |
| Arsenic | 179 | 164 | 8.4 |
| Copper | 241 | 220 | 8.7 |
| Lead | 14.8 | 9.09 | 38.6 |
| Zinc | 183 | 149 | 18.6 |

| Metals (ug/L) | UNTIAS04 SIR 300 (20%) Dissolved | UNTIAS06 SIR 300 (50%) Dissolved | % Decrease |
|---------------|-------------------------------------|-------------------------------------|------------|
| Aluminum | ND (<100) | ND (<100) | 0.0 |
| Arsenic | 180 | 159 | 11.7 |
| Copper | 69.4 | 59 | 15.0 |
| Lead | 3.86 | ND (<3.0) | 0.0 |
| Zinc | 62.3 | 19.7 | 68.4 |

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Table II-5. Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices, in µg/kg sediment.

| Contaminant | Threshold Effects Level (TEL) |
|--------------------|--------------------------------------|
| Arsenic | 7,240 |
| Cadmium | 676 |
| Chromium | 52,300 |
| Copper | 18,700 |
| Lead | 30,240 |
| Nickel | 15,900 |
| Silver | 730 |
| Zinc | 124,000 |

Table II-6. Mean number of neonates produced during a 7-day *Ceriodaphnia dubia* reproduction study with Finfeather Lake (11798) porewater and resins.

| 7-day Resin test initiated 8/25/01. | <i>C. dubia</i> 7-day Reproduction | | | | | Mean # | Stdev |
|--|---------------------------------------|------|------|------|------|--------|----------|
| | Rep1 | Rep2 | Rep3 | Rep4 | Rep5 | | |
| Reconstituted Hard Water (control) | 22 | 26 | 27 | 22 | 24 | 24.2 | 2.280351 |
| 11798 Baseline 25% | 29 | 29 | 27 | 22 | 25 | 26.4 | 2.966479 |
| 11798 Baseline 50% | 21 | 18 | 17 | 21 | 18 | 19 | 1.870829 |
| 11798 Baseline 100% | 16 | 13 | 17 | 13 | 14 | 14.6 | 1.81659 |
| 11798 SIR-900 25% | 24 | 26 | 31 | 26 | 28 | 27 | 2.645751 |
| 11798 SIR-900 50% | 21 | 28 | 24 | 23 | 21 | 23.4 | 2.880972 |
| 11798 SIR-900 100% | 20 | 20 | 13 | 18 | 18 | 17.8 | 2.863564 |
| 11798 TXI Shale 25% | 26 | 25 | 27 | 27 | 26 | 26.2 | 0.83666 |
| 11798 TXI Shale 50% | 29 | 22 | 21 | 25 | 20 | 23.4 | 3.646917 |
| 11798 TXI Shale 100% | 18 | 16 | 16 | 16 | 16 | 16.4 | 0.894427 |
| 11798 SIR-300 25% | 28 | 27 | 29 | 23 | 22 | 25.8 | 3.114482 |
| 11798 SIR-300 50% | 28 | 27 | 29 | 23 | 23 | 26 | 2.828427 |
| 11798 SIR-300 100% | 3 | 0 | 0 | 0 | 0 | 0.6 | 1.341641 |

Segment 1209A, Bryan Municipal Lake. Three stations total. 11792: Bryan Municipal Lake near dam spillway. 11793: Bryan Municipal Lake mainbody. 11794: Bryan Municipal Lake headwater. All statistical analyses were performed using TOXSTAT and followed USEPA guidelines for whole effluent toxicity tests.

Sample Event 1. Survival and growth of *Chironomus tentans* in Ten-day Sediment Exposures Conducted April 23 - May 3, 2001.
Samples collected on April 19, 2001.

| Sample ID | Number Surviving | Percent Survival | Percent Survival | Standard Deviation | p Value | Statistical Difference | AFDW Growth (mg) | Mean Growth (mg) | Standard Deviation | p Value | Statistical Difference | |
|----------------|------------------|------------------|------------------|--------------------|---------|------------------------|------------------|------------------|--------------------|---------|------------------------|-----|
| WRFS (Control) | 7 | 70 | 81.25 | 13.56 | N/A | N/A | 0.63 | 0.70625 | 0.23 | 0.05 | N/A | |
| | 7 | 70 | | | | | 0.87 | | | | | |
| | 7 | 70 | | | | | 0.91 | | | | | |
| | 9 | 90 | | | | | 0.83 | | | | | |
| | 10 | 100 | | | | | 0.42 | | | | | |
| | 7 | 70 | | | | | 1.01 | | | | | |
| | 8 | 80 | | | | | 0.41 | | | | | |
| | 10 | 100 | | | | | 0.57 | | | | | |
| | 11792 | 5 | 50 | 68.75 | 21.00 | 0.05 | NO | 0.66 | 0.455 | 0.10 | 0.05 | YES |
| | | 8 | 80 | | | | | 0.51 | | | | |
| | 6 | 60 | | | | | 0.4 | | | | | |
| | 9 | 90 | | | | | 0.46 | | | | | |
| | 8 | 80 | | | | | 0.45 | | | | | |
| | 3 | 30 | | | | | 0.47 | | | | | |
| | 9 | 90 | | | | | 0.33 | | | | | |
| | 7 | 70 | | | | | 0.36 | | | | | |
| 11793 | 6 | 60 | 31.25 | 26.96 | 0.05 | YES | N/A | N/A | N/A | 0.05 | N/A | |
| | 8 | 80 | | | | | N/A | | | | | |
| | 1 | 10 | | | | | N/A | | | | | |
| | 1 | 10 | | | | | N/A | | | | | |
| | 0 | 0 | | | | | N/A | | | | | |
| | 3 | 30 | | | | | N/A | | | | | |
| | 3 | 30 | | | | | N/A | | | | | |
| | 3 | 30 | | | | | N/A | | | | | |
| 11794 | 6 | 60 | 71.25 | 13.56 | 0.05 | NO | 0.393 | 0.36725 | 0.05 | 0.05 | YES | |
| | 6 | 60 | | | | | 0.395 | | | | | |
| | 8 | 80 | | | | | 0.44 | | | | | |
| | 6 | 60 | | | | | 0.29 | | | | | |
| | 7 | 70 | | | | | 0.31 | | | | | |
| | 9 | 90 | | | | | 0.39 | | | | | |
| | 6 | 60 | | | | | 0.39 | | | | | |
| | 9 | 90 | | | | | 0.33 | | | | | |

Segment 1209B, Fin Feather Lake. Three stations total. 11798: Fin Feather Lake near dam spillway. 11799: Fin Feather Lake mainbody. 11800: Fin Feather Lake headwater. All statistical analyses were performed using TOXSTAT and followed USEPA guidelines for whole effluent toxicity tests.

Sample Event 1. Survival and growth of *Chironomus tentans* in Ten-day Sediment Exposures Conducted 7 - 17 May, 2001.

Samples collected on April 19, 2001.

| Sample ID | Number Surviving | Percent Survival | Percent Survival | Standard Deviation | p Value | Statistical Difference | AFDW Growth (mg) | Mean Growth (mg) | Standard Deviation | p Value | Statistical Difference |
|-----------------------|------------------|------------------|------------------|--------------------|---------|------------------------|------------------|------------------|--------------------|---------|------------------------|
| WRFS (Control) | 8 | 80 | 73.75 | 9.16 | N/A | N/A | 0.3186 | 0.2442 | 0.08 | 0.05 | N/A |
| | 7 | 70 | | | | | 0.2443 | | | | |
| | 7 | 70 | | | | | 0.0756 | | | | |
| | 6 | 60 | | | | | 0.2853 | | | | |
| | 9 | 90 | | | | | 0.247 | | | | |
| | 7 | 70 | | | | | 0.2989 | | | | |
| | 8 | 80 | | | | | 0.2802 | | | | |
| | 7 | 70 | | | | | 0.2037 | | | | |
| | 4 | 40 | 32.5 | 21.21 | 0.05 | 0.05 | N/A | N/A | N/A | 0.05 | N/A |
| | 2 | 20 | | | | | N/A | | | | |
| 11798 | 6 | 60 | | | | | N/A | | | | |
| | 1 | 10 | | | | | N/A | | | | |
| | 0 | 0 | | | | | N/A | | | | |
| | 3 | 30 | | | | | N/A | | | | |
| | 5 | 50 | | | | | N/A | | | | |
| | 5 | 50 | | | | | N/A | | | | |
| | 3 | 30 | 57.5 | 16.69 | 0.05 | 0.05 | 0.2223 | 0.2023125 | 0.05 | 0.05 | NO |
| | 6 | 60 | | | | | 0.1512 | | | | |
| | 4 | 40 | | | | | 0.2662 | | | | |
| | 6 | 60 | | | | | 0.2463 | | | | |
| 11799 | 7 | 70 | | | | | 0.1841 | | | | |
| | 8 | 80 | | | | | 0.1193 | | | | |
| | 5 | 50 | | | | | 0.1928 | | | | |
| | 7 | 70 | | | | | 0.2363 | | | | |
| | 5 | 50 | 46.25 | 24.46 | 0.05 | 0.05 | N/A | N/A | N/A | 0.05 | N/A |
| | 4 | 40 | | | | | N/A | | | | |
| | 3 | 30 | | | | | N/A | | | | |
| | 1 | 10 | | | | | N/A | | | | |
| | 3 | 30 | | | | | N/A | | | | |
| | 5 | 50 | | | | | N/A | | | | |
| 11800 | 8 | 80 | | | | | N/A | | | | |
| | 8 | 80 | | | | | N/A | | | | |
| | 5 | 50 | 48.75 | 18.85 | 0.05 | 0.05 | N/A | N/A | N/A | 0.05 | N/A |
| | 5 | 50 | | | | | N/A | | | | |
| | 4 | 40 | | | | | N/A | | | | |
| | 6 | 60 | | | | | N/A | | | | |
| | 2 | 20 | | | | | N/A | | | | |
| | 6 | 60 | | | | | N/A | | | | |
| | 3 | 30 | | | | | N/A | | | | |
| | 8 | 80 | | | | | N/A | | | | |
| QA 1209B | 5 | 50 | | | | | N/A | | | | |
| | 5 | 50 | | | | | N/A | | | | |
| | 4 | 40 | | | | | N/A | | | | |
| | 6 | 60 | | | | | N/A | | | | |
| | 2 | 20 | | | | | N/A | | | | |
| | 6 | 60 | | | | | N/A | | | | |
| | 3 | 30 | | | | | N/A | | | | |
| | 8 | 80 | | | | | N/A | | | | |
| | 5 | 50 | | | | | N/A | | | | |
| | 8 | 80 | | | | | N/A | | | | |

Sample Event 2. Survival and growth of *Chironomus tentans* in Ten-day Sediment Exposures Conducted June 13 - 23, 2001.
 Samples collected on May 21, 2001.

| Sample ID | Number Surviving | Percent Survival | Percent Survival | Standard Deviation | p Value | Statistical Difference | AFDW Growth (mg) | Mean Growth (mg) | Standard Deviation | p Value | Statistical Difference |
|----------------|------------------|------------------|------------------|--------------------|---------|------------------------|------------------|------------------|--------------------|---------|------------------------|
| WRF5 (Control) | 9 | 90 | 75 | 32.07 | N/A | N/A | 0.2414 | 0.2687 | 0.13 | 0.05 | N/A |
| | 7 | 70 | | | | | 0.2293 | | | | |
| | 0 | 0 | | | | | 0 | | | | |
| | 9 | 90 | | | | | 0.331 | | | | |
| | 9 | 90 | | | | | 0.3766 | | | | |
| | 9 | 90 | | | | | 0.2223 | | | | |
| | 7 | 70 | | | | | 0.3429 | | | | |
| | 10 | 100 | | | | | 0.4061 | | | | |
| 11798 | 1 | 10 | 22.5 | 23.75 | 0.05 | YES | N/A | N/A | N/A | 0.05 | N/A |
| | 1 | 10 | | | | | N/A | | | | |
| | 4 | 40 | | | | | N/A | | | | |
| | 1 | 10 | | | | | N/A | | | | |
| | 5 | 50 | | | | | N/A | | | | |
| | 0 | 0 | | | | | N/A | | | | |
| | 0 | 0 | | | | | N/A | | | | |
| | 6 | 60 | | | | | N/A | | | | |
| 11799 | 7 | 70 | 78.75 | 19.59 | 0.05 | NO | 0.3186 | 0.23815 | 0.08 | 0.05 | NO |
| | 7 | 70 | | | | | 0.3156 | | | | |
| | 5 | 50 | | | | | 0.2304 | | | | |
| | 10 | 100 | | | | | 0.1493 | | | | |
| | 8 | 80 | | | | | 0.2899 | | | | |
| | 6 | 60 | | | | | 0.2745 | | | | |
| | 10 | 100 | | | | | 0.2187 | | | | |
| | 10 | 100 | | | | | 0.1082 | | | | |
| 11800 | 10 | 100 | 68.75 | 25.88 | 0.05 | NO | 0.2066 | 0.3033125 | 0.14 | 0.05 | NO |
| | 8 | 80 | | | | | 0.1875 | | | | |
| | 10 | 100 | | | | | 0.2219 | | | | |
| | 4 | 40 | | | | | 0.3565 | | | | |
| | 6 | 60 | | | | | 0.1662 | | | | |
| | 3 | 30 | | | | | 0.605 | | | | |
| | 6 | 60 | | | | | 0.362 | | | | |
| | 8 | 80 | | | | | 0.3208 | | | | |

**APPENDIX E
CHEMICAL TESTS LAB REPORTS AND DATA SUMMARY**

**Appendix E
Sediment Chemistry
Bryan Municipal Lake
Segment 1209A**

| PARAMETER | Station ID 11793 | | Station ID 11792 | Station ID 11793 | Station ID 11794 | Lowest | UNITS |
|-----------|---------------------------|-------------------|-------------------|-------------------|-------------------|-------------------------------|--------------------|
| | 5/21/01 | 7/18/01 | 7/12/2002 | 7/12/2002 | 7/12/2002 | | |
| Ions | Chloride | 115 | 73.6 | 26.9 | 102 | 161 | mg/Kg-dry wt |
| | Sulfate | 183 | 66.6 | 75.6 | 81.6 | 278 | mg/Kg-dry wt |
| Metals | Aluminum | 14400 | 11300 | 4990 | 11500 | 17900 | mg/Kg-dry wt |
| | Arsenic | 57.6 | 95.8 | 17.8 | 90.2 | 141 | 7.24 mg/Kg-dry wt |
| | Barium | 156 | 149 | 70.5 | 143 | 262 | mg/Kg-dry wt |
| | Cadmium | 0.736 | 0.619 | 0.249 | 0.658 | 1.41 | 0.676 mg/Kg-dry wt |
| | Calcium | 6730 | 7950 | 4100 | 6060 | 20300 | mg/Kg-dry wt |
| | Chromium | 32.8 | 36.9 | 10.8 | 43.9 | 90.8 | 52.3 mg/Kg-dry wt |
| | Copper | 52.5 | 40.6 | 13.9 | 44 | 178 | 18.7 mg/Kg-dry wt |
| | Iron | 10000 | 8430 | 5210 | 10200 | 16300 | mg/Kg-dry wt |
| | Lead | 36.4 | 37.1 | 21.8 | 42.3 | 99.7 | 30.2 mg/Kg-dry wt |
| | Magnesium | 1930 | 1540 | 791 | 1440 | 3870 | mg/Kg-dry wt |
| | Nickel | ND | ND | ND | ND | ND | mg/Kg-dry wt |
| | Potassium | 1040 | 817 | 527 | 887 | 1410 | mg/Kg-dry wt |
| | Selenium | ND | ND | 2.76 | 5.21 | 6.4 | mg/Kg-dry wt |
| | Silver | ND | ND | ND | ND | ND | mg/Kg-dry wt |
| | Sodium | 1100 | 917 | 273 | 716 | 1320 | mg/Kg-dry wt |
| | Zinc | 227 | 183 | 67.5 | 215 | 799 | 124 mg/Kg-dry wt |
| | Mercury | ND | ND | ND | ND | ND | 0.13 mg/Kg-dry wt |
| PARAMETER | 5/21/01 RESULT | 7/18/01 RESULT | 7/12/02 RESULT | 7/12/02 RESULT | 7/12/02 RESULT | Lowest Screening Value* | UNITS |
| Volatiles | 1,1,1-Trichloroethane | ND | ND | ND | ND | 30 | µg/Kg-dry wt |
| | 1,1,2,2-Tetrachloroethane | ND | ND | ND | ND | 940 | µg/Kg-dry wt |
| | 1,1,2-Trichloroethane | ND | ND | ND | ND | 1257 | µg/Kg-dry wt |
| | 1,1-Dichloroethane | ND | ND | ND | ND | 27 | µg/Kg-dry wt |
| | 1,1-Dichloroethene | ND | ND | ND | ND | 31 | µg/Kg-dry wt |
| | 1,2-Dibromoethane | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | 1,2-Dichloroethane | ND | ND | ND | ND | 256 | µg/Kg-dry wt |
| | 1,2-Dichloropropane | ND | ND | ND | ND | 2075 | µg/Kg-dry wt |
| | 2-Chloroethyl vinyl ether | ND | ND | ND | ND | 9727 | µg/Kg-dry wt |
| | Benzene | ND | ND | ND | ND | 57 | µg/Kg-dry wt |
| | Bromodichloromethane | ND | ND | ND | ND | 7426 | µg/Kg-dry wt |
| | Bromoform | ND | ND | ND | ND | 650 | µg/Kg-dry wt |
| | Bromomethane | ND | ND | ND | ND | 18 | µg/Kg-dry wt |
| | Carbon disulfide | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Carbon tetrachloride | ND | ND | ND | ND | 225 | µg/Kg-dry wt |
| | Chlorobenzene | ND | ND | ND | ND | 413 | µg/Kg-dry wt |
| | Chloroethane | ND | ND | ND | ND | 7937 | µg/Kg-dry wt |
| | Chloroform | ND | ND | ND | ND | 22 | µg/Kg-dry wt |
| | Chloromethane | ND | ND | ND | ND | 432 | µg/Kg-dry wt |
| | cis-1,2-Dichloroethene | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | cis-1,3-Dichloropropene | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Dibromochloromethane | ND | ND | ND | ND | 8701 | µg/Kg-dry wt |
| | Ethylbenzene | ND | ND | ND | ND | 10 | µg/Kg-dry wt |
| | Hexachlorobutadiene | ND | ND | ND | ND | 11 | µg/Kg-dry wt |
| | m,p-Xylene | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Methyl tert-butyl ether | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Methylene chloride | ND | ND | ND | ND | 374 | µg/Kg-dry wt |
| | o-Xylene | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Tetrachloroethene | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Toluene | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | trans-1,2-Dichloroethene | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | trans-1,3-Dichloropropene | ND | ND | ND | ND | 230 | µg/Kg-dry wt |
| | Trichloroethene | ND | ND | ND | ND | 215 | µg/Kg-dry wt |
| | Vinyl chloride | ND | ND | ND | ND | 691 | µg/Kg-dry wt |

Appendix E
Sediment Chemistry
Bryan Municipal Lake
Segment 1209A

| PARAMETER | 5/21/01 RESULT | 7/18/01 RESULT | 7/12/02 RESULT | 7/12/02 RESULT | 7/12/02 RESULT | Lowest Screening Values* | UNITS |
|----------------------------------|-------------------|-------------------|-------------------|-------------------|-------------------|--------------------------------|--------------|
| Semi-Vol. 1,2,4-Trichlorobenzene | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 1,2-Dichlorobenzene | ND | ND | ND | ND | ND | 50 | µg/Kg-dry wt |
| 1,3-Dichlorobenzene | ND | ND | ND | ND | ND | 1664 | µg/Kg-dry wt |
| 1,4-Dichlorobenzene | ND | ND | ND | ND | ND | 110 | µg/Kg-dry wt |
| 2,4,5-Trichlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4,6-Trichlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4-Dichlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4-Dimethylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4-Dinitrophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4-Dinitrotoluene | ND | ND | ND | ND | ND | 293 | µg/Kg-dry wt |
| 2,6-Dinitrotoluene | ND | ND | ND | ND | ND | 10341 | µg/Kg-dry wt |
| 2-Chloronaphthalene | ND | ND | ND | ND | ND | 267345 | µg/Kg-dry wt |
| 2-Chlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2-Methylnaphthalene | ND | ND | ND | ND | ND | 20.2 | µg/Kg-dry wt |
| 2-Methylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2-Nitrophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 3,3'-Dichlorobenzidine | ND | ND | ND | ND | ND | 20603 | µg/Kg-dry wt |
| 4,6-Dinitro-2-methylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 4-Bromophenyl phenyl ether | ND | ND | ND | ND | ND | 1248 | µg/Kg-dry wt |
| 4-Chloro-3-methylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 4-Chlorophenyl phenyl ether | ND | ND | ND | ND | ND | 456209 | µg/Kg-dry wt |
| 4-Methylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 4-Nitrophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Acenaphthene | ND | ND | ND | ND | ND | 6.71 | µg/Kg-dry wt |
| Acenaphthylene | ND | ND | ND | ND | ND | 5.87 | µg/Kg-dry wt |
| Anthracene | ND | ND | ND | ND | ND | 46.85 | µg/Kg-dry wt |
| Benzo(a)anthracene | ND | ND | ND | ND | ND | 74.8 | µg/Kg-dry wt |
| Benzo(a)pyrene | ND | ND | ND | ND | ND | 88.8 | µg/Kg-dry wt |
| Benzo(b)fluoranthene | ND | ND | ND | ND | ND | 27372 | µg/Kg-dry wt |
| Benzo(g,h,i)perylene | ND | ND | ND | ND | ND | 720 | µg/Kg-dry wt |
| Benzo(k)fluoranthene | ND | ND | ND | ND | ND | 3600 | µg/Kg-dry wt |
| Bis(2-chloroethoxy)methane | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Bis(2-chloroethyl)ether | ND | ND | ND | ND | ND | 368 | µg/Kg-dry wt |
| Bis(2-chloroisopropyl)ether | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Bis(2-ethylhexyl)phthalate | ND | ND | ND | ND | ND | 182 | µg/Kg-dry wt |
| Butyl benzyl phthalate | ND | ND | ND | ND | ND | 900 | µg/Kg-dry wt |
| Chrysene | ND | ND | ND | ND | ND | 108 | µg/Kg-dry wt |
| Di-n-butyl phthalate | ND | ND | ND | ND | ND | 11000 | µg/Kg-dry wt |
| Di-n-octylphthalate | ND | ND | ND | ND | ND | 885363 | µg/Kg-dry wt |
| Dibenzo(a,h)anthracene | ND | ND | ND | ND | ND | 6.22 | µg/Kg-dry wt |
| Diethyl phthalate | ND | ND | ND | ND | ND | 200 | µg/Kg-dry wt |
| Dimethyl phthalate | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Fluoranthene | ND | ND | ND | ND | ND | 113 | µg/Kg-dry wt |
| Fluorene | ND | ND | ND | ND | ND | 19 | µg/Kg-dry wt |
| Hexachlorobenzene | ND | ND | ND | ND | ND | 22 | µg/Kg-dry wt |
| Hexachlorocyclopentadiene | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Hexachloroethane | ND | ND | ND | ND | ND | 1000 | µg/Kg-dry wt |
| Indeno[1,2,3-cd]pyrene | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Isophorone | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| N-Nitrosodi-n-propylamine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| N-Nitrosodiphenylamine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Naphthalene | ND | ND | ND | ND | ND | 34.6 | µg/Kg-dry wt |
| Nitrobenzene | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Pentachlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Phenanthrene | ND | ND | ND | ND | ND | 86.7 | µg/Kg-dry wt |
| Phenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Pyrene | ND | ND | ND | ND | ND | 153 | µg/Kg-dry wt |

Appendix E
Sediment Chemistry
Bryan Municipal Lake
Segment 1209A

| | | | | | | | | |
|-----------------------------|--------------------|--------|-------|-------|-------|-------|--------------|--------------|
| Triazines | Atrazine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Cyanazine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Metolachlor | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Simazine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Pest/PCBs | a-BHC | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Alachlor | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Aldrin | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | b-BHC | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Chlordane | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | d-BHC | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | 4,4'-DDD | 7.7 | J | ND | ND | ND | 1.2 | µg/Kg-dry wt |
| | 4,4'-DDE | ND | ND | ND | ND | ND | 2.1 | µg/Kg-dry wt |
| | 4,4'-DDT | ND | ND | ND | ND | ND | 1 | µg/Kg-dry wt |
| | Dicofol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Dieldrin | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Endosulfan | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Endosulfan sulfate | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Endrin | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | g-BHC (Lindane) | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Heptachlor | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Heptachlor epoxide | ND | ND | ND | ND | ND | 0.6 | µg/Kg-dry wt |
| | Methoxychlor | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Mirex | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1016 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1221 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1232 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1242 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1248 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| PCB-1254 | ND | ND | ND | ND | ND | | µg/Kg-dry wt | |
| PCB-1260 | ND | ND | ND | ND | ND | | µg/Kg-dry wt | |
| Toxaphene | ND | ND | ND | ND | ND | | µg/Kg-dry wt | |
| Organo-phosphorus Compounds | Chloropyrifos | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Demeton (Total) | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Diazinon | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Guthion | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Malathion | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Parathion | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Chlorinated Herbicides | 2,4,5-T | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | 2,4,5-TP (Silvex) | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | 2,4-D | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Carbamates | Carbaryl | NA | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Diuron | NA | ND | ND | ND | ND | | µg/Kg-dry wt |
| SEM | Cadmium | 0.47 | 0.34 | ND | ND | ND | | µmol/dry g |
| | Copper | 1.83 | ND | ND | ND | ND | | µmol/dry g |
| | Lead | 30.67 | 22.00 | 0.06 | 0.15 | 0.26 | | µmol/dry g |
| | Mercury | 0.0008 | J | ND | ND | ND | | µmol/dry g |
| | Nickel | 1.46 | ND | 0.08 | 0.19 | 0.26 | | µmol/dry g |
| | Silver | 0.46 | J | ND | NA | NA | | µmol/dry g |
| Zinc | 160.45 | 100 | 0.86 | 2.5 | 8.4 | | µmol/dry g | |
| Total Organic Carbon (TOC) | | 56000 | 42880 | 19400 | 42900 | 57900 | | mg/Kg |
| Acid Volatile Sulfide (AVS) | | 560 | 78 | 34.6 | 110 | 140 | | µmol/dry g |
| Grain Size | Gravel | NA | NA | NA | 0.0 | 0.0 | | % |
| | Sand | 13.7 | 24.5 | 38.7 | 12.6 | 1.9 | | % |
| | Silt | 39.2 | 34.5 | 33.2 | 42.4 | 48.4 | | % |
| | Clay | 47.1 | 41.0 | 23.5 | 44.9 | 49.7 | | % |

Notes:

* Criteria is

J- result is estimated

ND- result was Not Detected

mg/kg-dry = milligrams per kilogram dry weight

ug/kg-dry = microgram per kilogram dry weight

umol/dry g = microgram per mole per dry gram

% = percent

**Appendix E
Sediment Chemistry
Finfeather Lake
Segment 1209B**

| | | Station ID 11799 | Station ID 11798 | Station ID 11799 | Station ID 11798 | Station ID 11800 | | |
|---------------------------|---------------------------|---------------------|---------------------|---------------------|---------------------|---------------------|--------------------------------|--------------|
| PARAMETER | | 5/21/01 RESULT | 7/18/01 RESULT | 7/18/01 RESULT | 5/9/02 RESULT | 5/9/02 RESULT | Lowest Screening Values* | UNITS |
| Ions | Chloride | 35.3 | 64.5 | 33.7 | 57.5 | 142 | | mg/Kg-dry wt |
| | Sulfate | 72.6 | 57.5 | 78.9 | 149 | 289 | | mg/Kg-dry wt |
| Metals | Aluminum | 7870 | 12800 | 4680 | 6600 | 15000 | | mg/Kg-dry wt |
| | Arsenic | 58.5 | 196 | 28.8 | 79.2 | 160 | 7.24 | mg/Kg-dry wt |
| | Barium | 113 | 641 | 96.6 | 164 | 207 | | mg/Kg-dry wt |
| | Cadmium | 0.33 | 0.71 | ND | 0.454 | 0.96 | 0.676 | mg/Kg-dry wt |
| | Calcium | 10100 | 19500 | 7750 | 8460 | 82500 | | mg/Kg-dry wt |
| | Chromium | 26.3 | 95.4 | 16.9 | 34.7 | 46.8 | 52.3 | mg/Kg-dry wt |
| | Copper | 65.4 | 575 | 44.5 | 171 | 113 | 18.7 | mg/Kg-dry wt |
| | Iron | 6740 | 14700 | 4220 | 6670 | 13800 | | mg/Kg-dry wt |
| | Lead | 17.5 | 56.9 | 12.6 | 33.3 | 51.8 | 30.24 | mg/Kg-dry wt |
| | Magnesium | 1340 | 2520 | 924 | 1130 | 4470 | | mg/Kg-dry wt |
| | Nickel | 6.91 | 76.7 | ND | 21.7 | ND | 15.9 | mg/Kg-dry wt |
| | Potassium | 652 | 882 | 402 | 464 | 1370 | | mg/Kg-dry wt |
| | Selenium | ND | ND | ND | ND | ND | | mg/Kg-dry wt |
| | Silver | ND | ND | ND | ND | ND | 0.73 | mg/Kg-dry wt |
| | Sodium | 462 | 1040 | 333 | 366 | 1030 | | mg/Kg-dry wt |
| | Zinc | 241 | 1280 | 151 | 447 | 466 | 124 | mg/Kg-dry wt |
| Mercury | ND | ND | ND | ND | ND | 0.13 | mg/Kg-dry wt | |
| Volatiles | 1,1,1-Trichloroethane | ND | ND | ND | ND | ND | 30 | µg/Kg-dry wt |
| | 1,1,2,2-Tetrachloroethane | ND | ND | ND | ND | ND | 940 | µg/Kg-dry wt |
| | 1,1,2-Trichloroethane | ND | ND | ND | ND | ND | 1257 | µg/Kg-dry wt |
| | 1,1-Dichloroethane | ND | ND | ND | ND | ND | 27 | µg/Kg-dry wt |
| | 1,1-Dichloroethene | ND | ND | UJ | UJ | ND | 31 | µg/Kg-dry wt |
| | 1,2-Dibromoethane | ND | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | 1,2-Dichloroethane | ND | ND | ND | ND | ND | 256 | µg/Kg-dry wt |
| | 1,2-Dichloropropane | ND | ND | ND | ND | ND | 2075 | µg/Kg-dry wt |
| | 2-Chloroethyl vinyl ether | ND | ND | ND | ND | ND | 9727 | µg/Kg-dry wt |
| | Benzene | ND | ND | ND | ND | ND | 57 | µg/Kg-dry wt |
| | Bromodichloromethane | ND | ND | ND | ND | ND | 7426 | µg/Kg-dry wt |
| | Bromoform | ND | ND | ND | ND | ND | 650 | µg/Kg-dry wt |
| | Bromomethane | ND | ND | ND | ND | ND | 18 | µg/Kg-dry wt |
| | Carbon disulfide | ND | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Carbon tetrachloride | ND | ND | ND | ND | ND | 225 | µg/Kg-dry wt |
| | Chlorobenzene | ND | ND | ND | ND | ND | 413 | µg/Kg-dry wt |
| | Chloroethane | ND | ND | ND | ND | ND | 7937 | µg/Kg-dry wt |
| | Chloroform | ND | ND | ND | ND | ND | 22 | µg/Kg-dry wt |
| | Chloromethane | ND | ND | ND | ND | ND | 432 | µg/Kg-dry wt |
| | cis-1,3-Dichloropropene | ND | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Dibromochloromethane | ND | ND | ND | ND | ND | 8701 | µg/Kg-dry wt |
| | Ethylbenzene | ND | ND | ND | ND | ND | 10 | µg/Kg-dry wt |
| | Hexachlorobutadiene | ND | ND | ND | ND | ND | 11 | µg/Kg-dry wt |
| | m,p-Xylene | ND | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Methyl tert-butyl ether | ND | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Methylene chloride | ND | ND | ND | ND | ND | 374 | µg/Kg-dry wt |
| | o-Xylene | ND | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Tetrachloroethene | ND | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | Toluene | ND | ND | ND | ND | ND | ND | µg/Kg-dry wt |
| | trans-1,2-Dichloroethene | ND | ND | UJ | UJ | ND | ND | µg/Kg-dry wt |
| trans-1,3-Dichloropropene | ND | ND | ND | ND | ND | 230 | µg/Kg-dry wt | |
| Trichloroethene | ND | ND | ND | ND | ND | 215 | µg/Kg-dry wt | |
| Vinyl chloride | ND | ND | ND | ND | ND | 691 | µg/Kg-dry wt | |

Appendix E
Sediment Chemistry
Finfeather Lake
Segment 1209B

| PARAMETER | 5/21/01 RESULT | 7/18/01 RESULT | 7/18/01 RESULT | 5/9/02 RESULT | 5/9/02 RESULT | Lowest Screening Values* | UNITS |
|----------------------------------|-------------------|-------------------|-------------------|---------------|---------------|--------------------------------|--------------|
| Semi-Vol. 1,2,4-Trichlorobenzene | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 1,2-Dichlorobenzene | ND | ND | ND | ND | ND | 50 | µg/Kg-dry wt |
| 1,3-Dichlorobenzene | ND | ND | ND | ND | ND | 1664 | µg/Kg-dry wt |
| 1,4-Dichlorobenzene | ND | ND | ND | ND | ND | 110 | µg/Kg-dry wt |
| 2,4,5-Trichlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4,6-Trichlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4-Dichlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4-Dimethylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4-Dinitrophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2,4-Dinitrotoluene | ND | ND | ND | ND | ND | 293 | µg/Kg-dry wt |
| 2,6-Dinitrotoluene | ND | ND | ND | ND | ND | 10341 | µg/Kg-dry wt |
| 2-Chloronaphthalene | ND | ND | ND | ND | ND | 267345 | µg/Kg-dry wt |
| 2-Chlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2-Methylnaphthalene | ND | ND | ND | ND | ND | 20.2 | µg/Kg-dry wt |
| 2-Methylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 2-Nitrophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 3,3'-Dichlorobenzidine | ND | ND | ND | ND | ND | 20603 | µg/Kg-dry wt |
| 4,6-Dinitro-2-methylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 4-Bromophenyl phenyl ether | ND | ND | ND | ND | ND | 1248 | µg/Kg-dry wt |
| 4-Chloro-3-methylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 4-Chlorophenyl phenyl ether | ND | ND | ND | ND | ND | 456209 | µg/Kg-dry wt |
| 4-Methylphenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| 4-Nitrophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Acenaphthene | ND | ND | ND | ND | ND | 6.71 | µg/Kg-dry wt |
| Acenaphthylene | ND | ND | ND | ND | ND | 5.87 | µg/Kg-dry wt |
| Anthracene | ND | ND | ND | ND | ND | 46.85 | µg/Kg-dry wt |
| Benzo(a)anthracene | ND | ND | ND | ND | ND | 74.8 | µg/Kg-dry wt |
| Benzo(a)pyrene | ND | ND | ND | ND | ND | 88.8 | µg/Kg-dry wt |
| Benzo(b)fluoranthene | ND | 0.17 | J | ND | ND | 27372 | µg/Kg-dry wt |
| Benzo(g,h,i)perylene | ND | ND | ND | ND | ND | 720 | µg/Kg-dry wt |
| Benzo(k)fluoranthene | ND | ND | ND | ND | ND | 3600 | µg/Kg-dry wt |
| Bis(2-chloroethoxy)methane | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Bis(2-chloroethyl)ether | ND | ND | ND | ND | ND | 368 | µg/Kg-dry wt |
| Bis(2-chloroisopropyl)ether | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Bis(2-ethylhexyl)phthalate | ND | ND | ND | ND | ND | 182 | µg/Kg-dry wt |
| Butyl benzyl phthalate | ND | ND | ND | ND | ND | 900 | µg/Kg-dry wt |
| Chrysene | ND | 0.14 | J | ND | ND | 108 | µg/Kg-dry wt |
| Di-n-butyl phthalate | ND | ND | ND | ND | ND | 11000 | µg/Kg-dry wt |
| Di-n-octylphthalate | ND | ND | ND | ND | ND | 885363 | µg/Kg-dry wt |
| Dibenzo(a,h)anthracene | ND | ND | ND | ND | ND | 6.22 | µg/Kg-dry wt |
| Diethyl phthalate | ND | ND | ND | ND | ND | 200 | µg/Kg-dry wt |
| Dimethyl phthalate | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Fluoranthene | ND | 0.17 | J | ND | ND | 113 | µg/Kg-dry wt |
| Fluorene | ND | ND | ND | ND | ND | 19 | µg/Kg-dry wt |
| Hexachlorobenzene | ND | ND | ND | ND | ND | 22 | µg/Kg-dry wt |
| Hexachlorocyclopentadiene | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Hexachloroethane | ND | ND | ND | ND | ND | 1000 | µg/Kg-dry wt |
| Indeno[1,2,3-cd]pyrene | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Isophorone | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| N-Nitrosodi-n-propylamine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| N-Nitrosodiphenylamine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Naphthalene | ND | ND | ND | ND | ND | 34.6 | µg/Kg-dry wt |
| Nitrobenzene | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Pentachlorophenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Phenanthrene | ND | ND | ND | ND | ND | 86.7 | µg/Kg-dry wt |
| Phenol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Pyrene | ND | 0.16 | J | ND | ND | 153 | µg/Kg-dry wt |

Appendix E
Sediment Chemistry
Finfeather Lake
Segment 1209B

| PARAMETER | | 5/21/01 RESULT | 7/18/01 RESULT | 7/18/01 RESULT | 5/9/02 RESULT | 5/9/02 RESULT | Lowest Screening Values* | UNITS |
|-----------------------------|--------------------|-------------------|-------------------|-------------------|------------------|------------------|--------------------------------|--------------|
| Triazines | Atrazine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Cyanazine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Metolachlor | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Simazine | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Pest/PCBs | a-BHC | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Alachlor | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Aldrin | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | b-BHC | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Chlordane | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | d-BHC | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | 4,4'-DDD | 15.0 J | ND | ND | ND | ND | 1.2 | µg/Kg-dry wt |
| | 4,4'-DDE | 24.0 J | ND | ND | ND | ND | 2.1 | µg/Kg-dry wt |
| | 4,4'-DDT | 5.2 J | 3.6 J | ND | ND | ND | 1 | µg/Kg-dry wt |
| | Dicofol | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Dieldrin | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Endosulfan | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Endosulfan sulfate | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Endrin | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | g-BHC (Lindane) | ND | 3.3 J | ND | ND | ND | | µg/Kg-dry wt |
| | Heptachlor | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Heptachlor epoxide | ND | ND | ND | ND | ND | 0.6 | µg/Kg-dry wt |
| | Methoxychlor | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Mirex | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1016 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1221 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1232 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1242 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | PCB-1248 | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| PCB-1254 | ND | ND | ND | ND | ND | | µg/Kg-dry wt | |
| PCB-1260 | ND | ND | ND | ND | ND | | µg/Kg-dry wt | |
| Toxaphene | ND | ND | ND | ND | ND | | µg/Kg-dry wt | |
| Organo-phosphorus Compounds | Chloropyrifos | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Demeton (Total) | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Diazinon | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Guthion | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Malathion | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Parathion | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Chlorinated Herbicides | 2,4,5-T | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | 2,4,5-TP (Silvex) | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| | 2,4-D | ND | ND | ND | ND | ND | | µg/Kg-dry wt |
| Carbamates | Carbaryl | NA | ND | ND | ND | ND | | µg/Kg-dry wt |
| | Diuron | NA | ND | ND | ND | ND | | µg/Kg-dry wt |
| PARAMETER | | 5/21/01 RESULT | 7/18/01 RESULT | 7/18/01 RESULT | 5/9/02 RESULT | 5/9/02 RESULT | Lowest Screening Values* | UNITS |
| SEM | Cadmium | 0.29 | 0.28 | ND | ND | ND | | µmol/dry g |
| | Copper | 2.13 | ND | 1.2 | 4.5 J | 3.9 J | | µmol/dry g |
| | Lead | 20.68 | 22.00 | 6.7 | 0.11 | 0.34 | | µmol/dry g |
| | Mercury | 0.0006 J | ND | ND | ND | ND | | µmol/dry g |
| | Nickel | 1.71 | 4.2 | ND | 0.23 | 0.19 | | µmol/dry g |
| | Silver | 0.407 J | ND | ND | NA | NA | | µmol/dry g |
| | Zinc | 270.9 | 660 | 90 | 7.9 | 12 | | µmol/dry g |
| Total Organic Carbon (TOC) | | 23100 | 26430 | 14710 | 18000 | 51300 | | mg/Kg |
| Acid Volatile Sulfide (AVS) | | 1061 | 260 | 51 | 29.6 | 112 | | µmol/dry g |
| Grain Size | Gravel | NA | NA | NA | 0 | 0 | | % |
| | Sand | 44.63 | 41.06 | 70.95 | 28.9 | 0.6 | | % |
| | Silt | 30.09 | 20.5 | 14.17 | 49.2 | 68.5 | | % |
| | Clay | 25.28 | 38.44 | 14.88 | 21.90 | 30.9 | | % |

Notes:

* Criteria is from *Equilibrium and Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices* tables. The value is the lowest value from the Indices as stated in the Appendix.

J- result is estimated

ND- result was Not Detected

mg/kg-dry = milligrams per kilogram dry weight

ug/kg-dry = microgram per kilogram dry weight

umol/dry g = microgram per mole per dry gram

% = percent

**APPENDIX F
DATA QUALITY OBJECTIVES AND VALIDATION REPORTS**

Appendix F Data Quality Objectives for Measurement Data

| Parameter | Units | Method Type | Method | Method Description | Storet | MAL | Precision of Laboratory Duplicates (RPD) | Accuracy of Matrix Spikes % Recovery | Accuracy crm | Percent Complete |
|--|--|--|--|-----------------------------|----------------------------|------|--|--------------------------------------|--------------|------------------|
| Field Parameters | | | | | | | | | | |
| pH | pH units | YSI Multi-Parameter Probe | EPA 150.1 or TNRCC SOP | probe | 00400 | 1.0 | 10 | NA | +/- 0.1 | 90 |
| Dissolved Oxygen (DO) | mg/L | YSI Multi-Parameter Probe | EPA 360.1 or TNRCC SOP | probe | 00300 | 1.0 | 10 | +/- 0.5 | NA | 90 |
| Conductivity | uS/cm | YSI Multi-Parameter Probe | EPA 120.1 or TNRCC SOP | probe | 00094 | 1 | 10 | +/- 5 | +/- 5 | 90 |
| Temperature | ° Celcius | YSI Multi-Parameter Probe | EPA 170.1 or TNRCC SOP | probe | 00010 | NA | 10 | NA | NA | 90 |
| Salinity | ppt | YSI Multi-Parameter Probe | TNRCC SOP | probe | 00480 | NA | NA | NA | NA | 90 |
| Instantaneous Stream Flow | cfs | Flowmeter | TNRCC SOP | sensor | 00061 | NA | NA | NA | NA | 90 |
| Flow Severity | 1-no flow, 2-low, 3-normal, 4-flood, 5-high, 6-dry | Observation | TNRCC SOP | Field observation | 01351 | NA | NA | NA | NA | 90 |
| Conventional Parameters | | | | | | | | | | |
| Total Residual Chlorine | mg/L | DPD | EPA 330.5 | colorimetric | 50060 | 0.1 | 20% | NA | NA | 90 |
| Sediment Grain-size | % particle size | Frac. Separation & gravimetric determination | EPA 3.4, 3.5 (600/2-78-054) | Separation and gravimetric | 89991, 82009, 82008, 80256 | NA | NA | NA | NA | 90 |
| Total Suspended Solids | mg/L | gravimetric | EPA 160.2 | gravimetric | 00530 | 4.0 | 20 | NA | +/- 10% | 90 |
| Total Organic Carbon (TOC) | mg/L | oxidation | EPA 415.1 | oxidation | 00680 | 1.0 | 20 | 78-120 | +/- 10% | 90 |
| Total Organic Carbon (TOC) in sediment | mg/kg | Combustion | B&B Laboratories SOP 1005 See Appendix I | Combustion | 81951 | 0.3 | 15 | 80-120 | +/- 5% | 90 |
| Oil & Grease | mg/L | Extraction Gravimetry | EPA 413.1 | Freon Extractable Material | 00556 | 1.0 | 20 | 80-120 | +/-10% | 90 |
| Dissolved Organic Carbon (DOC) | mg/L | oxidation | EPA 415.2 | oxidation | 00681 | 0.1 | 20 | 78-120 | +/- 10% | 90 |
| Total Alkalinity, as CaCO ₃ | mg/L | potentiometric | EPA 310.1-2 | potentiometri ^c | 00410 | 3.0 | 20 | 78-120 | NA | 90 |
| Total Dissolved Solids (TDS) | mg/L | residue gravimetric | EPA 160.1 | residue gravimetric | 70300 | 10.0 | 20 | NA | NA | 90 |
| Sulfate in water | mg/L | ion chromatoph ^{gry} | EPA 300.0/9056 | IC | 00945 | 3 | 20 | 70-113 | +/- 10% | 90 |
| Sulfate in sediment | mg/kg | ion chromatoph ^{gry} | EPA 300.0/9056 | IC | 85818 | 10 | 30 | 80-120 | 80-120 | 90 |
| Sulfide in water | mg/L | colorimetric | EPA 371.2 | colorimetric | 00745 | 1.0 | 20 | 80-120 | +/-10% | 90 |
| Flouride in water | mg/L | colorimetric | EPA 340.3/9056 | Colorimetric/ ^{IC} | 00950 | 0.5 | 20 | 80-120 | +/-10% | 90 |

Appendix F Data Quality Objectives for Measurement Data

| Parameter | Units | Method Type | Method | Method Description | Storet | MAL | Precision of Laboratory Duplicates (RPD) | Accuracy of Matrix Spikes % Recovery | Accuracy crm | Percent Complete |
|--|--------|---|-------------------------|--|----------------|-------------------|--|--------------------------------------|--------------|------------------|
| Chloride in water | mg/L | colorimetric | EPA 325.2/9256 | Colorimetric automated ferricyanide/IC | 00940 | 1.0 | 20 | 80-120 | | 90 |
| Chloride in sediment | mg/kg | IC | EPA 300.0 | IC | 00943 | 10 | 30 | 80-120 | 80-120 | 90 |
| Ammonia-N | mg/L | colorimetric | EPA 350.1 | colorimetric | 00610 | 0.02 | 20 | 68-135 | NA | 90 |
| o-Phosphorus | mg/L | colorimetric, absorbic acid | EPA 365.3 | IC | 00671 | 0.01 | 20 | 80-120 | NA | 90 |
| Potassium, total recoverable in water | mg/L | ICP/AES | EPA 200.7 | ICP/AES | 00937 | 0.05 | 20 | 80-149 | 90-110 | 90 |
| Potassium in sediment | mg/kg | ICP/MS | EPA 6020 | ICP/MS | 00938 | 25 | 25 | NA | 80-120 | 90 |
| Sodium, total recoverable in water | mg/L | ICP/AES | EPA 200.7 | ICP/AES | 00929 | 0.2 | 20 | 79-137 | 90-110 | 90 |
| Sodium in sediment | mg/kg | ICP/MS | EPA 6020 | ICP/MS | 00934 | 25 | 25 | NA | 80-120 | 90 |
| Nitrate/nitrite-N | mg/L | ion chromatography | EPA 353.2 | Colorimetric automated cadmium reduction | 00630 | 0.01 | 20 | 83-125 | +/- 10% | 90 |
| Total Kjeldahl Nitrogen | mg/L | colorimetric, automated phenate | EPA 351.2 | colorimetric | 00625 | 0.1 | 20 | 72-133 | +/- 10% | 90 |
| Total Phosphorus (TPO) | mg/L | colorimetric, automated, block digester | 365.1-4 | colorimetric | 00665 | 0.02 | 20 | 74-118 | +/- 10% | 90 |
| Cyanide | mg/L | spectrophotometric | EPA 335.2 | spectrophotometric | 00720 | 5 | 20 | 80-120 | +/- 10% | 90 |
| Turbidity | NTU | nephelometric | EPA 180.1 | nephelometric | 82079 | 0.05 | 20 | NA | +/- 10% | 90 |
| Carbonaceous Biochemical Oxygen Demand (BOD) | mg/L | potentiometric | EPA 405.1 | potentiometric | 00307 | 1.0 | 25 | NA | +/- 5% | 90 |
| Chemical Oxygen Demand (COD) | mg/L | colorimetric | EPA 410.1-3 | colorimetric | 00335 or 00340 | 10 | 25 | NA | +/- 5% | 90 |
| Acid volatile sulfide in sediment | umol/g | colorimetry | EPA Draft 1991 | Purge and trap, colorimetry | 50088 | 0.5 | 40 | 60-130 | NA | 90 |
| SEM Simultaneous extraction, sum of concentrations: Cd, Cu, Pb, Hg, Ni, Ag, and Zn | umol/g | CVAAS Hg, ICP Other elements | EPA 200.7/245.5 | Purge and Trap, Atomic spectroscopy | 50087 | 0.05-0.5 w/ metal | 40 | NA | NA | 90 |
| Metals, trace metals, and related parameters | | | | | | | | | | |
| Aluminum, dissolved in water | µg/L | ICP-MS | EPA 200.8 | ICP-MS | 01106 | 10 | 25 | 80-120 | 80-120 | 90 |
| Aluminum, total in water | µg/L | ICP-MS | EPA 200.8 | ICP-MS | 01105 | 10 | 25 | 80-120 | 80-120 | 90 |
| Aluminum in sediment | mg/kg | Primary Direct | EPA 200.8 or 6010B/6020 | ICP-MS | 01108 | 12.5 | 25 | NA | 80-120 | 90 |
| Arsenic, dissolved in water | µg/L | HGA/FS | EPA 200.8 | HGA/FS | 01000 | 10 | 25 | 55-146 | 55-146 | 90 |
| Arsenic, total in water | µg/L | HGA/FS | EPA 1632 | HGA/FS | 01002 | 0.5 | 25 | 55-146 | 55-146 | 90 |
| Arsenic in sediment | mg/kg | Primary Direct | EPA 6020/200.8 | ICP-MS | 01003 | 2.5 | 25 | 80-120 | 80-120 | 90 |
| Barium, dissolved in water | µg/L | Primary Direct | EPA 200.8 | ICP-MS | 01005 | 10 | 25 | 80-120 | 80-120 | 90 |

Appendix F Data Quality Objectives for Measurement Data

| Parameter | Units | Method Type | Method | Method Description | Storet | MAL | Precision of Laboratory Duplicates (RPD) | Accuracy of Matrix Spikes % Recovery | Accuracy crm | Percent Complete |
|---------------------------------------|-------|--------------------|-------------------------|--------------------|--------|---------------------------|--|--------------------------------------|--------------|------------------|
| Barium in sediment | mg/kg | Primary Direct | EPA 6020/200.8 | ICP-MS | 01008 | 2.5 | 25 | 80-120 | 80-120 | 90 |
| | | ICP-MS | EPA 200.8 | ICP-MS | 01025 | 0.1 | 25 | 80-120 | 80-120 | 90 |
| Cadmium, dissolved in water | µg/L | Alternate Direct | EPA 200.9 | GFAAS | 01025 | 0.05 | 25 | 64-145 | 64-145 | 90 |
| | | Primary Direct | EPA 200.8 | ICP-MS | 01027 | 0.1 | 25 | 84-113 | 84-113 | 90 |
| Cadmium in sediment | mg/kg | Alternate Direct | EPA 200.9 | GFAAS | 01027 | 0.05 | 25 | 64-145 | 64-145 | 90 |
| | | Primary Direct | EPA 200.8 or 6010B/6020 | ICP-MS | 01028 | 0.2 | 25 | 80-120 | 80-120 | 90 |
| Calcium, dissolved in water | mg/L | ICP/AES | EPA 200.7 | ICP-AES | 00915 | 0.05 | 20 | 84-113 | 84-113 | 90 |
| | | Alternate Direct | EPA 215.1 | Flame AAS | 00915 | 0.03 | 20 | 80-120 | 80-120 | 90 |
| Calcium, total recoverable in water | mg/L | ICP/AES | EPA 200.7 | ICP-AES | 00916 | 0.05 | 20 | 84-113 | 84-113 | 90 |
| | | Primary Direct | EPA 200.8 or 6010B/6020 | ICP-MS | 00917 | 12.5 | 25 | 80-120 | 80-120 | 90 |
| Chromium, dissolved in water | µg/L | ICP-MS | EPA 200.8 | ICP-MS | 01030 | 2.0 | 25 | 80-120 | 80-120 | 90 |
| | | Primary Direct | EPA 200.8 | ICP-MS | 01034 | 2.0 | 25 | 80-120 | 80-120 | 90 |
| Chromium (hexavalent), total in water | µg/L | Ion Chromatography | EPA 1636 | IC | 01032 | 5.0 | 20 | 79-122 | 79-122 | 90 |
| | | Primary Direct | EPA 6020/200.8 | ICP-MS | 01029 | 2 | 25 | 80-120 | 80-120 | 90 |
| Copper, dissolved in water | µg/L | ICP-MS | EPA 200.8 | ICP-MS | 01040 | 0.2 | 25 | 51-145 | 51-145 | 90 |
| | | Primary Direct | EPA 200.8 | ICP-MS | 01042 | 0.2 | 25 | 51-145 | 51-145 | 90 |
| Copper in sediment | mg/kg | Primary Direct | EPA 6020/200.8 | ICP-MS | 01043 | 2.5 | 25 | 80-120 | 80-120 | 90 |
| | | Primary Direct | EPA 130.1-2 | Titrametric EDTA | 00900 | 1.0, as CaCO ₃ | 20 | 80-120 | 80-120 | 90 |
| Iron, total recoverable in water | µg/L | ICP-AES | EPA 200.7 | ICP-AES | 01045 | 0.05 | | | | 90 |
| | | ICP/MS | EPA 6020A | ICP/MS | 01170 | 12.5 | | | | 90 |
| Lead, dissolved in water | µg/L | ICP-MS | EPA 200.8 | ICP-MS | 01049 | 0.05 | 25 | 72-143 | 72-143 | 90 |
| | | Primary Direct | EPA 200.8 | ICP-MS | 01051 | 0.05 | 25 | 72-143 | 72-143 | 90 |
| Lead, in sediment | mg/kg | Primary Direct | EPA 200.8 or 6010B/6020 | ICP-MS | 01052 | 2 | 25 | 80-120 | 80-120 | 90 |
| | | Alternate Direct | EPA 200.7 | ICP-AES | 00925 | 0.05 | 20 | 80-120 | 80-120 | 90 |
| Magnesium, dissolved in water | mg/L | ICP/AES | EPA 200.7 | ICP-AES | 00925 | 0.003 | 20 | 80-120 | 80-120 | 90 |
| | | Alternate Direct | EPA 242.1 | Flame AAS | 00925 | | | | | 90 |
| Magnesium, total recoverable in water | mg/L | ICP/AES | EPA 200.7 | ICP-AES | 00927 | 0.05 | 20 | 80-120 | 80-120 | 90 |
| | | ICP/MS | EPA 6020 | ICP/MS | 00924 | 2.5 | 25 | NA | 80-120 | 90 |
| Mercury, dissolved in water | µg/L | Primary Direct | EPA 1631 | P/T CVAF | 71890 | 0.0005 | 25 | 71-125 | 71-125 | 90 |
| | | Primary Direct | EPA 1631 | P/T CVAF | 71900 | 0.0005 | 25 | 71-125 | 71-125 | 90 |
| Mercury, total recoverable in water | µg/L | Primary Direct | EPA 245.5 | P/T CVAF CVAAS | 71921 | 0.05 | 25 | 80-120 | 80-120 | 90 |
| | | Primary Direct | EPA 245.5 | P/T CVAF CVAAS | 71921 | 0.05 | 25 | 80-120 | 80-120 | 90 |

Appendix F Data Quality Objectives for Measurement Data

| | | | | | | | | | | |
|--|-------|------------------|----------------------|---------|-------|--------|----|--------|--------|----|
| Nickel, dissolved in water | µg/L | ICP-MS | EPA 200.8 | ICP-MS | 01065 | 1.0 | 20 | 68-134 | 68-134 | 90 |
| | | Alternate Direct | EPA 200.9 | GFAAS | 01065 | 2.0 | 25 | 65-145 | 65-145 | 90 |
| | | Primary Direct | EPA 200.8 | ICP-MS | 01067 | 1.0 | 20 | 68-134 | 68-134 | 90 |
| | | Alternate Direct | EPA 200.9 | GFAAS | 01067 | 2.0 | 25 | 65-145 | 65-145 | 90 |
| Nickel in sediment | mg/kg | Primary Direct | EPA 6020/200.8 | ICP-MS | 01068 | 2.5 | 20 | 80-120 | 80-120 | 90 |
| | µg/L | Primary Direct | EPA 200.8 | ICP-MS | 01145 | 1 or 2 | 25 | 59-149 | 59-149 | 90 |
| Selenium, dissolved in water | µg/L | Alternate Direct | EPA 200.9 | GFAAS | 01145 | 2 | 25 | 56-131 | 56-131 | 90 |
| | | ICP-MS | EPA 200.8 | ICP-MS | 01147 | 2 | 25 | 59-149 | 59-149 | 90 |
| Selenium, total recoverable in water | µg/L | Alternate Direct | EPA 200.9 | GFAAS | 01147 | 2 | 25 | 56-131 | 56-131 | 90 |
| | | Primary Direct | EPA 6010B/6020/200.8 | ICP-MS | 01148 | 5 | 25 | 80-120 | 80-120 | 90 |
| Silver, dissolved in water | µg/L | ICP-MS | EPA 200.8 | ICP-MS | 01075 | 0.1 | 25 | 74-119 | 74-119 | 90 |
| | µg/L | Primary Direct | EPA 200.8 | ICP-MS | 01077 | 0.1 | 25 | 74-119 | 74-119 | 90 |
| Silver in sediment | mg/kg | Primary Direct | EPA 6020/200.8 | ICP-MS | 01078 | 1 | 25 | 75-125 | 75-125 | 90 |
| | µg/L | ICP-MS | EPA 200.8 | ICP-MS | 01090 | 0.5 | 25 | 46-146 | 46-146 | 90 |
| Zinc, dissolved in water | µg/L | Alternate Direct | EPA 200.7 | ICP-AES | 01090 | 5.0 | 25 | 67-142 | 67-142 | 90 |
| | | Alternate Direct | EPA 200.9 | GFAAS | 01090 | 0.5 | 25 | 67-142 | 67-142 | 90 |
| Zinc, total in water | µg/L | Primary Direct | EPA 200.8 | ICP-MS | 01092 | 0.5 | 25 | 46-146 | 46-146 | 90 |
| | | Alternate Direct | EPA 200.7 | ICP-MS | 01092 | 5.0 | 25 | 80-120 | 80-120 | 90 |
| Zinc, in sediment | mg/kg | Alternate Direct | EPA 200.9 | GFAAS | 01092 | 0.5 | 25 | 67-142 | 67-142 | 90 |
| | | Primary Direct | EPA 6020/200.8 | ICP-MS | 01093 | 2.5 | 25 | 80-120 | 80-120 | 90 |
| Organic and Organometal Compounds | | | | | | | | | | |
| Acenaphthene in water | µg/L | Primary | EPA 8270C | GC/MS | 34205 | 4 | 30 | 49-125 | 49-125 | 90 |
| Acenaphthene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34208 | 133 | 30 | 47-145 | 47-145 | 90 |
| Anthracene in water | µg/L | Primary | EPA 8270C | GC/MS | 34220 | 4 | 30 | 45-165 | 45-165 | 90 |
| Anthracene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34223 | 660 | 30 | 27-133 | 27-133 | 90 |
| Acenaphthylene in water | µg/L | Primary | EPA 8270C | GC/MS | 34200 | 4 | 30 | 47-125 | 47-125 | 90 |
| Acenaphthylene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34203 | 660 | 30 | 33-145 | 33-145 | 90 |
| Acrolein in sediment (Propenal) | µg/kg | Primary | EPA8260B | GC/MS | 34213 | 51 | 40 | 25-175 | 25-175 | 90 |
| Acrylonitrile in water | µg/L | Primary | EPA8260B | GC/MS | 34215 | 50 | 20 | 50-150 | 50-150 | 90 |
| Acrylonitrile in sediment | µg/kg | Primary | EPA8260B | GC/MS | 34218 | 3.71 | 40 | 25-175 | 25-175 | 90 |

Appendix F Data Quality Objectives for Measurement Data

| | | | | | | | | | | |
|---------------------------------------|-------|-----------|-----------|---|-------|------|----|--------|--------|----|
| Alachlor in water | µg/L | Primary | EPA 8081 | GC/ECD | 77825 | 0.10 | 25 | 50-150 | 50-150 | 90 |
| | | Alternate | EPA 525.1 | L/S Extraction + Capillary GC/MS | 77825 | 0.3 | 25 | | | 90 |
| Alachlor in sediment | | Alternate | EPA 645 | GC | | 0.6 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | | 0.06 | 25 | 23-101 | | 90 |
| Aldrin in water | µg/kg | Primary | EPA 8081 | GC/ECD | 75050 | 100 | 30 | 50-150 | 50-150 | 90 |
| | µg/L | Primary | EPA 8081 | GC/ECD | 39330 | 0.05 | 25 | 20-100 | 20-100 | 90 |
| Aldrin in sediment | µg/kg | Primary | EPA 8081 | GC/NPD | 39333 | 50 | 30 | 50-150 | 50-150 | 90 |
| | µg/L | Primary | EPA 619 | GC | 39630 | 0.15 | 25 | 62-191 | 62-191 | 90 |
| Atrazine in water | | Alternate | EPA 525.1 | L/S Extraction + Capillary GC/MS | | 0.42 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | | 1.5 | 25 | 31-132 | | 90 |
| Atrazine in sediment | µg/kg | Primary | EPA 8141 | GC/NPD | 39631 | 50 | 30 | | | 90 |
| | µg/L | Primary | EPA 8260B | GC/MS | 34030 | 1 | 20 | 75-125 | 75-125 | 90 |
| Benzene in water | µg/kg | Primary | EPA 8260B | GC/MS | 34237 | 10 | 40 | 25-165 | 25-165 | 90 |
| | µg/L | Primary | EPA 8260B | GC/MS | 32104 | 1 | 20 | 75-125 | 75-125 | 90 |
| Bromoform in water | µg/kg | Primary | EPA 8260B | GC/MS | 34290 | 10 | 40 | 30-180 | 30-180 | 90 |
| | µg/L | Primary | EPA 8260B | GC/MS | 30202 | 1 | 20 | 62-147 | 62-147 | 90 |
| Bromomethane in water | µg/kg | Primary | EPA 8260B | GC/MS | 88802 | 5 | 30 | 70-130 | 70-130 | 90 |
| | µg/L | Primary | EPA 8270C | GC/MS | 34526 | 4 | 30 | 51-133 | 51-133 | 90 |
| Benz (a) Anthracene in water | µg/kg | Primary | EPA 8270C | GC/MS | 34529 | 660 | 30 | 33-143 | 33-143 | 90 |
| | µg/L | Primary | EPA 8270C | GC/MS | 34247 | 4 | 30 | 41-125 | 41-125 | 90 |
| Benzo (a) Pyrene in water | µg/kg | Primary | EPA 8270C | GC/MS | 34250 | 660 | 30 | 17-163 | 17-163 | 90 |
| | µg/L | Primary | EPA 8270C | GC/MS | 34230 | 4 | 30 | 37-125 | 37-125 | 90 |
| Benzo (b) fluoranthene in water | µg/kg | Primary | EPA 8270C | GC/MS | 34233 | 133 | 30 | 24-159 | 24-159 | 90 |
| | µg/L | Primary | EPA 8270C | GC/MS | 34521 | 4 | 30 | 34-149 | 34-149 | 90 |
| Benzo (ghi) Perylene in water | µg/kg | Primary | EPA 8270C | GC/MS | 34524 | 660 | 30 | 15-219 | 15-219 | 90 |
| | µg/L | Primary | EPA 8270C | GC/MS | 34242 | 4 | 30 | 34-149 | 34-149 | 90 |
| Benzo (k) Fluoranthene in water | µg/kg | Primary | EPA 8270C | GC/MS | 34245 | 660 | 30 | 11-162 | 11-162 | 90 |
| | µg/L | Primary | EPA 8081 | GC/ECD | 39337 | 0.05 | 25 | 35-117 | 35-117 | 90 |
| BHC, alpha in water | µg/kg | Primary | EPA 8081 | GC/ECD | 39076 | 50 | 30 | 38-137 | 38-137 | 90 |
| | µg/L | Primary | EPA 8081 | GC/ECD | 39338 | 0.05 | 25 | 51-121 | 51-121 | 90 |
| BHC, beta in water | µg/kg | Primary | EPA 8081 | GC/ECD | 34257 | 50 | 30 | 51-133 | 51-133 | 90 |
| | µg/L | Primary | EPA 8081 | GC/ECD | 34259 | 0.05 | 25 | 32-121 | 32-121 | 90 |
| BHC, delta in water | µg/kg | Primary | EPA 8081 | GC/ECD | 34262 | 50 | 30 | 43-131 | 43-131 | 90 |
| | µg/L | Primary | EPA 8081 | GC/ECD | 39782 | 0.05 | 25 | 41-114 | 41-114 | 90 |
| BHC, gamma (Lindane) in water | µg/kg | Primary | EPA 8081 | GC/ECD | 39783 | 50 | 30 | 47-132 | 47-132 | 90 |
| | µg/L | Primary | EPA 8270C | GC/MS | 34278 | 4 | 30 | 49-125 | 49-125 | 90 |
| Bis (2-Chloroethoxy) Methane in water | µg/kg | Primary | EPA 8270C | GC/MS | 34281 | 660 | 30 | 33-184 | 33-184 | 90 |
| | µg/L | Primary | EPA 8270C | GC/MS | 34273 | 4 | 30 | 44-125 | 44-125 | 90 |
| Bis (2-Chloroethyl) Ether in water | µg/kg | Primary | EPA 8270C | GC/MS | 34276 | 133 | 30 | 12-158 | 12-158 | 90 |
| | µg/L | Primary | EPA 8270C | GC/MS | 34283 | 4 | 30 | 36-166 | 36-166 | 90 |

Appendix F Data Quality Objectives for Measurement Data

| | | | | | | | | | | |
|---|-------|-----------|-----------|----------------------------------|-------|------|----|--------|--------|----|
| Bis (2-Chloroisopropyl) Ether in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34286 | 133 | 30 | 36-166 | 36-166 | 90 |
| Bis (2-Ethylhexyl) Phthalate in water | µg/L | Primary | EPA 8270C | GC/MS | 39100 | 4 | 30 | 33-129 | 33-129 | 90 |
| Bis (2-Ethylhexyl) Phthalate in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 39102 | 660 | 30 | 8-158 | 8-158 | 90 |
| 4-Bromophenyl Phenyl Ether in water | µg/L | Primary | EPA 8270C | GC/MS | 34636 | 4 | 30 | 53-127 | 53-127 | 90 |
| 4-Bromophenyl Phenyl Ether in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34639 | 660 | 30 | 53-130 | 53-130 | 90 |
| N-Butylbenzyl Phthalate in water | µg/L | Primary | EPA 8270C | GC/MS | 34292 | 10 | 30 | 26-125 | 26-125 | 90 |
| N-Butylbenzyl Phthalate in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34295 | 660 | 30 | 15-152 | 15-152 | 90 |
| Carbaryl (Sevin) in water | µg/L | Primary | EPA 8321 | HPLC/MS | 39750 | 1 | 25 | 40-131 | 40-131 | 90 |
| Carbaryl (Sevin) in sediment | µg/kg | Primary | EPA 8321 | HPLC/MS | 81818 | 20 | 25 | 34-129 | 34-129 | 90 |
| Carbon disulfide in water | µg/L | Primary | EPA 8260B | GC/MS | 77041 | 25 | 20 | 50-150 | 50-150 | 90 |
| | | Alternate | EPA 1624 | Isotope Dilution GC/MS | 77041 | 25 | | | | 90 |
| Carbon disulfide in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 78544 | 50 | 30 | 50-150 | 50-150 | 90 |
| | | Alternate | EPA 1624 | Isotope Dilution GC/MS | 78544 | | 25 | | | 90 |
| Carbon Tetrachloride in water | µg/L | Primary | EPA 8260B | GC/MS | 32102 | 1 | 20 | 62-125 | 62-125 | 90 |
| Carbon Tetrachloride in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34299 | 10 | 40 | 60-150 | 60-150 | 90 |
| Chlorobenzene in water | µg/L | Primary | EPA 8260B | GC/MS | 34301 | 1 | 20 | 75-125 | 75-125 | 90 |
| Chlorobenzene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34304 | 10 | 40 | 20-175 | 20-175 | 90 |
| Chlorodibromomethane in water | µg/L | Primary | EPA 8260B | GC/MS | 32105 | 1 | 20 | 73-125 | 73-125 | 90 |
| Chlorodibromomethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34309 | 5 | 40 | 40-160 | 40-160 | 90 |
| Chloroethane in water | µg/L | Primary | EPA 8260B | GC/MS | 34311 | 1 | 50 | 53-145 | 53-145 | 90 |
| Chloroethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34314 | 5 | 40 | 15-255 | 15-255 | 90 |
| 2-Chloroethylvinyl ether in water | µg/L | Primary | EPA 8260B | GC/MS | 34576 | 50 | 20 | 50-150 | 50-150 | 90 |
| 2-Chloroethylvinyl ether in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34579 | 60 | 40 | 15-300 | 15-300 | 90 |
| Chloroform in water | µg/L | Primary | EPA 8260B | GC/MS | 32106 | 1 | 20 | 74-125 | 74-125 | 90 |
| Chloroform in sediment | µg/L | Primary | EPA 8260B | GC/MS | 34318 | 10 | 40 | 40-150 | 40-150 | 90 |
| Chlordane in water | µg/L | Primary | EPA 8081 | GC/ECD | 39350 | 0.05 | 25 | 45-122 | 45-122 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39350 | 1-2 | 25 | 69-133 | 69-133 | 90 |
| | | Alternate | EPA 525.1 | L/S Extraction + Capillary GC/MS | 39350 | 1-2 | 25 | | | 90 |
| Chlordane in sediment | µg/kg | Primary | EPA 8081 | GC/ECD | 39351 | 50 | 30 | 56-142 | 56-142 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | | | 25 | 69-133 | 69-133 | 90 |
| Chloromethane in water | µg/L | Primary | EPA 8260B | GC/MS | 30201 | 1 | 20 | 60-140 | 60-140 | 90 |
| Chloromethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 88835 | 10 | 30 | 70-130 | 70-130 | 90 |
| 2-Chloronaphthalene in water | µg/L | Primary | EPA 8270C | GC/MS | 34581 | 4 | 30 | 60-125 | 60-125 | 90 |
| 2-Chloronaphthalene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34584 | 660 | 30 | 60-130 | 60-130 | 90 |
| 2-Chlorophenol in water | µg/L | Primary | EPA 8270C | GC/MS | 34586 | 4 | 30 | 41-125 | 41-125 | 90 |
| 2-Chlorophenol in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34589 | 133 | 30 | 31-135 | 31-135 | 90 |
| 4-Chlorophenyl Phenyl Ether in water | µg/L | Primary | EPA 8270C | GC/MS | 34641 | 4 | 30 | 51-132 | 51-132 | 90 |
| 4-Chlorophenyl Phenyl Ether in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34644 | 133 | 30 | 25-158 | 25-158 | 90 |
| Chloropyrifos (Dursban) in water | µg/L | Primary | EPA 8141 | GC/NPD | 81403 | 0.5 | 25 | 45-118 | 45-118 | 90 |
| Chloropyrifos (Dursban) in sediment | µg/kg | Primary | EPA 8141 | GC/NPD | 81404 | 50 | 30 | 40-129 | 40-129 | 90 |
| Chrysene in water | µg/L | Primary | EPA 8270C | GC/MS | 34320 | 4 | 30 | 55-133 | 55-133 | 90 |
| Chrysene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34323 | 133 | 30 | 17-168 | 17-168 | 90 |
| Cyanazine in water | µg/L | Primary | EPA 619 | GC/NPD | 81757 | 0.5 | 25 | 30-232 | 30-232 | 90 |
| Cyanazine in sediment | µg/kg | Primary | EPA 619-m | GC/NPD | 03999 | 50 | 30 | | | 90 |

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|--------------------------------------|-------|-----------|-----------|---------|-------|-------|----|--------|--------|----|
| 2,4-D in water | µg/L | Primary | EPA 8151 | GC/ECD | 39730 | 0.5 | 25 | 72-146 | 72-146 | 90 |
| 2,4-D in sediment | µg/kg | Primary | EPA 8151 | GC/ECD | 39731 | 200 | 30 | 89-175 | 89-175 | 90 |
| Demeton in water | µg/L | Primary | EPA 8141 | GC/NPD | 39560 | 1 | 25 | 14-107 | 14-107 | 90 |
| Demeton in sediment | µg/kg | Primary | EPA 8141 | GC/NPD | 82400 | 100 | 30 | 5-108 | 5-108 | 90 |
| Diazinon in water | µg/L | Primary | EPA 8141 | GC/NPD | 39570 | 0.1 | 25 | 34-126 | 34-126 | 90 |
| Diazinon in sediment | µg/kg | Primary | EPA 8141 | GC/NPD | 39571 | 50 | 30 | 39-124 | 39-124 | 90 |
| 1,2-Dibromoethane in water | µg/L | Primary | EPA 8260B | GC/MS | 77651 | 1 | 20 | 75-125 | 75-125 | 90 |
| 1,2-Dibromoethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 88805 | 10 | 30 | 70-130 | 70-130 | 90 |
| Dicofof (Kelthane) in water | µg/L | Primary | EPA 8081 | GC/ECD | 39780 | 0.10 | 25 | | | 90 |
| Dicofof (Kelthane) in sediment | µg/kg | Primary | EPA 8081 | GC/ECD | 79799 | 100 | 30 | | | 90 |
| Dieldrin in water | µg/L | Primary | EPA 8081 | GC/ECD | 39380 | 0.02 | 25 | 52-120 | 52-120 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39380 | 0.02 | 25 | 48-158 | 48-158 | 90 |
| Dieldrin in sediment | µg/kg | Primary | EPA 8081 | GC/ECD | 39383 | 50 | 30 | 56-125 | 56-125 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 38383 | | 25 | 48-158 | 48-158 | 90 |
| BromoDichloromethane in water | µg/L | Primary | EPA 8260B | GC/MS | 32101 | 1 | 20 | 75-125 | 75-125 | 90 |
| BromoDichloromethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34330 | 10 | 40 | 40-160 | 40-160 | 90 |
| 1,1-Dichloroethane in water | µg/L | Primary | EPA 8260B | GC/MS | 34496 | 1 | 20 | 72-125 | 72-125 | 90 |
| 1,1-Dichloroethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34499 | 5 | 40 | 45-165 | 45-165 | 90 |
| 1,2-Dichloroethane in water | µg/L | Primary | EPA 8260B | GC/MS | 34531 | 1 | 20 | 68-127 | 68-127 | 90 |
| 1,2-Dichloroethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34534 | 5 | 40 | 40-165 | 40-165 | 90 |
| 1,1-Dichloroethylene in water | µg/L | Primary | EPA 8260B | GC/MS | 34501 | 1 | 20 | 75-125 | 75-125 | 90 |
| 1,1-Dichloroethylene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34504 | 5 | 40 | 15-260 | 15-260 | 90 |
| 1,2-Dichloropropane in water | µg/L | Primary | EPA 8260B | GC/MS | 34541 | 1 | 20 | 70-125 | 70-125 | 90 |
| 1,2-Dichloropropane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34544 | 5 | 40 | 15-255 | 15-255 | 90 |
| cis 1,3-Dichloropropene in water | µg/L | Primary | EPA 8260B | GC/MS | 34704 | 1 | 20 | 74-125 | 74-125 | 90 |
| cis 1,3-Dichloropropene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34702 | 10 | 30 | 70-130 | 70-130 | 90 |
| 1,3-Dichloropropylene in water | µg/L | Primary | EPA 8260B | GC/MS | 34565 | 10 | 40 | 15-280 | 15-280 | 90 |
| 1,3-Dichloropropylene in sediment | µg/kg | Primary | EPA 8321 | HPLC/MS | 39650 | 1 | 25 | 57-133 | 57-133 | 90 |
| Diuron (Karmex) in water | µg/L | Primary | EPA 8321 | HPLC/MS | 73030 | 20 | 25 | 25-133 | 25-133 | 90 |
| Diuron (Karmex) in sediment | µg/kg | Primary | EPA 8081 | GC/ECD | 39373 | 50 | 30 | 36-129 | 36-129 | 90 |
| DDT in sediment | µg/kg | Alternate | EPA 1656 | GC/ECD | 39373 | 12 | 25 | 79-119 | 79-119 | 90 |
| DDT in water | µg/L | Primary | EPA 8081 | GC/ECD | 39370 | 0.05 | 25 | 27-142 | 27-142 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39370 | 0.036 | 25 | 79-119 | 79-119 | 90 |
| DDE in sediment | µg/kg | Primary | EPA 8081 | GC/ECD | 39368 | 50 | 30 | 58-127 | 58-127 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39368 | 4 | 25 | 54-126 | 54-126 | 90 |
| DDE in water | µg/L | Primary | EPA 8081 | GC/ECD | 39365 | 0.05 | 25 | 29-120 | 29-120 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39365 | 0.030 | 25 | 54-126 | 54-126 | 90 |
| DDD in sediment | µg/kg | Primary | EPA 8081 | GC/ECD | 39363 | 50 | 30 | 51-129 | 51-129 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39363 | 11 | 25 | 57-129 | 57-129 | 90 |
| DDD in water | µg/L | Primary | EPA 8081 | GC/ECD | 39360 | 0.05 | 25 | 44-119 | 44-119 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39360 | 0.015 | 25 | 57-129 | 57-129 | 90 |
| Dibenzo (a,h) Anthracene in water | µg/L | Primary | EPA 8270C | GC/MS | 34556 | 4 | 30 | 50-125 | 50-125 | 90 |
| Dibenzo (a,h) Anthracene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34559 | 660 | 30 | 15-227 | 15-227 | 90 |
| 1,2-Dichlorobenzene in water | µg/L | Primary | EPA 8260B | GC/MS | 34536 | 4 | 30 | 42-155 | 42-155 | 90 |
| 1,2-Dichlorobenzene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34539 | 660 | 30 | 32-130 | 32-130 | 90 |
| 1,3-Dichlorobenzene in water | µg/L | Primary | EPA 8260B | GC/MS | 34566 | 4 | 30 | 36-125 | 36-125 | 90 |
| 1,3-Dichlorobenzene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34569 | 660 | 30 | 15-172 | 15-172 | 90 |
| 1,4-Dichlorobenzene in water | µg/L | Primary | EPA 8260B | GC/MS | 34571 | 4 | 30 | 30-125 | 30-125 | 90 |
| 1,4-Dichlorobenzene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34574 | 660 | 30 | 20-130 | 20-130 | 90 |
| 3,3-Dichlorobenzidine in water | µg/L | Primary | EPA 8270C | GC/MS | 34631 | 4 | 30 | 29-175 | 29-175 | 90 |

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|---------------------------------------|---------|-----------|--------|-------|------|----|--------|--------|----|
| 3,3-Dichlorobenzidine in sediment | Primary | EPA 8270C | GC/MS | 34634 | 133 | 30 | 15-262 | 15-262 | 90 |
| trans-1,2-Dichloroethene in water | Primary | EPA 8260B | GC/MS | 34546 | 1 | 20 | 75-125 | 75-125 | 90 |
| trans-1,2-Dichloroethene in sediment | Primary | EPA 8260B | GC/MS | 34549 | 10 | 30 | 75-125 | 75-125 | 90 |
| 2,4-Dichlorophenol in water | Primary | EPA 8270C | GC/MS | 34601 | 4 | 30 | 46-125 | 46-125 | 90 |
| 2,4-Dichlorophenol in sediment | Primary | EPA 8270C | GC/MS | 34604 | 133 | 30 | 36-135 | 36-135 | 90 |
| trans-1,3-Dichloropropene in water | Primary | EPA 8260B | GC/MS | 34699 | 1 | 20 | 66-125 | 66-125 | 90 |
| trans-1,3-Dichloropropene in sediment | Primary | EPA 8260B | GC/MS | 34697 | 10 | 30 | 70-130 | 70-130 | 90 |
| Diethyl Phthalate in water | Primary | EPA 8270C | GC/MS | 34336 | 10 | 30 | 37-125 | 37-125 | 90 |
| Diethyl Phthalate in sediment | Primary | EPA 8270C | GC/MS | 34339 | 660 | 30 | 15-130 | 15-130 | 90 |
| 2,4-Dimethylphenol in water | Primary | EPA 8270C | GC/MS | 34606 | 4 | 30 | 10-139 | 10-139 | 90 |
| 2,4-Dimethylphenol in sediment | Primary | EPA 8270C | GC/MS | 34609 | 133 | 30 | 30-149 | 30-149 | 90 |
| Dimethyl Phthalate in water | Primary | EPA 8270C | GC/MS | 34341 | 4 | 30 | 25-175 | 25-175 | 90 |
| Dimethyl Phthalate in sediment | Primary | EPA 8270C | GC/MS | 34344 | 660 | 30 | 15-130 | 15-130 | 90 |
| Di-n-Butyl Phthalate in water | Primary | EPA 8270C | GC/MS | 39110 | 10 | 30 | 34-136 | 34-136 | 90 |
| Di-n-Butyl Phthalate in sediment | Primary | EPA 8270C | GC/MS | 39112 | 330 | 30 | 1-130 | 1-130 | 90 |
| 4,6-Dinitro-ortho-cresol in water | Primary | EPA 8270C | GC/MS | 34657 | 10 | 30 | 26-134 | 26-134 | 90 |
| 4,6-Dinitro-ortho-cresol in sediment | Primary | EPA 8270C | GC/MS | 34660 | 330 | 30 | 25-144 | 25-144 | 90 |
| 2,4-Dinitrophenol in water | Primary | EPA 8270C | GC/MS | 34616 | 20 | 30 | 30-151 | 30-151 | 90 |
| 2,4-Dinitrophenol in sediment | Primary | EPA 8270C | GC/MS | 34619 | 660 | 30 | 25-161 | 25-161 | 90 |
| 2,4-Dinitrotoluene in water | Primary | EPA 8270C | GC/MS | 34611 | 4 | 30 | 39-139 | 39-139 | 90 |
| 2,4-Dinitrotoluene in sediment | Primary | EPA 8270C | GC/MS | 34614 | 133 | 30 | 39-139 | 39-139 | 90 |
| 2,6-Dinitrotoluene in water | Primary | EPA 8270C | GC/MS | 34626 | 4 | 30 | 51-125 | 51-125 | 90 |
| 2,6-Dinitrotoluene in sediment | Primary | EPA 8270C | GC/MS | 34629 | 133 | 30 | 50-158 | 50-158 | 90 |
| Di-n-Octyl Phthalate in water | Primary | EPA 8270C | GC/MS | 34596 | 10 | 30 | 38-127 | 38-127 | 90 |
| Di-n-Octyl Phthalate in sediment | Primary | EPA 8270C | GC/MS | 34599 | 660 | 30 | 4-146 | 4-146 | 90 |
| Endosulfan in water | Primary | EPA 8081 | GC/ECD | 39388 | 0.05 | 25 | 55-123 | 55-123 | 90 |
| Endosulfan in sediment | Primary | EPA 8081 | GC/ECD | 39389 | 50 | 30 | 56-142 | 56-142 | 90 |
| Endosulfan Sulfate in water | Primary | EPA 8081 | GC/ECD | 34351 | 0.05 | 25 | 51-126 | 51-126 | 90 |
| Endosulfan Sulfate in sediment | Primary | EPA 8081 | GC/ECD | 34354 | 50 | 30 | 25-153 | 25-153 | 90 |
| Endrin in water | Primary | EPA 8081 | GC/ECD | 39390 | 0.05 | 25 | 40-138 | 40-138 | 90 |
| Endrin in sediment | Primary | EPA 8081 | GC/ECD | 39393 | 50 | 30 | 44-129 | 44-129 | 90 |
| Ethylbenzene in water | Primary | EPA 8260B | GC/MS | 34371 | 1 | 20 | 75-125 | 75-125 | 90 |
| Ethylbenzene in sediment | Primary | EPA 8260B | GC/MS | 34374 | 5 | 40 | 25-175 | 25-175 | 90 |
| Fluorene in water | Primary | EPA 8270C | GC/MS | 34381 | 4 | 30 | 48-139 | 48-139 | 90 |
| Fluorene in sediment | Primary | EPA 8270C | GC/MS | 34384 | 660 | 30 | 59-130 | 59-130 | 90 |
| Fluoranthene in water | Primary | EPA 8270C | GC/MS | 34376 | 4 | 30 | 26-137 | 26-137 | 90 |
| Fluoranthene in sediment | Primary | EPA 8270C | GC/MS | 34379 | 133 | 30 | 26-137 | 26-137 | 90 |
| Guthion (Azinphos methyl) in water | Primary | EPA 8141 | GC/NPD | 39580 | 5.0 | 25 | 13-155 | 13-155 | 90 |
| Guthion (Azinphos methyl) in sediment | Primary | EPA 8141 | GC/NPD | 39581 | 500 | 30 | 36-153 | 36-153 | 90 |
| Heptachlor in water | Primary | EPA 8081 | GC/ECD | 39410 | 0.05 | 25 | 12-122 | 12-122 | 90 |
| Heptachlor in sediment | Primary | EPA 8081 | GC/ECD | 39413 | 50 | 30 | 37-149 | 37-149 | 90 |

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|---------------------------------------|------------------------|-----------------------|---|----------------|-------------|----------|------------------|------------------|----------|
| Heptachlor epoxide in water | Primary | EPA 8081 | GC/ECD | 39420 | 0.05 | 25 | 52-121 | 52-121 | 90 |
| | Alternate/Confirmatory | EPA 1656 EPA 525.1 | GC/ECD L/S Extraction + Capillary GC/MS | 39420 39420 | 0.04 0.7 | 25 25 | 49-131 49-131 | 48-158 48-158 | 90 90 |
| | Primary | EPA 8081 | GC/ECD | 39423 | 50 | 30 | 55-140 | 55-140 | 90 |
| Hexachlorobenzene in water | Alternate | EPA 1656 | GC/ECD | 39423 | 1.0 | 25 | 49-131 | 49-131 | 90 |
| | Primary | EPA 8270C | GC/MS | 39700 | 4 | 30 | 46-133 | 46-133 | 90 |
| | Primary | EPA 8270C | GC/MS | 39701 | 133 | 30 | 15-152 | 15-152 | 90 |
| Hexachlorobutadiene in water | Primary | EPA 8260B | GC/MS | 34391 | 1 | 20 | 59-128 | 59-128 | 90 |
| | Primary | EPA 8260B | GC/MS | 39705 | 5 | 30 | 24-130 | 24-130 | 90 |
| | Primary | EPA 8270C | GC/MS | 34386 | 10 | 30 | 20-125 | 20-125 | 90 |
| Hexachlorocyclopentadiene in sediment | Primary | EPA 8270C | GC/MS | 34389 | 330 | 30 | 31-135 | 31-135 | 90 |
| | Primary | EPA 8270C | GC/MS | 34396 | 4 | 30 | 25-153 | 25-153 | 90 |
| | Primary | EPA 8270C | GC/MS | 34399 | 133 | 30 | 40-130 | 40-130 | 90 |
| Hexachloroethane in water | Primary | EPA 8270C | GC/MS | 34403 | 4 | 30 | 27-160 | 27-160 | 90 |
| | Primary | EPA 8270C | GC/MS | 34406 | 133 | 30 | 25-170 | 25-170 | 90 |
| | Primary | EPA 8270C | GC/MS | 34408 | 4 | 30 | 26-175 | 26-175 | 90 |
| Indeno[1,2,3-cd]pyrene in sediment | Primary | EPA 8270C | GC/MS | 34411 | 133 | 30 | 25-175 | 25-175 | 90 |
| | Primary | EPA 8270C | GC/MS | 39530 | 0.5 | 25 | 40-132 | 40-132 | 90 |
| | Primary | EPA 8141 | GC/NPD | 39531 | 50 | 30 | 45-127 | 45-127 | 90 |
| Isophorone in water | Primary | EPA 8141 | GC/NPD | 39480 | 0.05 | 25 | 39-160 | 39-160 | 90 |
| | Primary | EPA 8081 | GC/ECD | 39481 | 50 | 30 | 37-144 | 37-144 | 90 |
| | Primary | EPA 8260B | GC/MS | 34416 | 5 | 40 | 15-305 | 15-305 | 90 |
| Malathion in sediment | Primary | EPA 8260B | GC/MS | 34421 | 5 | 40 | 15-320 | 15-320 | 90 |
| | Primary | EPA 8260B | GC/MS | 34423 | 1 | 20 | 75-125 | 75-125 | 90 |
| | Primary | EPA 8260B | GC/MS | 34426 | 5 | 40 | 15-250 | 15-250 | 90 |
| Methoxychlor in water | Primary | EPA 8270C | GC/MS | 34452 | 4 | 30 | 44-125 | 44-125 | 90 |
| | Primary | EPA 8270C | GC/MS | 34455 | 133 | 30 | 34-135 | 34-135 | 90 |
| | Primary | EPA 8270C | GC/MS | 45502 | 660 | 30 | 21-133 | 21-133 | 90 |
| Methylene Chloride in sediment | Primary | EPA 8270C | GC/MS | 77152 | 4 | 30 | 25-125 | 25-125 | 90 |
| | Primary | EPA 8270C | GC/MS | 77146 | 4 | 30 | 25-125 | 25-125 | 90 |
| | Primary | EPA 8270C | GC/MS | 78872 | 134 | 30 | 25-135 | 25-135 | 90 |
| 3-Methyl-4-Chlorophenol in water | Primary | EPA 8270C | GC/MS | 78803 | 134 | 30 | 25-135 | 25-135 | 90 |
| | Primary | EPA 8260B | GC/MS | 46491 | 5 | 20 | 65-135 | 65-135 | 90 |
| | Primary | EPA 8260B | GC/MS | 50928 | 10 | 30 | 70-130 | 70-130 | 90 |
| Methyl Chloride in sediment | Primary | EPA 8141 | GC/NPD | 82612 | 0.5 | 25 | | | 90 |
| | Primary | EPA 8141 | GC/NPD | 38923 | 50 | 30 | | | 90 |
| | Primary | EPA 8081 | GC/ECD | 39755 | 0.1 | 25 | | | 90 |
| Methyl tert-butyl ether in water | Primary | EPA 8081 | GC/ECD | 79800 | 100 | 30 | | | 90 |
| | Primary | EPA 8270C | GC/MS | 34696 | 4 | 30 | 50-125 | 50-125 | 90 |
| | Primary | EPA 8270C | GC/MS | 34445 | 660 | 30 | 21-133 | 21-133 | 90 |
| Metolachlor in water | Primary | EPA 8270C | GC/MS | 34447 | 4 | 30 | 46-133 | 46-133 | 90 |
| | Primary | EPA 8270C | GC/MS | 34450 | 133 | 30 | 36-143 | 36-143 | 90 |
| | Primary | EPA 8270C | GC/MS | 34433 | 4 | 30 | 27-125 | 27-125 | 90 |
| Mirex in sediment | Primary | EPA 8270C | GC/MS | 34436 | 133 | 30 | 25-135 | 25-135 | 90 |
| | Primary | EPA 8270C | GC/MS | 34696 | 4 | 30 | 50-125 | 50-125 | 90 |
| | Primary | EPA 8270C | GC/MS | 34445 | 660 | 30 | 21-133 | 21-133 | 90 |
| Naphthalene in water | Primary | EPA 8270C | GC/MS | 34447 | 4 | 30 | 46-133 | 46-133 | 90 |
| | Primary | EPA 8270C | GC/MS | 34450 | 133 | 30 | 36-143 | 36-143 | 90 |
| | Primary | EPA 8270C | GC/MS | 34433 | 4 | 30 | 27-125 | 27-125 | 90 |
| Naphthalene in sediment | Primary | EPA 8270C | GC/MS | 34436 | 133 | 30 | 25-135 | 25-135 | 90 |
| | Primary | EPA 8270C | GC/MS | 34696 | 4 | 30 | 50-125 | 50-125 | 90 |
| | Primary | EPA 8270C | GC/MS | 34445 | 660 | 30 | 21-133 | 21-133 | 90 |
| Nitrobenzene in water | Primary | EPA 8270C | GC/MS | 34447 | 4 | 30 | 46-133 | 46-133 | 90 |
| | Primary | EPA 8270C | GC/MS | 34450 | 133 | 30 | 36-143 | 36-143 | 90 |
| | Primary | EPA 8270C | GC/MS | 34433 | 4 | 30 | 27-125 | 27-125 | 90 |
| Nitrobenzene in sediment | Primary | EPA 8270C | GC/MS | 34436 | 133 | 30 | 25-135 | 25-135 | 90 |
| | Primary | EPA 8270C | GC/MS | 34696 | 4 | 30 | 50-125 | 50-125 | 90 |
| | Primary | EPA 8270C | GC/MS | 34445 | 660 | 30 | 21-133 | 21-133 | 90 |
| N-Nitrosodiphenylamine in water | Primary | EPA 8270C | GC/MS | 34447 | 4 | 30 | 46-133 | 46-133 | 90 |
| | Primary | EPA 8270C | GC/MS | 34450 | 133 | 30 | 36-143 | 36-143 | 90 |
| | Primary | EPA 8270C | GC/MS | 34433 | 4 | 30 | 27-125 | 27-125 | 90 |
| N-Nitrosodiphenylamine in sediment | Primary | EPA 8270C | GC/MS | 34436 | 133 | 30 | 25-135 | 25-135 | 90 |
| | Primary | EPA 8270C | GC/MS | 34696 | 4 | 30 | 50-125 | 50-125 | 90 |
| | Primary | EPA 8270C | GC/MS | 34445 | 660 | 30 | 21-133 | 21-133 | 90 |

Appendix F Data Quality Objectives for Measurement Data

| | | | | | | | | | | |
|---------------------------------------|-------|-----------|-----------|--------|-------|-------|----|--------|--------|----|
| N-Nitrosodi-n-propylamine in water | µg/L | Primary | EPA 8270C | GC/MS | 34428 | 4 | 30 | 37-125 | 37-125 | 90 |
| N-Nitrosodi-n-propylamine in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34431 | 133 | 30 | 27-135 | 27-135 | 90 |
| 2-Nitrophenol in water | µg/L | Primary | EPA 8270C | GC/MS | 34591 | 4 | 30 | 44-125 | 44-125 | 90 |
| 2-Nitrophenol in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34594 | 133 | 30 | 34-135 | 34-135 | 90 |
| 4-Nitrophenol in water | µg/L | Primary | EPA 8270C | GC/MS | 34646 | 4 | 30 | 15-131 | 15-131 | 90 |
| 4-Nitrophenol in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34649 | 133 | 30 | 25-141 | 25-141 | 90 |
| Parathion in water | µg/L | Primary | EPA 8141 | GC/NPD | 39540 | 0.5 | 25 | 39-136 | 39-136 | 90 |
| Parathion in sediment | µg/kg | Primary | EPA 8141 | GC/NPD | 39541 | 50 | 30 | 33-139 | 33-139 | 90 |
| Pentachlorophenol in water | µg/L | Primary | EPA 8270C | GC/MS | 39032 | 4 | 30 | 28-136 | 28-136 | 90 |
| Pentachlorophenol in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 39061 | 133 | 30 | 38-146 | 38-146 | 90 |
| Pyrene in water | µg/L | Primary | EPA 8270C | GC/MS | 34469 | 4 | 30 | 47-136 | 47-136 | 90 |
| Pyrene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34472 | 660 | 30 | 52-130 | 52-130 | 90 |
| Phenanthrene in water | µg/L | Primary | EPA 8270C | GC/MS | 34461 | 4 | 30 | 54-125 | 54-125 | 90 |
| Phenanthrene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34464 | 13310 | 30 | 54-130 | 54-130 | 90 |
| Phenol in water | µg/L | Primary | EPA 8270C | GC/MS | 34694 | 4 | 30 | 15-125 | 15-125 | 90 |
| Phenol in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34695 | 133 | 30 | 25-135 | 25-135 | 90 |
| PCBs in water | µg/L | Primary | EPA 8082 | GC/ECD | 39516 | 0.5 | 25 | 30-117 | 30-117 | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39516 | 0.35 | 25 | 75-119 | 75-119 | 90 |
| PCB-1242 in water | µg/L | Primary | EPA 8082 | GC/ECD | 39496 | 0.35 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39496 | 0.35 | 25 | 75-119 | 75-119 | 90 |
| PCB-1254 in water | µg/L | Primary | EPA 8082 | GC/ECD | 39504 | 0.35 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39504 | 0.35 | 25 | 75-119 | 75-119 | 90 |
| PCB-1221 in water | µg/L | Primary | EPA 8082 | GC/ECD | 39488 | 0.35 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39488 | 0.35 | 25 | 75-119 | 75-119 | 90 |
| PCB-1232 in water | µg/L | Primary | EPA 8082 | GC/ECD | 39492 | 0.35 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39492 | 0.35 | 25 | 75-119 | 75-119 | 90 |
| PCB-1248 in water | µg/L | Primary | EPA 8082 | GC/ECD | 39500 | 0.35 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39500 | 0.35 | 25 | 75-119 | 75-119 | 90 |
| PCB-1260 in water | µg/L | Primary | EPA 8082 | GC/ECD | 39508 | 0.35 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39508 | 0.35 | 25 | 75-119 | 75-119 | 90 |
| PCB-1016 in water | µg/L | Primary | EPA 8082 | GC/ECD | 34671 | 0.35 | 25 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 34671 | 0.35 | 25 | 75-119 | 75-119 | 90 |
| PCBs in sediment total | µg/kg | Primary | EPA 8082 | GC/ECD | 39519 | 200 | 30 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39519 | 1.0 | 25 | 75-119 | 75-119 | 90 |
| PCB-1242 In Sediment | µg/kg | Primary | EPA 8082 | GC/ECD | 39499 | 200 | 30 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39499 | 1.0 | 25 | 75-119 | 75-119 | 90 |
| PCB-1254 In Sediment | µg/kg | Primary | EPA 8082 | GC/ECD | 39507 | 200 | 30 | | | 90 |
| | | Alternate | EPA 1656 | GC/ECD | 39507 | 1.0 | 25 | 75-119 | 75-119 | 90 |

Appendix F Data Quality Objectives for Measurement Data

| | | | | | | | | | | |
|---------------------------------------|-------|------------------------|-----------|---|-------|------|----|--------|--------|----|
| PCB-1221 In Sediment | µg/kg | Primary | EPA 8082 | GC/ECD | 39491 | 200 | 30 | | | 90 |
| PCB-1221 In Sediment | µg/kg | Alternate | EPA 1656 | GC/ECD | 39491 | 1.0 | 25 | 75-119 | 75-119 | 90 |
| PCB-1232 In Sediment | µg/kg | Primary | EPA 8082 | GC/ECD | 39495 | 200 | 30 | | | 90 |
| | µg/kg | Alternate | EPA 1656 | GC/ECD | 39495 | 1.0 | 25 | 75-119 | 75-119 | 90 |
| PCB-1248 In Sediment | µg/kg | Primary | EPA 8082 | GC/ECD | 39503 | 200 | 30 | | | 90 |
| | µg/kg | Alternate | EPA 1656 | GC/ECD | 39503 | 1.0 | 25 | 75-119 | 75-119 | 90 |
| PCB-1260 In Sediment | µg/kg | Primary | EPA 8082 | GC/ECD | 39511 | 200 | 30 | 61-118 | 61-118 | 90 |
| | µg/kg | Alternate | EPA 1656 | GC/ECD | 39511 | 1.0 | 25 | 75-119 | 75-119 | 90 |
| PCB-1016 In Sediment | µg/kg | Primary | EPA 8082 | GC/ECD | 39514 | 200 | 30 | 56-113 | 56-113 | 90 |
| | µg/kg | Alternate | EPA 1656 | GC/ECD | 39514 | 1.0 | 25 | 75-119 | 75-119 | 90 |
| Simazine in water | µg/L | Primary | EPA 8141 | GC/NPD | 39055 | 0.5 | 25 | 35-135 | 35-135 | 90 |
| Simazine in sediments | µg/L | Primary | EPA 8141 | GC/NPD | 39046 | 50 | 30 | 35-135 | 35-135 | 90 |
| 2,4,5-T in water | µg/L | Primary | EPA 8151 | GC/ECD | 39740 | 0.10 | 25 | 45-134 | 45-134 | 90 |
| 2,4,5-T in sediment | µg/kg | Primary | EPA 8151 | GC/ECD | 39741 | 40 | 30 | 48-153 | 48-153 | 90 |
| 2,4,5-TP (Silvex) in water | µg/L | Primary | EPA 8151 | GC/ECD | 39760 | 0.1 | 25 | 46-125 | 46-125 | 90 |
| 2,4,5-TP (Silvex) in sediment | µg/kg | Primary | EPA 8151 | GC/ECD | 39761 | 40 | 30 | 54-145 | 54-145 | 90 |
| 1,1,2,2-Tetrachloroethane in water | µg/L | Primary | EPA 8260B | GC/MS | 34516 | 1 | 20 | 74-125 | 74-125 | 90 |
| 1,1,2,2-Tetrachloroethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34519 | 5 | 40 | 35-170 | 35-170 | 90 |
| Tetrachloroethene in water | µg/L | Primary | EPA 8260B | GC/MS | 34475 | 1 | 20 | 71-125 | 71-125 | 90 |
| Tetrachloroethene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34478 | 10 | 30 | 70-130 | 70-130 | 90 |
| 1,2,4-Trichlorobenzene in water | µg/L | Primary | EPA 8270C | GC/MS | 34551 | 4 | 30 | 44-142 | 44-142 | 90 |
| 1,2,4-Trichlorobenzene in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34554 | 133 | 30 | 34-152 | 34-152 | 90 |
| Trichloroethylene in water | µg/L | Primary | EPA 8260B | GC/MS | 39180 | 1 | 20 | 71-125 | 71-125 | 90 |
| Trichloroethylene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34487 | 10 | 40 | 60-170 | 60-170 | 90 |
| 1,1,1-trichloro-ethane in water | µg/L | Primary | EPA 8260B | GC/MS | 34506 | 1 | 20 | 75-125 | 75-125 | 90 |
| 1,1,1-trichloro-ethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34509 | 5 | 25 | 70-130 | 70-130 | 90 |
| 1,1,2-trichloro-ethane in water | µg/L | Primary | EPA 8260B | GC/MS | 34511 | 1 | 20 | 75-127 | 75-127 | 90 |
| 1,1,2-trichloro-ethane in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34514 | 5 | 25 | 70-130 | 70-130 | 90 |
| 2,4,5-Trichlorophenol in water | µg/L | Primary | EPA 8270C | GC/MS | 77687 | 4 | 30 | 25-175 | 25-175 | 90 |
| 2,4,5-Trichlorophenol in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 78401 | 133 | 30 | 25-175 | 25-175 | 90 |
| 2,4,6-Trichlorophenol in water | µg/L | Primary | EPA 8270C | GC/MS | 34621 | 4 | 30 | 39-128 | 39-128 | 90 |
| 2,4,6-Trichlorophenol in sediment | µg/kg | Primary | EPA 8270C | GC/MS | 34624 | 133 | 30 | 29-138 | 29-138 | 90 |
| Toluene in water | µg/L | Primary | EPA 8260B | GC/MS | 34010 | 1 | 20 | 74-125 | 74-125 | 90 |
| Toluene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34483 | 10 | 30 | | | 90 |
| Toxaphene in water | µg/L | Primary | EPA 8081 | GC/ECD | 39400 | 1.0 | 25 | 28-131 | 28-131 | 90 |
| | | Alternate/Confirmatory | EPA 1656 | GC/ECD | 39400 | 2.7 | 25 | 76-122 | 76-122 | 90 |
| | | Alternate/Confirmatory | EPA 525.1 | L/S Extraction + Capillary GC/MS | 39400 | 20 | 25 | | | 90 |
| Toxaphene in sediment | µg/kg | Primary | EPA 8081 | GC/ECD | 39403 | 500 | 30 | 21-113 | 21-113 | 90 |
| | µg/kg | Alternate | EPA 1656 | GC/ECD | 39403 | 5.0 | 25 | 76-122 | 76-122 | 90 |

Appendix F Data Quality Objectives for Measurement Data

| | | | | | | | | | | |
|---|--------------------|----------------------------|--------------------------|--|----------------------|-------|----|--------|--------|----|
| Vinyl Chloride in water | µg/L | Primary | EPA 8260B | GC/MS | 39175 | 1 | 20 | 46-134 | 46-134 | 90 |
| Vinyl Chloride in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 34495 | 10 | 40 | 15-325 | 15-325 | 90 |
| m,p-xylene in water | µg/L | Primary | EPA 8260B | GC/MS | 85795 | 1 | 20 | 75-125 | 75-125 | 90 |
| o-xylene in water | µg/L | Primary | EPA 8260B | GC/MS | 77135 | 1 | 20 | 75-125 | 75-125 | 90 |
| m,p-xylene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 45516 | 10 | 30 | 70-130 | 70-130 | 90 |
| o-xylene in sediment | µg/kg | Primary | EPA 8260B | GC/MS | 78402 | 10 | 30 | 70-130 | 70-130 | 90 |
| Tributyltin in water | µg/L | Primary | EV-024/025 | | 30340 | 0.010 | 25 | | | 90 |
| Toxicity in ambient marine water | % Survival Yes/No* | <i>Mysidopsis bahia</i> | EPA 600-4-91-003; 1007.0 | Chronic Toxicity Screening Test | 89805 | NA | NA | NA | NA | 90 |
| Toxicity in ambient marine water | % Survival Yes/No* | <i>Menidia Berrylina</i> | EPA 600-4-91-003; 1006.0 | Chronic Toxicity Screening Test | 89806 | NA | NA | NA | NA | 90 |
| Toxicity in marine sediment | % Survival Yes/No* | <i>Leptocheirus</i> | EPA 600-R-94-025; 100.4 | Whole Sediment Toxicity Test | 89815 | NA | NA | NA | NA | 90 |
| Toxicity in marine sediment | % Survival Yes/No* | <i>Neanthes</i> | EPA 823-B-98-004 | Whole Sediment Toxicity Test | 89816 | NA | NA | NA | NA | 90 |
| Freshwater toxicity | % Survival Yes/No* | <i>Ceriodaphnia dubia</i> | EPA 600-4-91-002; 1002.0 | 7-day subchronic test for survival, reproduction | 89802 | NA | NA | NA | NA | 90 |
| Freshwater toxicity | % Survival Yes/No* | <i>Pimephales promelas</i> | EPA 600-4-91-002; 1000.0 | 7-day test for larval survival, growth | 89803 | NA | NA | NA | NA | 90 |
| Toxicity for freshwater whole sediments | % Survival Yes/No | <i>Hyallela azteca</i> | EPA 600-R-94-024; 100.1 | 10-day survival test for sediments | 89813 | NA | NA | NA | NA | 90 |
| Toxicity for freshwater whole sediments | % Survival Yes/No | <i>Chironomus tentans</i> | EPA 600-R-94-024; 100.2 | 10-day survival and growth tests for sediments | 89814 | NA | NA | NA | NA | 90 |
| Benthic Macro invertebrate sampling | number | counts | TNRCC SOP | TNRCC SOP | Texas Species Code** | NA | NA | NA | NA | 90 |
| Nekton Sampling | number | counts | TNRCC SOP | TNRCC SOP | Texas Species Code** | NA | NA | NA | NA | 90 |
| Stream Habitat | NA | Counts | TNRCC SOP | TNRCC SOP | NA | NA | NA | NA | NA | 90 |
| Sediment Core Upper Depth | Inches | Grab | TNRCC SOP | TNRCC SOP | 81900 | NA | NA | NA | NA | 90 |
| Sediment Core Lower Depth | Inches | Grab | TNRCC SOP | TNRCC SOP | 81901 | NA | NA | NA | NA | 90 |

* 1 = toxic; 2 = sublethal; 3 = none

** Individual species will be reported by TNRCC species code (TNRCC 1999)

DATA VERIFICATION REPORT
for sediment samples collected from Segment 1209A
BRYAN MUNICIPAL LAKE TMDL SITE

May 21, 2001

Data Verification by: Sandra de las Fuentes

The following data verification summary report covers environmental sediment samples collected from the Bryan Municipal Lake Segment 1209A, Station 11793, on May 21, 2001.

A Chemist with Parsons has reviewed the data submitted by DHL Analytical, B&B Laboratories, APPL, Inc. and The University of North Texas.

The sample in this event was analyzed for volatiles, semivolatiles, pesticides (including triazines, PCBs, organophosphorus compounds and herbicides), total metals, anions, simultaneously extracted metals (SEM), acid volatile sulfide (AVS), total organic carbon (TOC) and grain size.

Analysis for carbamates by USEPA SW846 Method 8321A (which includes carbaryl and diuron) was not performed due to a laboratory oversight.

There were no field quality control samples collected at this site. No trip blanks were analyzed for volatiles and no field blanks or equipment blanks were collected in association with the sediment samples in this DVR. Therefore, the possibility of contamination during sampling or handling could not be evaluated for these samples.

All samples were collected by Parsons and were analyzed by the various laboratories following procedures outlined in the Assessment of the Presence and Causes of Ambient Toxicity Quality Assurance Project Plan (QAPP).

REVIEW CRITERIA

All data submitted by the various laboratories has been reviewed. Field and laboratory QC sample information was examined, including: laboratory blanks, laboratory control samples (LCS), laboratory duplicates, standard reference material (SRM) samples, matrix spikes and matrix spike duplicate (MS and MSD) samples, surrogate spikes and Chain-of-Custody (COC) forms. The findings presented in this report are based on the reviewed information and whether the requirements specified in the project QAPP were met.

VOLATILES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001 and was analyzed for volatile organic compounds (VOCs). The VOC analyses were performed using USEPA SW846 Method 8260B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results LCS sample and surrogate spikes. A sample (11799-3) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group. It should be noted that only a small subset of analytes was reported for the MS/MSD.

The percent recoveries for the LCS were all within acceptance criteria.

The percent recoveries for the MS/MSD were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All volatile results for the samples in this report were considered usable. The completeness for the VOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEMIVOLATILES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for semivolatile organic compounds (SVOCs). The SVOC analyses were performed using USEPA SW846 Method 8270C.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS samples, and the surrogate spikes. A sample (10643-2) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group. It should be noted that only a small subset of analytes was reported for the MS/MSD.

All MS/MSD and surrogate %Rs were within acceptance criteria.

All LCS %Rs were within acceptance criteria.

All of the surrogate recoveries were within laboratory specified acceptance criteria for the LCS and MB except for the following:

| Sample | Analyte | %R | QC Criteria |
|--------|----------------------|-----|-------------|
| LCS | 2,4,6-Tribromophenol | 135 | 19-122 |
| MB | 4-terphenyl-d14 | 141 | 18-137 |

Since this surrogate compound was above control limits and all the percent recoveries for the LCS compounds were within acceptance criteria, no corrective action was taken. No action was taken for the non-compliant surrogate recovery in the MB since this surrogate compound was only slightly above control limits.

All of the surrogate recoveries for sample 11793-3 were within laboratory specified acceptance criteria except for the following:

| Sample | Analyte | %R | QC Criteria |
|---------|----------------------|-----|-------------|
| 11793-3 | 2,4,6-Tribromophenol | 158 | 19-122 |
| | 4-Terphenyl-d14 | 147 | 18-137 |

The sample was not flagged for the non-conformance surrogate compounds since the surrogates were above control limits and the sample was non-detect for all semi-volatile compounds.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria with the exception of the following:

| Analyte | MS %R | MSD %R | %RPD | QC Criteria |
|-------------------|-------|--------|------|-------------|
| pentachlorophenol | 72.5 | 53.2 | 30.7 | 30% |

Pentachlorophenol was only slightly above laboratory specified acceptance criteria. No corrective action was taken since the sample spiked was from a different TMDL site.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All semivolatile results for the samples in this report were considered usable. The completeness for the SVOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TRIAZINES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for triazine. The triazine compounds, atrazine, cyanazine, metolachlor and simazine, were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS sample and surrogate spikes. Sample, 10643-2(ARF 35491) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the triazine analyses. The blank was free of any triazines above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All triazine results for the sample in this report were considered usable. The completeness for the triazine portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

PESTICIDES / PCBS

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for pesticides and PCBs. The pesticide/PCB analyses were performed using USEPA SW846 Method 8081A/8082.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and surrogate spikes. Sample, 10643-2(ARF 35491) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria except for the following:

| Analyte | LCS %R | Lab Tolerance |
|---------|--------|---------------|
| Dicofol | 240 | 50-150 |

Dicofol was recovered high in the LCS by laboratory acceptance criteria. The QAPP did not provide accuracy acceptance criteria, therefore non-detect results in the sample were not flagged.

All MS/MSD percent recoveries were within acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | Tolerance |
|--------------|--------|--------|-----------|
| Aldrin | 42.5 | 37.4 | 46-155 |
| b-BHC | (55.2) | 46.0 | 51-133 |
| chlordane | (56.9) | 52.4 | 56-142 |
| DDE | (64.3) | 53.6 | 58-127 |
| DDT | (41.8) | 34.1 | 36-129 |
| Endosulfan | (61.7) | 51.2 | 56-142 |
| Methoxychlor | (39.8) | 33.2 | 37-144 |
| PCB-1016 | 120 | 135 | 56-113 |

() indicates recovery met criteria.

The sample in this data set not flagged for the non-compliant %Rs since the MS/MSD sample was taken from another TMDL site.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the pesticide/PCB analyses. The blank was free of any pesticides or PCBs of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All pesticide/PCB results for the samples in this report were considered usable. The completeness for the pesticide/PCB portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ORGANOPHOSPHORUS COMPOUNDS

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for organophosphorus compounds. The organophosphorus compounds, Chloropyrifos, Demeton, Diazinon, Guthion, Malathion and Parathion were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples, and surrogate spikes. Sample, 10643-2(ARF 35491) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the organophosphorus compound analyses. The blank was free of any organophosphorus compounds above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All organophosphorus compound results for the sample in this report were considered usable. The completeness for the organophosphorus compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

HERBICIDES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for herbicides. Herbicides, 2,4,5-T, 2,4,5-TP (Silvex) and 2,4-D, were analyzed using USEPA SW846 Method 8151A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and the surrogate spike. Sample, 10643-2(ARF 35491) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria with the exception of the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|----------------|--------------|---------------|--------------------|
| 2,4-D | 69.1 | 69.8 | 89-175 |

The MS/MSD %Rs were below acceptance criteria, although no flags were applied to the non-detected results for this compound since the MS/MSD sample was taken from another TMDL site.

The surrogate spike recovery met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

The method blank was run in association with the herbicide analyses. The blank was free of any herbicides above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All herbicide results for the samples in this report were considered usable. The completeness for the herbicide portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOTAL METALS AND IONS

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for total metals (aluminum, arsenic, barium, cadmium, calcium, chromium, copper, iron, lead, magnesium, mercury, nickel, potassium, selenium, silver, sodium and zinc). The mercury analyses were performed using USEPA SW846 Method 7471A. All other metals were determined using USEPA SW846 Method 6020B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. A sample (10643-2) from another TMDL site was selected as the MS/MSD for this QC batch for total metals. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group. A sample from another client was used as the batch QC for the MS/MSD for mercury.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria except for the following:

| Analyte | MS %R | MS %R | QC Criteria |
|-----------|-------|-------|-------------|
| Aluminum | -131 | -111 | 80-120% |
| Barium | 73.2 | 78.8 | |
| Calcium | 49.6 | 55.5 | |
| Iron | -77.4 | -45.2 | |
| Lead | 69.6 | 58.7 | |
| Magnesium | 58.2 | 60.5 | |
| Potassium | 62.5 | 65.7 | |
| Sodium | 53.2 | 54.3 | |
| Zinc | 76.1 | 78.6 | |

There were no flags added since the sample used for the MS/MSD was from a different TMDL site as the sample in this data set.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

No calibration, analytical spike or dilution test information was provided for the analyses.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ANIONS (CHLORIDE AND SULFATE)

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for chloride and sulfate using USEPA SW846 Method 9056.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and LCSD samples.

All LCS and LCSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the LCS/LCSD recoveries.

LCS/LCSD RPDs were within laboratory specified acceptance criteria for chloride and sulfate.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEM IN SEDIMENT

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for Simultaneously Extracted Metals (SEM), including cadmium, copper, lead, mercury, nickel, silver and zinc.

The metals analyses were performed using a modified EPA 1620 method, which is equivalent to EPA 200.7 and EPA 245.5.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. Another client's sample was used for the MS/MSD for the batch QC for this group. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met QAPP acceptance criteria.

There was no accuracy data provided for silver and mercury.

No accuracy criteria for the MS/MSD samples were listed in the QAPP for the SEM analyses. The tolerances listed for metals analyses were used to evaluate the MS/MSD samples.

All MS %Rs met the QAPP metals acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|---------|-------|--------|-------------|
| Copper | 76 | 79 | 80-120% |
| Lead | (109) | 265 | |
| Zinc | 136 | (101) | |

() indicates recovery met criteria

Because no tolerances were specified in the QAPP for SEM matrix spike accuracy and since this sample is from another client, no corrective action was necessary.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | RPD | QC Limits |
|---------|-------|--------|-----|-----------|
| Lead | 109 | 265 | 84% | 20% |

Since this sample is from another client, no corrective action was necessary.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time specified in the QAPP.

All laboratory blanks were reviewed and found to be free of SEM above the MAL, except for the following:

| Sample ID | Analyte | Conc. (ug/dry g) | MDL (ug/dry g) |
|-----------|---------|---------------------|-------------------|
| MB | Zinc | 3.09 | 0.24 |

No flags were applied since the result for zinc in the sample was greater than 5 times the result in the method blank.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All SEM results for the samples in this report were considered usable. The completeness for the SEM portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

AVS IN SEDIMENT

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for Acid Volatile Sulfide (AVS). The AVS analyses were performed using EPA method 376.3.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. Another client's sample was used for the MS/MSD for the batch QC for this group. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the QAPP.

All laboratory blanks were reviewed and found to be free of AVS at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All AVS results for the samples in this report were considered usable. The completeness for the AVS portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOC

General

This sample group consisted of two (2) samples, one (1) environmental sediment sample and one laboratory duplicate. The samples were collected on May 21 2001, and were analyzed for total organic carbon (TOC).). The TOC analyses were performed using B&B Laboratories, Inc. Standard Operating Procedure 1005.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the standard reference material (SRM) samples.

TOC met acceptance criteria in both SRM samples analyzed.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the laboratory duplicate. Sample "Dup(11793-3)" was randomly selected by the laboratory and analyzed as a laboratory duplicate of sample "11793-3."

The laboratory duplicate RPD was within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

Two method blanks were analyzed in association with the samples. Both blanks were free of TOC at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All TOC results for the samples in this report were considered usable. The completeness for the TOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

GRAIN SIZE

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for grain size by GS-92-01-B&B Method. Grain size results are reported as a percent of sand, silt or clay based on the weight of the sample.

Accuracy

Accuracy could not be evaluated by this method.

Precision

Precision could not be evaluated by this method.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

There were no method blanks required by this method.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All results for grain size for the sample in this report were considered usable. The completeness for the grain size compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

DATA VERIFICATION REPORT
for sediment samples collected from Segment 1209A
BRYAN MUNICIPAL LAKE TMDL SITE

July 18, 2001

Data Verification by: Sandra de las Fuentes

The following data verification summary report covers environmental sediment samples collected from the Bryan Municipal Lake Segment 1209A, Station 11793, on July 18, 2001.

A Chemist with Parsons has reviewed the data submitted by DHL Analytical, B&B Laboratories, APPL, Inc. and The University of North Texas.

The sample in this event was analyzed for volatiles, semivolatiles, pesticides (including triazines, PCBs, organophosphorus compounds and herbicides), total metals, anions, simultaneously extracted metals (SEM), acid volatile sulfide (AVS), total organic carbon (TOC) and grain size.

There were no field quality control samples collected at this site. No trip blanks were analyzed for volatiles and no field blanks or equipment blanks were collected in association with the sediment samples in this DVR. Therefore, the possibility of contamination during sampling or handling could not be evaluated for these samples.

All samples were collected by Parsons and were analyzed by the various laboratories following procedures outlined in the Assessment of the Presence and Causes of Ambient Toxicity Quality Assurance Project Plan (QAPP).

REVIEW CRITERIA

All data submitted by the various laboratories has been reviewed. Field and laboratory QC sample information was examined, including: laboratory blanks, laboratory control samples (LCS), laboratory duplicates, standard reference material (SRM) samples, matrix spikes and matrix spike duplicate (MS and MSD) samples, surrogate spikes and Chain-of-Custody (COC) forms. The findings presented in this report are based on the reviewed information and whether the requirements specified in the project QAPP were met.

VOLATILES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001 and was analyzed for volatile organic compounds (VOCs). The VOC analyses were performed using USEPA SW846 Method 8260B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results LCS sample and surrogate spikes. Another client's sample was used for the MS/MSD for the batch QC for this group. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group. It should be noted that only a small subset of analytes was reported for the MS/MSD.

The percent recoveries for the LCS were all within acceptance criteria, except for the following:

| Analyte | LCS %R | QC Criteria |
|---------------------------|--------|-------------|
| 1,1-Dichloroethene | 66.5 | 70-130% |
| Trans-1, 2-dichloroethene | 68.5 | 70-130% |

The reported concentrations for these analytes were considered estimated (possibly biased low) and were flagged "J" if detected or "UJ" if non-detect.

The percent recoveries for the MS/MSD were within acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|--------------------|--------|--------|-------------|
| 1,1-Dichloroethene | (75.8) | 69.2 | 70-130% |
| Toluene | (74.8) | 66.4 | 70-130% |

() indicates recovery met criteria.

No action was taken since the sample spiked was taken from another client.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All volatile results for the samples in this report were considered usable. The completeness for the VOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEMIVOLATILES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for semivolatile organic compounds (SVOCs). The SVOC analyses were performed using USEPA SW846 Method 8270C.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS samples, and the surrogate spikes. A sample from another TMDL site (11799-5) was selected by the laboratory as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group. It should be noted that only a small subset of analytes was reported for the MS/MSD.

All MS/MSD and surrogate %Rs were within acceptance criteria.

All LCS %Rs were within acceptance criteria.

All of the surrogate recoveries were within laboratory specified acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All semivolatile results for the samples in this report were considered usable. The completeness for the SVOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TRIAZINES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for triazine. The triazine compounds, atrazine, cyanazine, metolachlor and simazine, were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS sample and surrogate spikes. A sample from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the triazine analyses. The blank was free of any triazines above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All triazine results for the sample in this report were considered usable. The completeness for the triazine portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

PESTICIDES / PCBS

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for pesticides and PCBs. The pesticide/PCB analyses were performed using USEPA SW846 Method 8081A/8082.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and surrogate spikes. A sample from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | Tolerance |
|--------------|-------|--------|-----------|
| Methoxychlor | 34.3 | (41.6) | 37-144 |
| DDT | 26.5 | 32.6 | 36-129 |

() indicates recovery met criteria.

The sample in this data set was not flagged for the non-compliant %Rs since the spiked sample was taken from another TMDL site.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the pesticide/PCB analyses. The blank was free of any pesticides or PCBs of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All pesticide/PCB results for the samples in this report were considered usable. The completeness for the pesticide/PCB portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ORGANOPHOSPHORUS COMPOUNDS

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for organophosphorus compounds. The organophosphorus compounds, Chloropyrifos, Demeton, Diazinon, Guthion, Malathion and Parathion were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples, and surrogate spikes. A sample from another TMDL site was

selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the organophosphorus compound analyses. The blank was free of any organophosphorus compounds above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All organophosphorus compound results for the sample in this report were considered usable. The completeness for the organophosphorus compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

HERBICIDES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for herbicides. Herbicides, 2,4,5-T, 2,4,5-TP (Silvex) and 2,4-D, were analyzed using USEPA SW846 Method 8151A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and the surrogate spike. A sample from another TMDL site was

selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

The surrogate spike recovery met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

The method blank was run in association with the herbicide analyses. The blank was free of any herbicides above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All herbicide results for the samples in this report were considered usable. The completeness for the herbicide portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

CARBAMATES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for carbamates. The carbamate compounds, carbaryl and diuron were analyzed using USEPA SW846 Method 8321A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and surrogate spikes. A sample from another TMDL site was selected

as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria except for the following:

| Analyte | MS %R | Tolerance |
|---------|-------|-----------|
| Diuron | 163 | 25-133 |

The sample in this data set was not flagged for the non-compliant %Rs since the spiked sample was taken from another TMDL site.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

The MS/MSD RPDs were outside of laboratory specified acceptance criteria as indicated in the following:

| Analyte | MS %R | MSD %R | RPD | Lab Tolerance |
|----------|-------|--------|------|---------------|
| Carbaryl | 41.4 | 63.7 | 42.3 | 25% |
| Diuron | 100 | 163 | 47.9 | |

The sample in this data set was not flagged for the non-compliant %Rs since the spiked sample was taken from another TMDL site.

All field duplicate RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the carbamate analyses. The blank was free of any carbamates of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All carbamate results for the samples in this report were considered usable. The completeness for the carbamates portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOTAL METALS AND IONS

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for total metals (aluminum, arsenic, barium, cadmium, calcium, chromium, copper, iron, lead, magnesium, mercury, nickel, potassium, selenium, silver, sodium and zinc). The mercury analyses were performed using USEPA SW846 Method 7471A. All other metals were determined using USEPA SW846 Method 6020B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. Another client's sample was used for the MS/MSD for the batch QC for both the total metals and mercury analyses. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria except for the following:

| Analyte | MS %R | MS %R | QC Criteria |
|-----------|-------|--------|-------------|
| Aluminum | -141 | 202 | 80-120% |
| Barium | 79.3 | (99.2) | 80-120% |
| Calcium | -174 | -253 | 80-120% |
| Iron | -167 | -11.8 | 80-120% |
| Magnesium | 67.4 | (114) | 80-120% |
| Mercury | (113) | 127 | 77-120% |
| Potassium | 65.4 | (80) | 80-120% |

() indicates recovery met criteria.

For aluminum, calcium, iron and magnesium, the sample concentration was significantly greater (over 4 times) than the spike concentration. The result for barium and potassium may be biased low in the sample, although no flag was applied since the sample spiked

was taken from another client. The result for mercury may be biased high, although no flag was applied since the sample was taken from another client.

There were no flags added since the sample used for the MS/MSD was from another client sample.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria except for the following:

| Analyte | MS Conc. (mg/Kg-dry wt) | MSD Conc. (mg/Kg-dry wt) | RPD | RPD Limits |
|----------|----------------------------|-----------------------------|-------|------------|
| Arsenic | 101 | 76.25 | 28% | 25% |
| Cadmium | 98.47 | 73.84 | 28.6% | |
| Selenium | 87.48 | 67.33 | 26% | |

The laboratory calculates the %RPD using the concentration results for the metals based on dry weight. The actual recoveries for the metals listed above were all well within acceptance limits, therefore no corrective action were required.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

No calibration, analytical spike or dilution test information was provided for the analyses.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ANIONS (CHLORIDE AND SULFATE)

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for chloride and sulfate using USEPA SW846 Method 9056.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and LCSD samples.

All LCS and LCSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the LCS/LCSD recoveries.

LCS/LCSD RPDs were within laboratory specified acceptance criteria for chloride and sulfate.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEM IN SEDIMENT

General

This sample group consisted of four (4) samples, including one environmental sediment sample, one field duplicate sample and one pair of MS/MSD samples. The samples were collected on July 18, 2001, and were analyzed for Simultaneously Extracted Metals (SEM), including cadmium, copper, lead, mercury, nickel, silver and zinc.

The metals analyses were performed using a modified EPA 1620 method, which is equivalent to EPA 200.7 and EPA 245.5.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. A sample from another TMDL site was analyzed as the MS/MSD sample for this data set. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met QAPP acceptance criteria.

No accuracy criteria for the MS/MSD samples were listed in the QAPP for the SEM analyses. The tolerances listed for metals analyses were used to evaluate the MS/MSD samples.

All MS/MSD %Rs met the QAPP metals acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|---------|-------|--------|-------------|
| Silver | 0 | 0 | |
| Cadmium | 72 | (86) | |
| Copper | 0 | 0 | 80-120% |
| Lead | 0 | 52 | |
| Zinc | 65 | 147 | |

() indicates recovery met criteria

The laboratory explained the observed variances as a product of sample inhomogeneity and matrix interference. This sample was analyzed in duplicate as shown below. As a result of the high variances in both the MS/MSD spike results and the duplicate data, the concentrations for the above compounds were considered estimated although no flags were applies since the sample spiked was taken from a different TMDL site.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria with the exception of the following:

| Analyte | MS Conc (ug/kg) | MSD Conc. (ug/kg) | RPD | QC Limits |
|---------|--------------------|----------------------|-----|--------------|
| Lead | 21.6 | 33.1 | 42% | 20% |

There were no flags applied to the samples since the sample spiked was taken from a different TMDL site.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time specified in the QAPP.

All laboratory blanks were reviewed and found to be free of SEM above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All SEM results for the samples in this report were considered usable. The completeness for the SEM portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

AVS IN SEDIMENT

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for Acid Volatile Sulfide (AVS). The AVS analyses were performed using EPA method 376.3.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS. A sample from another TMDL site was used for the MS/MSD for the batch QC for this group. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the QAPP.

All laboratory blanks were reviewed and found to be free of AVS at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All AVS results for the samples in this report were considered usable. The completeness for the AVS portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOC

General

This sample group consisted of two (2) samples, one (1) environmental sediment sample and one laboratory duplicate. The samples were collected on July 18, 2001, and were analyzed for total organic carbon (TOC). The TOC analyses were performed using B&B Laboratories, Inc. Standard Operating Procedure 1005.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the standard reference material (SRM) samples.

TOC met acceptance criteria in both SRM samples analyzed.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the laboratory duplicate. Sample, 11793-5 (PAR0018D), was randomly selected by the laboratory and analyzed as a laboratory duplicate of sample, 11793-5 (PAR0018).

The laboratory duplicate RPD was within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

Two method blanks were analyzed in association with the samples. Both blanks were free of TOC at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All TOC results for the samples in this report were considered usable. The completeness for the TOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

GRAIN SIZE

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for grain size by GS-92-01-B&B Method. Grain size results are reported as a percent of sand, silt or clay based on the weight of the sample.

Accuracy

Accuracy could not be evaluated by this method.

Precision

Precision could not be evaluated by this method.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

There were no method blanks required by this method.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All results for grain size for the sample in this report were considered usable. The completeness for the grain size compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

DATA VERIFICATION REPORT
for sediment samples collected from Segment 1209A
BRYAN MUNICIPAL LAKE TMDL SITE

July 12, 2002

Data Verification by: Sandra de las Fuentes

The following data verification summary report covers environmental sediment samples collected from the Bryan Municipal Lake Segment 1209A, Stations 11792, 11793 and 11794, on July 12, 2002.

A Chemist with Parsons has reviewed the data submitted by DHL Analytical, APPL, Inc. and The University of North Texas.

The sample in this event was analyzed for volatiles, semivolatiles, pesticides (including triazines, PCBs, organophosphorus compounds and herbicides), total metals, anions, simultaneously extracted metals (SEM), acid volatile sulfide (AVS), total organic carbon (TOC) and grain size.

There were no field quality control samples collected at this site. No trip blanks were analyzed for volatiles and no field blanks or equipment blanks were collected in association with the sediment samples in this DVR. Therefore, the possibility of contamination during sampling or handling could not be evaluated for these samples.

All samples were collected by Parsons and were analyzed by the various laboratories following procedures outlined in the Assessment of the Presence and Causes of Ambient Toxicity Quality Assurance Project Plan (QAPP).

REVIEW CRITERIA

All data submitted by the various laboratories has been reviewed. Field and laboratory QC sample information was examined, including: laboratory blanks, laboratory control samples (LCS), laboratory duplicates, matrix spikes and matrix spike duplicate (MS and MSD) samples, surrogate spikes and Chain-of-Custody (COC) forms. The findings presented in this report are based on the reviewed information and whether the requirements specified in the project QAPP were met.

VOLATILES

General

This sample group consisted of five (5) samples, including three (3) environmental sediment samples, and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on July 12, 2002 and were analyzed for volatile organic compounds (VOCs). The VOC analyses were performed using USEPA SW846 Method 8260B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results LCS sample and surrogate spikes. Sample 11792-10 was randomly selected by the laboratory as the MS/MSD for this QC batch. It should be noted that only a small subset of analytes was reported for the MS/MSD.

The percent recoveries for the LCS were all within acceptance criteria.

The percent recoveries for the MS/MSD were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All volatile results for the samples in this report were considered usable. The completeness for the VOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEMIVOLATILES

General

This sample group consisted of five (5) samples, including three (3) environmental sediment samples, and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on July 12, 2002, and were analyzed for semivolatile organic compounds (SVOCs). The SVOC analyses were performed using USEPA SW846 Method 8270C.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS samples, and the surrogate spikes. Sample 11792 was randomly selected by the laboratory as the MS/MSD for this QC batch. It should be noted that only a small subset of analytes was reported for the MS/MSD.

All MS/MSD and surrogate %Rs were within acceptance criteria.

All LCS %Rs were within acceptance criteria.

All of the surrogate recoveries were within laboratory specified acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All semivolatile results for the samples in this report were considered usable. The completeness for the SVOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TRIAZINES

General

This sample group consisted of three (3) environmental sediment samples. The samples were collected on July 12, 2002, and were analyzed for triazine. The triazine compounds, atrazine, cyanazine, metolachlor and simazine, were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample and surrogate spikes.

The LCS percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the triazine analyses. The blank was free of any triazines above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All triazine results for the sample in this report were considered usable. The completeness for the triazine portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

PESTICIDES / PCBS

General

This sample group consisted of three (3) environmental sediment samples. The samples were collected on July 12, 2002, and were analyzed for pesticides and PCBs. The pesticide/PCB analyses were performed using USEPA SW846 Method 8081A/8082.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample and surrogate spikes.

The LCS percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the pesticide/PCB analyses. The blank was free of any pesticides or PCBs of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All pesticide/PCB results for the samples in this report were considered usable. The completeness for the pesticide/PCB portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ORGANOPHOSPHORUS COMPOUNDS

General

This sample group consisted of three (3) environmental sediment samples. The samples were collected on July 12, 2002, and were analyzed for organophosphorus compounds. The organophosphorus compounds, Chlorpyrifos, Demeton, Diazinon, Guthion, Malathion and Parathion were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample and surrogate spikes.

The LCS percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the organophosphorus compound analyses. The blank was free of any organophosphorus compounds above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All organophosphorus compound results for the sample in this report were considered usable. The completeness for the organophosphorus compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

HERBICIDES

General

This sample group consisted of five (5) samples, including three (3) environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on July 12, 2002, and were analyzed for herbicides. Herbicides, 2,4,5-T, 2,4,5-TP (Silvex) and 2,4-D, were analyzed using USEPA SW846 Method 8151A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and the surrogate spike. Sample 11793-10 was randomly selected by the laboratory as the MS/MSD for this QC batch.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

The surrogate spike recovery met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

The method blank was run in association with the herbicide analyses. The blank was free of any herbicides above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All herbicide results for the samples in this report were considered usable. The completeness for the herbicide portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

CARBAMATES

General

This sample group consisted of three (3) environmental sediment samples. The samples were collected on July 12, 2002, and were analyzed for carbamates. The carbamate compounds, carbaryl and diuron were analyzed using USEPA SW846 Method 8321A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample and surrogate spikes.

The LCS percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the carbamate analyses. The blank was free of any carbamates of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All carbamate results for the samples in this report were considered usable. The completeness for the carbamates portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOTAL METALS AND IONS

General

This sample group consisted of three (3) environmental sediment samples. The samples were collected on July 12, 2002, and were analyzed for total metals (aluminum, arsenic, barium, cadmium, calcium, chromium, copper, iron, lead, magnesium, mercury, nickel, potassium, selenium, silver, sodium and zinc). The mercury analyses were performed using USEPA SW846 Method 7471A. All other metals were determined using USEPA SW846 Method 6020B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS sample.

All LCS %Rs met acceptance criteria.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

No calibration, analytical spike or dilution test information was provided for the analyses.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ANIONS (CHLORIDE AND SULFATE)

General

This sample group consisted of four (4) samples, including three (3) environmental sediment samples and one laboratory duplicate sample, randomly selected by the laboratory. The samples were collected on July 12, 2002, and were analyzed for chloride and sulfate using USEPA SW846 Method 9056.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and LCSD samples.

All LCS and LCSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the LCS/LCSD recoveries and the laboratory duplicate sample results. Sample 11794 was randomly selected by the laboratory as the laboratory duplicate sample for this QC batch.

LCS/LCSD RPDs were within laboratory specified acceptance criteria for chloride and sulfate.

The laboratory duplicate RPDs were within QAPP acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEM IN SEDIMENT

General

This sample group consisted of three (3) environmental sediment samples. The samples were collected on July 12, 2002, and were analyzed for Simultaneously Extracted Metals (SEM), including cadmium, copper, lead, mercury, nickel and zinc.

The metals analyses were performed using a modified EPA 821 draft method, which is equivalent to EPA 200.7 and EPA 245.5.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS sample.

All LCS %Rs met QAPP acceptance criteria.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.
 All samples were prepared and analyzed within the hold time specified in the QAPP.
 All laboratory blanks were reviewed and found to be free of SEM above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All SEM results for the samples in this report were considered usable. The completeness for the SEM portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

AVS IN SEDIMENT

General

This sample group consisted of three (3) environmental sediment samples. The samples were collected on July 12, 2002, and were analyzed for Acid Volatile Sulfide (AVS). The AVS analyses were performed using EPA 821 draft method.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and LCSD samples.

The LCS and LCSD %Rs was slightly above laboratory acceptance criteria as shown in the following table:

| Analyte | LCS %R | LCSD %R | Laboratory Tolerance |
|---------|--------|---------|----------------------|
| AVS | 112 | (87) | 85-105% |

() indicates criteria was met

The QAPP doesn't specify tolerance criteria for the LCS for AVS. Since the %R for the LCS is only slightly above laboratory tolerance, there were no flags applied to the sample results for AVS.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the LCS/LCSD recoveries.

All LCS/LCSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;

- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the QAPP.

All laboratory blanks were reviewed and found to be free of AVS at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All AVS results for the samples in this report were considered usable. The completeness for the AVS portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOC

General

This sample group consisted of five (5) samples, including three (3) environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on July 12, 2002, and were analyzed for total organic carbon (TOC). The TOC analyses were performed using EPA 415.1.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS sample and the MS/MSD samples. Sample 11792-10 was randomly selected by the laboratory and analyzed as the MS/MSD for this data set.

TOC met acceptance criteria in the LCS sample analyzed.

TOC met acceptance criteria in the MS/MSD samples.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

MS/MSD RPDs were within laboratory specified acceptance criteria for TOC.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

The method blank was analyzed in association with the samples and was free of TOC at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All TOC results for the samples in this report were considered usable. The completeness for the TOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

GRAIN SIZE

General

This sample group consisted of three (3) environmental sediment samples. The samples were collected on July 12, 2002, and were analyzed for grain size by EPA 3.4, 3.5 (600/2-78-054). Grain size results are reported as a percent of gravel, sand, silt or clay based on the weight of the sample.

Accuracy

Accuracy could not be evaluated by this method.

Precision

Precision could not be evaluated by this method.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

There were no method blanks required by this method.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All results for grain size for the sample in this report were considered usable. The completeness for the grain size compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

DATA VERIFICATION REPORT
for sediment samples collected from Segment 1209B
FINFEATHER LAKE TMDL SITE

May 21, 2001

Data Verification by: Sandra de las Fuentes

The following data verification summary report covers environmental sediment samples collected from the Finfeather Lake Segment 1209B, Station 11799, on May 21, 2001.

A Chemist with Parsons has reviewed the data submitted by DHL Analytical, B&B Laboratories, APPL, Inc. and The University of North Texas.

The sample in this event was analyzed for volatiles, semivolatiles, pesticides (including triazines, PCBs, organophosphorus compounds and herbicides), total metals, anions, simultaneously extracted metals (SEM), acid volatile sulfide (AVS), total organic carbon (TOC) and grain size.

Analysis for carbamates by USEPA SW846 Method 8321A (which includes carbaryl and diuron) was not performed due to a laboratory oversight.

There were no field quality control samples collected at this site. No trip blanks were analyzed for volatiles and no field blanks or equipment blanks were collected in association with the sediment samples in this DVR. Therefore, the possibility of contamination during sampling or handling could not be evaluated for these samples.

All samples were collected by Parsons and were analyzed by the various laboratories following procedures outlined in the Assessment of the Presence and Causes of Ambient Toxicity Quality Assurance Project Plan (QAPP).

REVIEW CRITERIA

All data submitted by the various laboratories has been reviewed. Field and laboratory QC sample information was examined, including: laboratory blanks, laboratory control samples (LCS), laboratory duplicates, standard reference material (SRM) samples, matrix spikes and matrix spike duplicate (MS and MSD) samples, surrogate spikes and Chain-of-Custody (COC) forms. The findings presented in this report are based on the reviewed information and whether the requirements specified in the project QAPP were met.

VOLATILES

General

This sample group consisted of three (3) samples including one environmental sediment sample and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on May 21, 2001 and were analyzed for volatile organic compounds (VOCs). The VOC analyses were performed using USEPA SW846 Method 8260B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results LCS sample, MS and MSD recoveries and surrogate spikes. Sample 11799-3 was randomly selected by the laboratory as the MS/MSD for this QC batch. It should be noted that only a small subset of analytes was reported for the MS/MSD.

The percent recoveries for the LCS were all within acceptance criteria.

The percent recoveries for the MS/MSD were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All volatile results for the samples in this report were considered usable. The completeness for the VOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEMIVOLATILES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for semivolatile organic compounds (SVOCs). The SVOC analyses were performed using USEPA SW846 Method 8270C.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS samples, and the surrogate spikes. A sample (10643-2) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group. It should be noted that only a small subset of analytes was reported for the MS/MSD.

All MS/MSD and surrogate %Rs were within acceptance criteria.

All LCS %Rs were within acceptance criteria.

All of the surrogate recoveries were within laboratory specified acceptance criteria for the LCS and MB except for the following:

| Sample | Analyte | %R | QC Criteria |
|--------|----------------------|-----|-------------|
| LCS | 2,4,6-Tribromophenol | 135 | 19-122% |
| MB | 4-terphenyl-d14 | 141 | 18-137% |

Since this surrogate compound was above control limits and all the percent recoveries for the LCS compounds were within acceptance criteria, no corrective action was taken. No action was taken for the non-compliant surrogate recovery in the MB since this surrogate compound was only slightly above control limits.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria with the exception of the following:

| Analyte | MS %R | MSD %R | %RPD | QC Criteria |
|-------------------|-------|--------|------|-------------|
| pentachlorophenol | 72.5 | 53.2 | 30.7 | 30% |

Pentachlorophenol was only slightly above laboratory specified acceptance criteria. No corrective action was taken since the sample spiked was from a different TMDL site.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All semivolatile results for the samples in this report were considered usable. The completeness for the SVOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TRIAZINES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for triazine. The triazine compounds, atrazine, cyanazine, metolachlor and simazine, were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS sample and surrogate spikes. Sample, 10643-2(ARF 35491) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the triazine analyses. The blank was free of any triazines above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All triazine results for the sample in this report were considered usable. The completeness for the triazine portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

PESTICIDES / PCBs

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for pesticides and PCBs. The pesticide/PCB analyses were performed using USEPA SW846 Method 8081A/8082.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and surrogate spikes. Sample, 10643-2(ARF 35491) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria except for the following:

| Analyte | LCS %R | Lab Tolerance |
|---------|--------|---------------|
| Dicofol | 240 | 50-150% |

Dicofol was recovered high in the LCS by laboratory acceptance criteria. The QAPP did not provide accuracy acceptance criteria, therefore non-detect results in the sample were not flagged.

All MS/MSD percent recoveries were within acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | Tolerance |
|--------------|--------|--------|-----------|
| Aldrin | 42.5 | 37.4 | 46-155% |
| b-BHC | (55.2) | 46.0 | 51-133% |
| chlordane | (56.9) | 52.4 | 56-142% |
| DDE | (64.3) | 53.6 | 58-127% |
| DDT | (41.8) | 34.1 | 36-129% |
| Endosulfan | (61.7) | 51.2 | 56-142% |
| Methoxychlor | (39.8) | 33.2 | 37-144% |
| PCB-1016 | 120 | 135 | 56-113% |

() indicates recovery met criteria.

The sample in this data set not flagged for the non-compliant %Rs since the MS/MSD sample was taken from another TMDL site

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the pesticide/PCB analyses. The blank was free of any pesticides or PCBs of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All pesticide/PCB results for the samples in this report were considered usable. The completeness for the pesticide/PCB portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ORGANOPHOSPHORUS COMPOUNDS

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for organophosphorus compounds. The organophosphorus compounds, Chlorpyrifos, Demeton, Diazinon, Guthion, Malathion and Parathion were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples, and surrogate spikes. Sample, 10643-2 (ARF 35491) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the organophosphorus compound analyses. The blank was free of any organophosphorus compounds above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All organophosphorus compound results for the sample in this report were considered usable. The completeness for the organophosphorus compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

HERBICIDES

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for herbicides. Herbicides, 2,4,5-T, 2,4,5-TP (Silvex) and 2,4-D, were analyzed using USEPA SW846 Method 8151A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and the surrogate spike. Sample, 10643-2 (ARF 35491) from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria with the exception of the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|---------|-------|--------|-------------|
| 2,4-D | 69.1 | 69.8 | 89-175% |

The MS/MSD %R for 2,4-D were below acceptance criteria, although no flags were applied to the non-detected results for this compound since the MS/MSD sample was taken from another TMDL site.

The surrogate spike recovery met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

The method blank was run in association with the herbicide analyses. The blank was free of any herbicides above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All herbicide results for the samples in this report were considered usable. The completeness for the herbicide portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOTAL METALS AND IONS

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for total metals (aluminum, arsenic, barium, cadmium, calcium, chromium, copper, iron, lead, magnesium, mercury, nickel, potassium, selenium, silver, sodium and zinc). The mercury analyses were performed using USEPA SW846 Method 7471A. All other metals were determined using USEPA SW846 Method 6020B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. A sample (10643-2) from another TMDL site was selected as the MS/MSD for this QC batch for total metals. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group. A sample from another client was used as the batch QC for the MS/MSD for mercury.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|-----------|-------|--------|-------------|
| Aluminum | -131 | -111 | 80-120% |
| Barium | 73.2 | 78.8 | |
| Calcium | 49.6 | 55.5 | |
| Iron | -77.4 | -45.2 | |
| Lead | 69.6 | 58.7 | |
| Magnesium | 58.2 | 60.5 | |
| Potassium | 62.5 | 65.7 | |
| Sodium | 53.2 | 54.3 | |
| Zinc | 76.1 | 78.6 | |

There were no flags added since the sample used for the MS/MSD was from a different TMDL site as the sample in this group.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

No calibration, analytical spike or dilution test information was provided for the analyses.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ANIONS (CHLORIDE AND SULFATE)

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for chloride and sulfate using USEPA SW846 Method 9056.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and LCSD samples.

All LCS and LCSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the LCS/LCSD recoveries.

LCS/LCSD RPDs were within laboratory specified acceptance criteria for chloride and sulfate.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEM IN SEDIMENT

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for Simultaneously Extracted Metals (SEM), including cadmium, copper, lead, mercury, nickel, silver and zinc.

The metals analyses were performed using a modified EPA 1620 method, which is equivalent to EPA 200.7 and EPA 245.5.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. Another client’s sample was used for the MS/MSD for the batch QC for this group. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met QAPP acceptance criteria.

There was no accuracy data provided for silver and mercury.

No accuracy criteria for the MS/MSD samples were listed in the QAPP for the SEM analyses. The tolerances listed for metals analyses were used to evaluate the MS/MSD samples.

All MS %Rs met the QAPP metals acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|---------|-------|--------|-------------|
| Copper | 76 | 79 | 80-120% |
| Lead | (109) | 265 | |
| Zinc | 136 | (101) | |

() indicates recovery met criteria

Because no tolerances were specified in the QAPP for SEM matrix spike accuracy and since this sample is from another client, no corrective action was necessary.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | RPD | QC Tolerance |
|---------|-------|--------|-----|--------------|
| Lead | 109 | 265 | 84% | 20% |

Since this sample is from another client, no corrective action was necessary.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time specified in the QAPP.

All laboratory blanks were reviewed and found to be free of SEM above the MAL, except for the following:

| Sample ID | Analyte | Conc. (ug/dry g) | MDL (ug/dry g) |
|-----------|---------|---------------------|-------------------|
| MB | Zinc | 3.09 | 0.24 |

No flags were applied since the result for zinc in the sample was greater than 5 times the result in the method blank.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All SEM results for the samples in this report were considered usable. The completeness for the SEM portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

AVS IN SEDIMENT

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for Acid Volatile Sulfide (AVS). The AVS analyses were performed using EPA method 376.3.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. Another client's sample was used for the MS/MSD for the batch QC for this group. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the QAPP.

All laboratory blanks were reviewed and found to be free of AVS at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All AVS results for the samples in this report were considered usable. The completeness for the AVS portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOC

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for total organic carbon (TOC). The TOC analyses were performed using B&B Laboratories, Inc. Standard Operating Procedure 1005.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the standard reference material (SRM) samples.

TOC met acceptance criteria in both SRM samples analyzed.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

Two method blanks were analyzed in association with the samples. Both blanks were free of TOC at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All TOC results for the samples in this report were considered usable. The completeness for the TOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

GRAIN SIZE

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on May 21, 2001, and was analyzed for grain size by GS-92-01-B&B Method. Grain size results are reported as a percent of sand, silt or clay based on the weight of the sample.

Accuracy

Accuracy could not be evaluated by this method.

Precision

Precision could not be evaluated by this method.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

There were no method blanks required by this method.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All results for grain size for the sample in this report were considered usable. The completeness for the grain size compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

DATA VERIFICATION REPORT
for sediment samples collected from Segment 1209B
FINFEATHER LAKE TMDL SITE

July 18, 2001

Data Verification by: Sandra de las Fuentes

The following data verification summary report covers environmental sediment samples collected from the Finfeather Lake Segment 1209B, Stations 11798 and 11799, on July 18, 2001.

A Chemist with Parsons has reviewed the data submitted by DHL Analytical, B&B Laboratories, APPL, Inc. and The University of North Texas.

The sample in this event was analyzed for volatiles, semivolatiles, pesticides (including triazines, PCBs, organophosphorus compounds and herbicides), total metals, anions, simultaneously extracted metals (SEM), acid volatile sulfide (AVS), total organic carbon (TOC) and grain size.

There were no field quality control samples collected at this site. No trip blanks were analyzed for volatiles and no field blanks or equipment blanks were collected in association with the sediment samples in this DVR. Therefore, the possibility of contamination during sampling or handling could not be evaluated for these samples.

All samples were collected by Parsons and were analyzed by the various laboratories following procedures outlined in the Assessment of the Presence and Causes of Ambient Toxicity Quality Assurance Project Plan (QAPP).

REVIEW CRITERIA

All data submitted by the various laboratories has been reviewed. Field and laboratory QC sample information was examined, including: laboratory blanks, laboratory control samples (LCS), laboratory duplicates, standard reference material (SRM) samples, matrix spikes and matrix spike duplicate (MS and MSD) samples, surrogate spikes and Chain-of-Custody (COC) forms. The findings presented in this report are based on the reviewed information and whether the requirements specified in the project QAPP were met.

VOLATILES

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on July 18, 2001 and were analyzed for volatile organic compounds (VOCs). The VOC analyses were performed using USEPA SW846 Method 8260B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results LCS sample and surrogate spikes. Another client's sample was used for the MS/MSD for the batch QC for this group. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group. It should be noted that only a small subset of analytes was reported for the MS/MSD.

The percent recoveries for the LCS were all within acceptance criteria, except for the following:

| Analyte | LCS %R | QC Criteria |
|---------------------------|--------|-------------|
| 1,1-Dichloroethene | 66.5 | 70-130% |
| Trans-1, 2-dichloroethene | 68.5 | 70-130% |

The reported concentrations for these analytes were considered estimated (possibly biased low) and were flagged "J" for the samples if detected or "UJ" if non-detect.

The percent recoveries for the MS/MSD were within acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|--------------------|--------|--------|-------------|
| 1,1-Dichloroethene | (75.8) | 69.2 | 70-130% |
| Toluene | (74.8) | 66.4 | 70-130% |

() indicates recovery met criteria.

No action was taken since the sample spiked was taken from another client.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All volatile results for the samples in this report were considered usable. The completeness for the VOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEMIVOLATILES

General

This sample group consisted of four (4) samples, including two (2) environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on July 18, 2001, and were analyzed for semivolatile organic compounds (SVOCs). The SVOC analyses were performed using USEPA SW846 Method 8270C.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS samples, and the surrogate spikes. Sample 11799-5 was randomly selected by the laboratory as the MS/MSD for this QC batch. It should be noted that only a small subset of analytes was reported for the MS/MSD.

All MS/MSD and surrogate %Rs were within acceptance criteria.

All LCS %Rs were within acceptance criteria.

All of the surrogate recoveries were within laboratory specified acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All semivolatile results for the samples in this report were considered usable. The completeness for the SVOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TRIAZINES

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on July 18, 2001, and were analyzed for triazine. The triazine compounds, atrazine, cyanazine, metolachlor and simazine, were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the MS/MSD samples, LCS sample and surrogate spikes. A sample from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the triazine analyses. The blank was free of any triazines above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All triazine results for the sample in this report were considered usable. The completeness for the triazine portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

PESTICIDES / PCBS

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on July 18, 2001, and were analyzed for pesticides and PCBs. The pesticide/PCB analyses were performed using USEPA SW846 Method 8081A/8082.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and surrogate spikes. A sample from another TMDL site was selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | Tolerance |
|--------------|-------|--------|-----------|
| Methoxychlor | 34.3 | (41.6) | 37-144 |
| DDT | 26.5 | 32.6 | 36-129 |

() indicates recovery met criteria.

The samples in this data set were not flagged for the non-compliant %Rs since the spiked sample was taken from another TMDL site.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the pesticide/PCB analyses. The blank was free of any pesticides or PCBs of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All pesticide/PCB results for the samples in this report were considered usable. The completeness for the pesticide/PCB portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ORGANOPHOSPHORUS COMPOUNDS

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on July 18, 2001, and were analyzed for organophosphorus compounds. The organophosphorus compounds, Chloropyrifos, Demeton, Diazinon, Guthion, Malathion and Parathion were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples, and surrogate spikes. A sample from another TMDL site was

selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the organophosphorus compound analyses. The blank was free of any organophosphorus compounds above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All organophosphorus compound results for the sample in this report were considered usable. The completeness for the organophosphorus compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

HERBICIDES

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on July 18, 2001, and were analyzed for herbicides. Herbicides, 2,4,5-T, 2,4,5-TP (Silvex) and 2,4-D, were analyzed using USEPA SW846 Method 8151A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and the surrogate spike. A sample from another TMDL site was

selected as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

The surrogate spike recovery met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

The method blank was run in association with the herbicide analyses. The blank was free of any herbicides above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All herbicide results for the samples in this report were considered usable. The completeness for the herbicide portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

CARBAMATES

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on July 18, 2001, and were analyzed for carbamates. The carbamate compounds, carbaryl and diuron were analyzed using USEPA SW846 Method 8321A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and surrogate spikes. A sample from another TMDL site was selected

as the MS/MSD for this QC batch. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this data group.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria except for the following:

| Analyte | MS %R | Tolerance |
|---------|-------|-----------|
| Diuron | 163 | 25-133 % |

The samples in this data set were not flagged for the non-compliant %Rs since the spiked sample was taken from another TMDL site.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

The MS/MSD RPDs were outside of laboratory specified acceptance criteria as indicated in the following:

| Analyte | MS %R | MSD %R | RPD | Lab Tolerance |
|----------|-------|--------|------|---------------|
| Carbaryl | 41.4 | 63.7 | 42.3 | 25% |
| Diuron | 100 | 163 | 47.9 | |

The samples in this data set were not flagged for the non-compliant %Rs since the spiked sample was taken from another TMDL site.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the carbamate analyses. The blank was free of any carbamates of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All carbamate results for the samples in this report were considered usable. The completeness for the carbamates portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOTAL METALS AND IONS

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on July 18, 2001, and were analyzed for total metals (aluminum, arsenic, barium, cadmium, calcium, chromium, copper, iron, lead, magnesium, mercury, nickel, potassium, selenium, silver, sodium and zinc). The mercury analyses were performed using USEPA SW846 Method 7471A. All other metals were determined using USEPA SW846 Method 6020B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. Another client's sample was used for the MS/MSD for the batch QC for both the total metals and mercury analyses. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria except for the following:

| Analyte | MS %R | MS %R | QC Criteria |
|-----------|-------|--------|-------------|
| Aluminum | -141 | 202 | 80-120% |
| Barium | 79.3 | (99.2) | 80-120% |
| Calcium | -174 | -253 | 80-120% |
| Iron | -167 | -11.8 | 80-120% |
| Magnesium | 67.4 | (114) | 80-120% |
| Mercury | (113) | 127 | 77-120% |
| Potassium | 65.4 | (80) | 80-120% |

() indicates recovery met criteria.

For aluminum, calcium, iron and magnesium, the sample concentration was significantly greater (over 4 times) than the spike concentration. The result for barium and potassium may be biased low in the sample, although no flag was applied since the sample spiked

was taken from another client. The result for mercury may be biased high, although no flag was applied since the sample was taken from another client.

There were no flags added since the sample used for the MS/MSD was from another client sample.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria except for the following:

| Analyte | MS Conc. (mg/Kg-dry wt) | MSD Conc. (mg/Kg-dry wt) | RPD | RPD Limits |
|----------|----------------------------|-----------------------------|-------|------------|
| Arsenic | 101 | 76.25 | 28% | 25% |
| Cadmium | 98.47 | 73.84 | 28.6% | |
| Selenium | 87.48 | 67.33 | 26% | |

The laboratory calculates the %RPD using the concentration results for the metals based on dry weight. The actual recoveries for the metals listed above were all well within acceptance limits, therefore no corrective action were required.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

No calibration, analytical spike or dilution test information was provided for the analyses.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ANIONS (CHLORIDE AND SULFATE)

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on July 18, 2001, and were analyzed for chloride and sulfate using USEPA SW846 Method 9056.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and LCSD samples.

All LCS and LCSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the LCS/LCSD recoveries.

LCS/LCSD RPDs were within laboratory specified acceptance criteria for chloride and sulfate.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEM IN SEDIMENT

General

This sample group consisted of four (4) samples, including one environmental sediment sample, one field duplicate sample and one pair of MS/MSD samples. The samples were collected on July 18, 2001, and were analyzed for Simultaneously Extracted Metals (SEM), including cadmium, copper, lead, mercury, nickel, silver and zinc.

The metals analyses were performed using a modified EPA 1620 method, which is equivalent to EPA 200.7 and EPA 245.5.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. A sample from another TMDL site was analyzed as the MS/MSD sample for this data set. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met QAPP acceptance criteria.

No accuracy criteria for the MS/MSD samples were listed in the QAPP for the SEM analyses. The tolerances listed for metals analyses were used to evaluate the MS/MSD samples.

All MS/MSD %Rs met the QAPP metals acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|---------|-------|--------|-------------|
| Silver | 0 | 0 | 80-120% |
| Cadmium | 72 | (86) | |
| Copper | 0 | 0 | |
| Lead | 0 | 52 | |
| Zinc | 65 | 147 | |

() indicates recovery met criteria

The laboratory explained the observed variances as a product of sample inhomogeneity and matrix interference. This sample was analyzed in duplicate as shown below. As a result of the high variances in both the MS/MSD spike results and the duplicate data, the concentrations for the above compounds were considered estimated although no flags were applies since the sample spiked was taken from a different TMDL site.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria with the exception of the following:

| Analyte | MS Conc (ug/kg) | MSD Conc. (ug/kg) | RPD | QC Limits |
|---------|-----------------|-------------------|-----|-----------|
| Lead | 21.6 | 33.1 | 42% | 20% |

There were no flags applied to the samples since the sample spiked was taken from a different TMDL site.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time specified in the QAPP.

All laboratory blanks were reviewed and found to be free of SEM above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All SEM results for the samples in this report were considered usable. The completeness for the SEM portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

AVS IN SEDIMENT

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for Acid Volatile Sulfide (AVS). The AVS analyses were performed using EPA method 376.3.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS. A sample from another TMDL site was used for the MS/MSD for the batch QC for this group. The results for the MS/MSD will be discussed although not used to qualify the data for the sample in this group.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the QAPP.

All laboratory blanks were reviewed and found to be free of AVS at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All AVS results for the samples in this report were considered usable. The completeness for the AVS portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOC

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for total organic carbon (TOC). The TOC analyses were performed using B&B Laboratories, Inc. Standard Operating Procedure 1005.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the standard reference material (SRM) samples.

TOC met acceptance criteria in both SRM samples analyzed.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

Two method blanks were analyzed in association with the samples. Both blanks were free of TOC at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All TOC results for the samples in this report were considered usable. The completeness for the TOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

GRAIN SIZE

General

This sample group consisted of one (1) environmental sediment sample. The sample was collected on July 18, 2001, and was analyzed for grain size by GS-92-01-B&B Method. Grain size results are reported as a percent of sand, silt or clay based on the weight of the sample.

Accuracy

Accuracy could not be evaluated by this method.

Precision

Precision could not be evaluated by this method.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

There were no method blanks required by this method.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All results for grain size for the sample in this report were considered usable. The completeness for the grain size compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

DATA VERIFICATION REPORT
for sediment samples collected from Segment 1209B
FINFEATHER LAKE TMDL SITE

May 9, 2002

Data Verification by: Sandra de las Fuentes

The following data verification summary report covers environmental sediment samples collected from the Finfeather Lake Segment 1209B, Stations 11798 and 11800, on May 9, 2002.

A Chemist with Parsons has reviewed the data submitted by DHL Analytical, Laboratories, APPL, Inc. and The University of North Texas.

The sample in this event was analyzed for volatiles, semivolatiles, pesticides (including triazines, PCBs, organophosphorus compounds and herbicides), total metals, anions, simultaneously extracted metals (SEM), acid volatile sulfide (AVS), total organic carbon (TOC) and grain size.

There were no field quality control samples collected at this site. No trip blanks were analyzed for volatiles and no field blanks or equipment blanks were collected in association with the sediment samples in this DVR. Therefore, the possibility of contamination during sampling or handling could not be evaluated for these samples.

All samples were collected by Parsons and were analyzed by the various laboratories following procedures outlined in the Assessment of the Presence and Causes of Ambient Toxicity Quality Assurance Project Plan (QAPP).

REVIEW CRITERIA

All data submitted by the various laboratories has been reviewed. Field and laboratory QC sample information was examined, including: laboratory blanks, laboratory control samples (LCS), laboratory duplicates, matrix spikes and matrix spike duplicate (MS and MSD) samples, surrogate spikes and Chain-of-Custody (COC) forms. The findings presented in this report are based on the reviewed information and whether the requirements specified in the project QAPP were met.

VOLATILES

General

This sample group consisted of four (4) samples, including two (2) environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on May 9, 2002 and were analyzed for volatile organic compounds (VOCs). The VOC analyses were performed using USEPA SW846 Method 8260B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results LCS sample and surrogate spikes. Sample 11798-9 was used for the MS/MSD for the batch QC for this group. It should be noted that only a small subset of analytes was reported for the MS/MSD.

The percent recoveries for the LCS were all within acceptance criteria.

The percent recoveries for the MS/MSD were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All volatile results for the samples in this report were considered usable. The completeness for the VOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEMIVOLATILES

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on May 9, 2002, and were analyzed for semivolatile organic compounds (SVOCs). The SVOC analyses were performed using USEPA SW846 Method 8270C.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS samples and the surrogate spikes.

All LCS %Rs were within acceptance criteria.

All of the surrogate recoveries were within laboratory specified acceptance criteria.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was analyzed in association with the samples. The blank was free of target analytes above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All semivolatile results for the samples in this report were considered usable. The completeness for the SVOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TRIAZINES

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on May 9, 2002, and were analyzed for triazine. The triazine compounds, atrazine, cyanazine, metolachlor and simazine, were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample and surrogate spikes.

The LCS percent recoveries were within acceptance criteria, except for the following:

| Analyte | % R | QC Criteria |
|-----------|-----|-------------|
| Cyanazine | 152 | 50-150% |

No flags were applied to the sample concentrations for Cyanazine since the recovery in the LCS was only slightly above the QAPP tolerance.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

No precision data was available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the triazine analyses. The blank was free of any triazines above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All triazine results for the sample in this report were considered usable. The completeness for the triazine portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

PESTICIDES / PCBS

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on May 9, 2002, and were analyzed for pesticides and PCBs. The pesticide/PCB analyses were performed using USEPA SW846 Method 8081A/8082.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample and surrogate spikes.

The LCS percent recoveries were within acceptance criteria except for the following:

| Analyte | %R | Tolerance |
|---------------------|-----|-----------|
| Chlordane | 126 | 61-125 |
| d-BHC | 137 | 55-124 |
| DDE | 129 | 58-122 |
| Dicofol | 244 | 70-130 |
| Dieldrin | 134 | 45-126 |
| Endosulfane | 140 | 60-122 |
| Endosulfane Sulfide | 126 | 57-120 |
| Endrin | 126 | 43-124 |
| g-BHC | 125 | 57-123 |
| Heptachlor Epoxide | 137 | 60-124 |

The samples in this data set were not flagged for the non-compliant %Rs since the spiked sample was taken from another TMDL site.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks, except for the following:

| Sample | Surrogate | %R | QC Criteria |
|--------|-----------------------------|-----|-------------|
| LCS | Tetrachloro-m-xylene (TCmX) | 116 | 50-112% |

No flags were applied to the samples since the surrogate recovery in the LCS was only slightly above QC tolerance. The second surrogate, Decachlorobiphenyl, was within QC tolerance.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the pesticide/PCB analyses. The blank was free of any pesticides or PCBs of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All pesticide/PCB results for the samples in this report were considered usable. The completeness for the pesticide/PCB portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ORGANOPHOSPHORUS COMPOUNDS

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on May 9, 2002, and were analyzed for organophosphorus compounds. The organophosphorus compounds, Chloropyrifos, Demeton, Diazinon, Guthion, Malathion and Parathion were analyzed using USEPA SW846 Method 8141A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample and surrogate spikes.

The LCS percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

There was no precision data available for evaluation.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;

- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the organophosphorus compound analyses. The blank was free of any organophosphorus compounds above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All organophosphorus compound results for the sample in this report were considered usable. The completeness for the organophosphorus compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

HERBICIDES

General

This sample group consisted of four (4) samples, including two (2) environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on May 9, 2002, and were analyzed for herbicides. Herbicides, 2,4,5-T, 2,4,5-TP (Silvex) and 2,4-D, were analyzed using USEPA SW846 Method 8151A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and the surrogate spike. Sample 11800-9 was randomly selected by the laboratory as the MS/MSD for this QC batch.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

The surrogate spike recovery met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

The method blank was run in association with the herbicide analyses. The blank was free of any herbicides above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All herbicide results for the samples in this report were considered usable. The completeness for the herbicide portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

CARBAMATES

General

This sample group consisted of four (4) samples, including two (2) environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on May 9, 2002, and were analyzed for carbamates. The carbamate compounds, carbaryl and diuron were analyzed using USEPA SW846 Method 8321A.

Accuracy

Accuracy was evaluated using the percent recovery (%R) results for the LCS sample, MS/MSD samples and surrogate spikes. Sample 11798-9 was randomly selected by the laboratory as the MS/MSD for this QC batch.

The LCS percent recoveries were within acceptance criteria.

All MS/MSD percent recoveries were within acceptance criteria.

All surrogate spike recoveries met laboratory specified tolerance in the samples, QC and method blanks.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD.

The MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blank was run in association with the carbamate analyses. The blank was free of any carbamates of concern above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All carbamate results for the samples in this report were considered usable. The completeness for the carbamates portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOTAL METALS AND IONS

General

This sample group consisted of four (4) samples, including two (2) environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on May 9, 2002, and were analyzed for total metals (aluminum, arsenic, barium, cadmium, calcium, chromium, copper, iron, lead, magnesium, mercury, nickel, potassium, selenium, silver, sodium and zinc). The mercury analyses were performed using USEPA SW846 Method 7471A. All other metals were determined using USEPA SW846 Method 6020B.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. Sample 11798-9 was used for the MS/MSD for the batch QC for both the total metals and mercury analyses.

All LCS %Rs met acceptance criteria.

All MS and MSD %Rs met acceptance criteria except for the following:

| Analyte | MS %R | MS %R | QC Criteria |
|---------|-------|-------|----------------|
|---------|-------|-------|----------------|

| | | | |
|-----------|--------|-----|---------|
| Aluminum | 213 | 427 | 80-120% |
| Barium | (114) | 149 | 80-120% |
| Calcium | 178 | 430 | 80-120% |
| Copper | (98.1) | 129 | 80-120% |
| Iron | 207 | 407 | 80-120% |
| Magnesium | (112) | 141 | 80-120% |
| Mercury | 125 | 123 | 80-120% |
| Potassium | (106) | 127 | 80-120% |
| Zinc | 78.5 | 122 | 80-120% |

() indicates recovery met criteria.

For aluminum, calcium, and iron, the sample concentration was significantly greater (over 4 times) than the spike concentration therefore no flags were applied to the sample results. No flags were applied to the sample results for barium, copper, magnesium and potassium since the MS was within limits and the MSD was biased high. No flags were applied to the zinc results in the samples since the MS was only slightly below control limits and the MSD was only slightly above control limits.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

No calibration, analytical spike or dilution test information was provided for the analyses.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

ANIONS (CHLORIDE AND SULFATE)

General

This sample group consisted of three (3) samples, including two environmental sediment samples and one laboratory duplicate, randomly selected by the laboratory. The samples were collected on May 9, 2002, and were analyzed for chloride and sulfate using USEPA SW846 Method 9056.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and LCSD samples.

All LCS and LCSD %Rs met acceptance criteria.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the LCS/LCSD recoveries and laboratory duplicate analyte values. Sample 11800-9 was randomly selected by the laboratory as the laboratory duplicate sample.

LCS/LCSD RPDs were within laboratory specified acceptance criteria for chloride and sulfate.

Chloride and sulfate met the QAPP tolerance for the laboratory duplicate samples.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time required by the method.

All laboratory blanks were free of target analytes above the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All metals results for the samples in this report were considered usable. The completeness for the metals portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

SEM IN SEDIMENT

General

This sample group consisted of four (4) samples, including two environmental sediment samples and one pair of MS/MSD samples. The samples were collected on May 9, 2002, and were analyzed for Simultaneously Extracted Metals (SEM), including cadmium, copper, lead, mercury, nickel and zinc.

The metals analyses were performed using a modified EPA 821 draft method, which is equivalent to EPA 200.7 and EPA 245.5.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS and MS/MSD samples. Sample 11798-9 was analyzed as the MS/MSD sample for this data set.

All LCS %Rs met QAPP acceptance criteria.

No accuracy criteria for the MS/MSD samples were listed in the QAPP for the SEM analyses. The tolerances listed for metals analyses were used to evaluate the MS/MSD samples.

All MS/MSD %Rs met the QAPP metals acceptance criteria except for the following:

| Analyte | MS %R | MSD %R | QC Criteria |
|---------|-------|--------|-------------|
| Copper | 14.6 | 22.2 | 80-120% |
| Zinc | 38.6 | 74.4 | |

() indicates recovery met criteria

The laboratory explained the low copper recovery as product of sample inhomogeneity and/or matrix interference. The concentrations for copper were considered estimated and flagged “J” for detected copper results. For zinc, the sample concentration was significantly greater (over 4 times) than the spike concentration, so no corrective action was necessary.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP.

All samples were prepared and analyzed within the hold time specified in the QAPP.

All laboratory blanks were reviewed and found to be free of SEM above the MAL

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All SEM results for the samples in this report were considered usable. The completeness for the SEM portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

AVS IN SEDIMENT

General

This sample group consisted of four (4) samples, including two (2) environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on May 9, 2002, and were analyzed for Acid Volatile Sulfide (AVS). The AVS analyses were performed using EPA method 821 Draft.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS. Sample 11798-9 was used for the MS/MSD for the batch QC for this group.

All LCS %Rs met acceptance criteria.

The results for the MS and MSD %Rs are as follows:

| Analyte | MS %R | MSD %R | QC Criteria |
|----------------|--------------|---------------|--------------------|
| AVS | 0 | 58.6 | 80-120% |

For AVS, the sample concentration was significantly greater (over 4 times) than the spike concentration, so no corrective action was necessary.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

All MS/MSD RPDs were within laboratory specified acceptance criteria.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and
- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the procedures outlined in the QAPP with the exceptions noted above.

All samples were prepared and analyzed within the hold time required by the QAPP.

All laboratory blanks were reviewed and found to be free of AVS at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All AVS results for the samples in this report were considered usable. The completeness for the AVS portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

TOC

General

This sample group consisted of four (4) samples, including two environmental sediment samples and one pair of MS/MSD samples, randomly selected by the laboratory. The samples were collected on May 9, 2002, and were analyzed for total organic carbon (TOC). The TOC analyses were performed using EPA 415.1.

Accuracy

Accuracy was evaluated using the percent recovery (%R) for the LCS sample and the MS/MSD samples. Sample 11800-9 was randomly selected by the laboratory and analyzed as the MS/MSD for this data set.

TOC met acceptance criteria in the LCS sample analyzed.

TOC met acceptance criteria in the MS/MSD samples.

Precision

Precision was evaluated using the Relative Percent Difference (RPD) obtained from the MS/MSD recoveries.

MS/MSD RPDs were within laboratory specified acceptance criteria for TOC.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

- Examining laboratory blanks for contamination of samples during analysis.

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

One method blanks were analyzed in association with the samples. The blank was free of TOC at the MAL.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All TOC results for the samples in this report were considered usable. The completeness for the TOC portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

GRAIN SIZE

General

This sample group consisted of two (2) environmental sediment samples. The samples were collected on May 9, 2002, and were analyzed for grain size by EPA 3.4, 3.5 (600/2-78-054). Grain size results are reported as a percent of gravel, sand, silt or clay based on the weight of the sample.

Accuracy

Accuracy could not be evaluated by this method.

Precision

Precision could not be evaluated by this method.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represents actual site conditions. Representativeness has been evaluated by:

- Comparing actual analytical procedures to those described in the QAPP;
- Evaluating holding times; and

All samples were prepared and analyzed following the QAPP and within the hold time required by the method.

There were no method blanks required by this method.

Completeness

Completeness was evaluated by comparing the total number of samples collected with the total number of samples with valid analytical data.

All results for grain size for the sample in this report were considered usable. The completeness for the grain size compound portion of this data set is 100%, which meets the minimum QAPP acceptance criteria of 90%.

**APPENDIX G
TECHNICAL MEMOS**

TECHNICAL MEMORANDUM 1

February 13, 2002

Suggested Criteria For Assessing Ambient Sediment And Water Toxicity Testing Results

INTRODUCTION

This technical memorandum recommends criteria for assessing ambient sediment and water chronic toxicity testing results. It is recommended that the lethal and sublethal end-point criteria described in this memorandum be used to identify waterbodies with varying degrees of impairment of aquatic life uses. Ambient toxicity tests exceeding the recommended criteria indicate the waterbody needs additional assessment and/or should be listed on the 303(d) and 305(b) List.

The following criteria recommendations and supporting information are divided into criteria for assessing sediment and ambient water toxicity data.

SEDIMENT RECOMMENDATIONS

Sediment Criteria 1 – Use an alpha = 0.05 when the number of replicates is less than 20. Use an alpha = 0.01 when the number of replicates is 20 or more.

To maintain a high power, 20 or more replicates should be used before using an alpha = 0.01. Otherwise, use an alpha = 0.05.

Sediment Criteria 2 – The whole-sediment toxicity test is recommended for use with ambient sediment samples. Use elutriate tests only on dredge material or when testing the effects of an activity that will cause excessive resuspension of the instream sediment.

Whole sediment toxicity testing is the preferred method because of its consistency and better approximation of actual instream conditions than elutriate testing. For gathering sediment data for aquatic life use attainment determinations, comparing whole sediment test to whole sediment test are preferred. Comparing a combination of whole sediment tests to elutriate tests is like comparing apples to oranges. Both tests are good for their intended purpose; however, for consistency, whole sediment tests are recommended rather than instream sediment testing. Use elutriate tests only on dredge material or when testing the effects of an activity that will cause excessive resuspension of the sediment.

Sediment Criteria 3 – In general, sublethal effects testing is not appropriate to short-duration sediment toxicity tests. Sublethal effects sediment toxicity test methods have not been fully developed. Long-term sublethal effects testing is new and more data are needed to assess this method. Therefore, sublethal effects testing will not be used to assess attainment of aquatic life uses at this time.

More data are needed before sublethal whole sediment toxicity tests can be considered appropriate for assessing aquatic life use attainment for instream sediment. According to EPA's freshwater sediment toxicity testing manual, "*Additional studies are ongoing to more thoroughly evaluate the relative sensitivity between lethal and sublethal endpoints measured in 10-d tests and between sublethal endpoints measured in the long-term tests (28-d). Results of these studies and additional applications of the methods described in Section 14 and 15 will provide data that can be used to assist in determining where application of long-term tests will be most appropriate.*"(1)

Sediment Criteria 4 - Mortality in the sample must also be less than the minimum control mortality allowed according to the EPA method.

For ambient sediment toxicity testing, if the conditions of test acceptability are met and survival of the test organism is equal to or greater than 80 percent of the original number of test organisms, the test shall be considered to not have demonstrated significant lethality.

The first WET test "Statistical Interpretation" provision in recent TPDES permits states, "*If the conditions of test acceptability are met and the survival of the test organism is equal to or greater than 80% in the critical dilution and all dilutions below that, the test shall be considered to not have demonstrated significant lethality.*" It is recommended that similar criteria be applied to sediment toxicity testing.

Sediment Criteria 5 – The minimum significant difference (MSD) or the minimum detectable difference (MDD) should not less than 20 percent.

In general, protocols applicable to sediment toxicity are not as well established as those for water methods. However, a 1992 EPA Region 6/ Galveston Corps of Engineers Regional Implementation Agreement for the Ocean Disposal of Dredged Material Off the Texas Coast states:

"Dredged material does not meet the LPC for benthic toxicity when bioassay organism mortality (1) is statistically greater than in the reference sediment, and (2) exceeds mortality in the reference sediment by at least 10% or exceeds the reference mortality by 20% when amphipods are used."

These approaches document ample justification for the selection of a minimum significant difference in survival of the test organism relative to the control.

A.1 WATER RECOMMENDATIONS

The following criteria are recommended:

Water Criteria 1 - Use the Fisher's Exact statistical test and the t-Test for ambient water toxicity testing for survival and sublethal effects, respectively.

Use of the Fisher's Exact statistical test and the t-Test for ambient water toxicity testing for survival and sublethal effects, respectively, is recommended. The EPA Region 6

Laboratory uses the Fisher's Exact and t-Test for determining the MSD for chronic survival and sublethal effects in ambient water toxicity testing. Although EPA's chronic whole effluent toxicity (WET) test manual allows for different statistical tests and reasonable arguments can be made for using different tests, the same statistical tests should be used to allow for a more direct comparison of results from one lab to another.

Water Criteria 2 - For ambient water survival and sublethal toxicity testing, if the conditions of test acceptability are met and survival of the test organism is equal to or greater than 80 percent of the number of test organisms at the beginning of the test, the test should be considered to not have demonstrated significant lethality.

For ambient water toxicity testing, if the conditions of test acceptability are met and survival of the test organism is equal to or greater than 80 percent of the original number of test organisms, it is recommended that the test be considered to not have demonstrated significant lethality.

The first WET test "Statistical Interpretation" provision in recent TPDES permits states, *"If the conditions of test acceptability are met and the survival of the test organism is equal to or greater than 80% in the critical dilution and all dilutions below that, the test shall be considered to not have demonstrated significant lethality."* It is recommended that similar criteria be applied to ambient water toxicity testing.

Water Criteria 3 - Use an alpha = 0.05 for determining the minimum significant difference in lethal toxicity testing and an alpha = 0.01 in sublethal toxicity testing. Sublethal toxicity test failure rates of less than 30 percent, by themselves, provide inconclusive data. The waterbody should continue to be judged as fully supporting aquatic life uses if previously designated as such. Sublethal toxicity test failure rates greater than 31 percent but less than 50 percent, by themselves, provide inconclusive evidence that the stream is not supporting aquatic life uses. Nevertheless, tests failures in the above range do indicate the stream is partially supporting the use, but additional testing is warranted. Sublethal toxicity test failure rates greater than 50 percent, by themselves, provide evidence that toxicity probably exists and the stream should be designated as not supporting aquatic life uses and that additional testing and potential toxicant identification are warranted.

The current debate between U.S. Environmental Protection Agency (EPA) and the regulated community over the interlaboratory variability of WET testing and the correlation of WET test failures with instream impairment, has spurred much interest and research. In 1995 EPA amended 40 CFR Part 136 – "Guidelines Establishing Test Procedures for the Analysis of Pollutants" to include WET testing. In 1996 the City of San Bernardino, United Water Florida, and City of Washington, Georgia sued EPA over these methods. Several items identified by the plaintiffs were clarification of the WET method procedures, guidance for use of WET test in permits, and guidance addressing when and under what circumstances a TIE/TRE should be initiated. Lone Star Steel Company also sued EPA in 1996 concerning issues related to WET test failures due to pathogens. In 1997 EPA amended and added new WET method procedures. Shortly after issuing the final WET rule, EPA was sued by the Edison Electric Institute, *et al.*,

and Western Coalition of Arid States(2). These plaintiffs claimed, among other things, that the variability of the WET tests exacerbated results because of unaccounted Type I errors. A Type I error occurs when an effluent is shown to be toxic when it is, in fact, not toxic, or when an ambient toxicity test indicates impairment of aquatic life uses when, in fact, the stream is fully supportive of aquatic life uses. All these suits were settled out of court in 1998 contingent upon separate agreements(2).

EPA's Wet Variability Study

The settlement agreements required EPA to amend most of the WET test methods and issue clarifications and new guidance. Additionally, EPA was required to perform an interlaboratory WET variability study subject to independent peer review. The final Interlaboratory WET Variability Study was published in September 2001(5). Revised WET methods were proposed in October 2001 with the comment period ending January 11, 2002.

Following the 1998 settlements through proposal of the latest revisions of the WET methods, a number of reports and professional articles were published. A study published in 2000 entitled "Investigating the Incidence of Type I Errors for Chronic Whole Effluent Toxicity Testing Using *Ceriodaphnia Dubia*"(3) sought to determine the frequency of Type I errors in *C. dubia* survival and reproductive toxicity tests. Non-toxic synthetic fresh water created using EPA's recommendations(4) was sent by participating wastewater treatment plant operators to 16 laboratories. The laboratories were not aware that the samples were non-toxic. The paper's abstract contained the following conclusion:

"Of the 16 tests completed by the biomonitoring laboratories, two did not meet control performance criteria. Six of the remaining 14 valid tests (43%) indicated toxicity ($TUc > 1$) in the sample (i.e., no-observed-effect concentration or $IC25 < 100%$ (Interpreted to mean $NOEC < 100%$ and $IC25 < 100%$)). This incidence of false positives was six times higher than expected when the critical value (alpha) was set to 0.05. No plausible causes for this discrepancy were found. Various alternatives for reducing the rate of Type I errors are recommended, including greater reliance on survival endpoints and use of additional test acceptance criteria."

The survival end-points between the control and the test for the 16 labs were not significantly different. All the false-positives mentioned above were observed in the *C. dubia* reproduction tests.

Results of this study, in part, caused EPA to propose changes(6) to the method of calculating the MSD between the control and the test for both sublethal endpoints for *C. dubia* and the fathead minnow toxicity tests. EPA is proposing to allow NPDES permit holders to reduce the nominal (Type I) error rate "alpha" from 0.05 to 0.01 when results of the test are reported as a condition of the permit or when WET permit limits are

derived without allowing for receiving water dilution. EPA set an additional condition, in the revised chronic WET manual, of not exceeding the Maximum-Minimum Significant Difference (Mx-MSD) using an alpha = 0.01. The Mx-MSD for *C. dubia* reproduction and fathead growth tests is 37 percent and 35 percent, respectively. In other words, the maximum MSD for *C. dubia* reproduction test cannot exceed 37 percent of the mean young per female in the control when using an alpha = 0.01. Insufficient replicates can cause the calculated MSD to exceed the Mx-MSD.

EPA made the decision to allow permittees to change the alpha to 0.01, not because the WET test was theoretically flawed, but because, in practice, WET test results were being used to make “yes or no” regulatory decisions. The NPDES permit holders did not want to be falsely accused by EPA of harming the environment. The same can be argued when a stream segment is listed as partially or not supporting aquatic life uses in the 305(b) Report based solely on ambient-water sublethal toxicity testing results. Stream segments listed in the 305(b) report as not supporting aquatic life uses are placed on the state’s 303(d) List.

In October 2000, EPA published preliminary results of their Interlaboratory WET Variability Study required in the above mentioned out-of-court settlement. In February 2001, the Western Coalition of Arid States (West-CAS), one of the plaintiffs in the out-of-court settlement, provided EPA its comments to the preliminary variability study(7). One comment provided by West-CAS relative to this memorandum is:

“EPA underestimated the true rate of false positives by misinterpreting results from the reference toxicant tests. The Agency acknowledged that many laboratories failed to observe toxicity in the chronic Ceriodaphnia tests on reference toxicant samples. The agency asserts, incorrectly, that the failure was due to “differences in test sensitivity between laboratories.” In fact, 9 of the 11 most sensitive tests (based on percent minimum significant difference) indicated that the reference toxicant sample was not toxic. Conversely, 9 of the 11 least sensitive tests showed the sample was toxic. On average, tests that indicated toxicity(,) were 50% less sensitive than tests that indicated no toxicity. The difference in test sensitivity was statistically-significant (p=.05). If the reference toxicant sample was actually toxic, then the most sensitive tests would be the most likely to confirm the presence of toxicity. Because that did not occur in EPA’s study, and because two-thirds of the laboratories (including the referee lab) reported no statistically-significant difference in Ceriodaphnia reproduction, the only logical conclusion is that the sample was not toxic. Therefore, the laboratories observing test failures were, in fact, reporting false positives. Based on data from the nontoxic reference toxicant tests, the true rate of Type-I error exceeds 33% for the chronic Ceriodaphnia reproduction method.”

Risk Science and West-CAS provided additional comments after the final version of the variability study was published in September 2001. The following is a comment that expands on the one provided above(8).

“Two-thirds of the laboratories failed to observe a toxic response for the reference toxicant samples during the chronic Ceriodaphnia dubia tests. Given that the most sensitive c. dubia tests indicated no toxicity and the least sensitive c. dubia tests showed toxicity, how should the true nature of the original sample be classified: toxic or non-toxic?”

In March 2001, EPA published peer review comments to the variability study. The following are some of the more interesting comments from the three reviewers, X, Y and Z, on EPA’s WET Variability Study, 2001(9).

Peer Reviewer X:

Question: *Are the results scientifically acceptable within the context of the intended regulatory use?*

Answer: “Yes and No. The data are there, though they need clarifications as noted in this review. However, I am not convinced that the Study Plan allowed for direct comparisons with regulatory use. For example, test concentrations were regimented and had larger than normal gradations, and false positives were not evaluated in terms of ecological significance but rather in terms of testing only. These tests are applied, to often, as decisive when (see Section 5 of this review, below) they are far from such.”

Comment: “First, single species toxicity tests (e.g., WET tests) are valuable first tier assessments. Results should then be used as guidance for additional studies such as exposure characterizations to provide insight on causality (e.g., TIEs), or biological assessments to provide data for detecting ecological impairment. As noted by Hall and Gidding (2000) and Chapman (2000), WET tests are the beginning, not the end of evaluations.”

Peer Reviewer Z

Question: Are the results scientifically acceptable within the context of the intended regulatory use?

Answer: “YES/NO. The results are scientifically acceptable within any context since the approach was scientifically rigorous. However, there is a distinction between scientifically acceptable in terms of accepting the results versus whether or not the results are acceptable for regulatory use. This is reminiscent of the following story: “*The operation was a success, but the patient died!*” The results should be accepted, but the results seem to show that some of these tests should not be used in the regulatory context because the successful completion rate is too low and the CV values are too high.”

Additional comment by West-CAS and the peer review committee and EPA’s response to their comments may be viewed at <http://www.toxicity.com/>

Reducing Type I Errors

Many scientific articles have been published that state or infer that WET or ambient toxicity tests in and by themselves do not necessarily indicate aquatic life uses are impaired (10, 11, 12). For *C. dubia* reproductive tests, Type I errors appear to occur, in practice, in greater than 5 percent ($\alpha = 0.05$) of the tests. Reasons include sampling and laboratory contamination, improper food preparation or contamination, individually poor performing females, not discarding results following a procedural error, parasites, pH drift, poor training, inexperience, and others (6, 11, 13). Not discarding results following a procedural error is more common than expected (7, 8). As an example, in EPA's final WET variability study, the successful *C. dubia* reproductive test completion rate for labs that met the Test Acceptance Criteria was 82 percent. Nevertheless, the successful completion rate for labs that met all non-discretionary conditions in 40 CFR Part 136 was 40 percent (7). There is also much debate as to whether WET testing correlates with instream aquatic conditions. In Section 3.5.5 of the Water Environment Research Foundation report (10) it was stated that "*Ceriodaphnia chronic reproduction NOEC showed no relationship with instream biological conditions.*" This report and specifically this statement focused on comparing results of WET testing of permitted point-source discharges to instream biological (benthic macroinvertebrate) assessments. Although this report compares WET test results from discharged effluent and not ambient water, the above quote was based, in part, on results from effluent dominated streams.

The following quote summarizes the views of many scientist and toxicologist.

"Rather than relying on a discrete, yes/no decision based on hypothesis testing of ambient toxicity tests at (α) levels of 0.1, 0.05 or 0.01, statistical interpretation of toxicity data and scientific judgement should be incorporated into the decision making process of determining when a stream segment or waterbody is impaired and considered for TMDL development." (14) Nevertheless, yes or no regulatory decisions are made on scientific evidence that may not support the regulatory action taken.

CONCLUSION

The recommended Sediment Criteria mirror previously established criteria established by the U.S. Corps of Engineers or are similar to the recommended water criteria. Water Criteria 1 and 2 are minor modifications to existing TNRCC policy. The reasons for these recommendations are noted above. Water Criteria 3 is more likely to be controversial. Unfortunately, there must be a line drawn where yes or no regulatory decisions concerning toxicity testing and attainment of aquatic life uses are made. Water Criteria 3 through 6 provide this line.

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Appendix G

Technical Memos

This document provides a list of sediment quality indices which have been compiled for screening purposes. A brief discussion of the indices generally available and the methodology used to complete the table follows.

Measured concentrations of contaminants may be compared to sediment quality screening indices to indicate whether a measured concentrations of a compound may have the potential to cause toxicity. There are many ways to derive sediment quality indices. Therefore, a discussion of the ways in which indices are derived is necessary to understand the various types of indices and how they differ.

The bulk concentration of contaminants in sediment is measured. Typically most of the bulk measured contaminant is bound in organic matter (in the case of organic compounds) and acid-volatile sulfides (in the case of metals), and not biologically available to cause toxicity in sediment. In general, organic matter has a much higher capacity for binding organic contaminants than inorganic matter. The composition of the sediments governs the bioavailability and expressed toxicity of a contaminant.

Organisms differ greatly in their sensitivity to contaminants. Toxic effects may include, but are not limited to changes in growth rates, number of offspring, behavior, physiology, and mortality. Thus, a broad range of concentrations is reported to cause toxicity. For example, DDT has been observed to cause small reductions in growth of oysters at concentrations of 0.01 µg/L in water, while fireworms (*Eurythroë complanata*) will live at 1,000 µg/L of DDT. For many contaminants, toxic effects have only been measured with a few types of organisms. Water and sediment quality indices are designed to protect all organisms from any biological effects, therefore, they are typically set well below the level that has been observed to be toxic in order to include a substantial margin of safety. Thus, contaminant levels in sediments that exceed screening indices do not necessarily indicate the presence of biological effects to the indigenous species present.

Equilibrium-Partitioning Sediment Quality Indices for Organic Compounds

Sediment quality indices based on “equilibrium partitioning” are provided in this summary. This term refers to the division, at equilibrium, of organic contaminants between sediment organic matter and the pore water present between the grains of sediments. The sediment pore water fraction is assumed to be mostly bioavailable. This approach has been used in numerous studies. The USEPA (1993) recommends it as one component of the sediment quality triad. It allows consideration of site-specific bioavailability of contaminants.

Four different equilibrium partitioning-based screening indices for the organic compounds measured in this study are listed in Table 1. While equilibrium partitioning-based indices must be calculated for each location using the site-specific organic carbon concentration, these indices

are illustrated using a sediment organic carbon content of 1 percent. The illustrative value of 1 percent is typically used for general publications, since it can be easily multiplied to address site-specific organic carbon. The indices would be twice as high for a sediment with 2 percent organic carbon, three times as high for a sediment with 3 percent organic carbon, and so forth.

There is a broad range in values for those contaminants for which multiple equilibrium partitioning-based indices can be calculated. This is caused by differing assumptions used in the calculations, as well as considerable uncertainties in the data sources. In Table 1, the indices are labeled as Tier 1, Tier 2, predicted, and acute. Tier 1 sediment quality indices are available for only a few contaminants. Tier 1 indices are based on an aquatic chronic toxicity data set and were verified by EPA using whole sediment toxicity tests. The toxicity is calculated as a draft EPA final chronic value, which is based on the chronic toxicity to the most sensitive species and incorporates a substantial margin of safety. Tier 2 sediment quality indices are similar to draft Tier 1 indices, but were based on draft EPA secondary chronic values, which are based on less extensive toxicity data sets. Because there is more uncertainty regarding toxicity, EPA lowered Tier 2 indices by a factor ranging from 4 to 22 to be more protective. For some measured contaminants, Tier 1 or Tier 2 indices were not available. Therefore, “Predicted” sediment quality indices were calculated in the same way that EPA developed Tier 1 and Tier 2 indices. In some cases, these “Predicted” indices were based on expected (rather than measured) partitioning behavior, and/or very limited chronic toxicity datasets. Primary data sources used for this data set was obtained from a broad range of sources, such as EPA Region 4, EPA Office of Solid Waste and Emergency Response, and others. Thus, there is substantial uncertainty in “Predicted” sediment quality indices. Finally, no chronic toxicity information was available for several compounds. Thus, “Acute” sediment quality indices were calculated based on observed acute lethal toxicity to the most sensitive aquatic organisms. Marine acute toxicity measurements were used if available. As expected, calculations based on acute toxicity are higher than those based on chronic toxicity.

Other Sediment Quality Indices for Organic Compounds

In the absence of information about the bioavailability of contaminants, several different types of other sediment quality screening indices have been developed. To determine whether there is cause for further investigation of sediment contaminants, the State of Texas Surface Water Quality Monitoring Program applies the simplest approach. They compare individual sediment contaminant measurements at a particular location to the 85th percentile of all concentrations of that contaminant measured in all Texas tidal streams and estuaries. This technique focuses more on sediment quality relative to other locations than the toxicity and bioavailability of a particular compound.

Another slightly more refined approach than the one described above is based on empirical relationships between bulk sediment contaminant concentrations and observed biological effects. Indices based on this approach also do not consider site-specific conditions affecting contaminant bioavailability. They are applied without knowledge of the organic carbon content of the sediment. Several government agencies have used this method to develop sediment quality indices to screen sediments for potential biological effects. No single set of such indices has been accepted by all scientific and regulatory communities. The National Oceanic and Atmospheric

Administration developed the Effects Range-Median (ER-M) and Effects Range-Low (ER-L) indices (Long and Morgan, 1991; Long et al., 1995). The ER-M is the median of the range of contaminant concentrations at which adverse biological effects were observed, while the ER-L is the tenth percentile. A second set of indices, the Probable Effects Levels (PELs) and Threshold Effects Levels (TELs), were developed for the Florida Department of Environmental Protection (MacDonald, 1994). The PEL is defined as the average of: 1) the median of the range of contaminant concentrations at which biological effects were observed; and 2) the eighty-fifth percentile of the range of concentrations at which biological effects were not observed. Thus, the PEL is similar to, but slightly lower than the ER-M. The TEL is the average of: 1) the fifteenth percentile of concentrations having biological effects; and 2) the fiftieth percentile of concentrations having no effects. The Apparent Effects Threshold (AET), developed for the State of Washington, is the highest sediment chemical concentration at which statistically significant differences in observed adverse biological effects from reference conditions do not occur. This is equivalent to the concentration above which adverse biological effects typically always occur for a given site. AETs also vary with the biological indicator examined. The AET-low is the lowest AET among multiple biological indicators (e.g., growth and reproduction effects), while the AET-high is the highest AET measured, typically mortality.

Summary of Sediment Quality Indices for Organic Compounds

Various sediment quality indices are available and each of the indices was developed with a given set of assumptions. As discussed, four types of equilibrium partitioning-based indices are presented in Table 1. These types of indices are based upon USEPA protocols. This information is provided for reference. Specific data analysis methodologies that will be applied to the sediment data for organic compounds will be based upon analysis of all of the site-specific data collected, including indigenous benthic organisms.

Sediment Quality Screening for Metals

The metals lead, cadmium, nickel, silver, zinc, and copper, form strong and biologically unavailable compounds with sulfides in sediments. Numerous studies have shown that when molar concentrations of these metals in sediments do not exceed the molar concentration of acid volatile sulfide (AVS), metal toxicity is seldom observed (Pesch et al, 1995; Casas and Crecilius, 1994; DiToro et al, 1990; Hansen et al, 1996; Berry et al, 1996). AVS is the solid-phase sulfide in sediments that is soluble in cold acid (typically 1 N hydrochloric acid). Organic matter and sediment particle surfaces may provide secondary sorbent phases to reduce the bioavailability and toxicity of metals in sediments.

The equilibrium partitioning approach will be applied to predict the toxicity of divalent metals by the method recommended by the USEPA (1994). Briefly, the sum of molar concentrations of mercury, silver, copper, lead, cadmium, zinc, and nickel extracted with the AVS (simultaneously extracted metals, or SEM) is compared to the AVS concentration. If the SEM is less than AVS, it will be assumed that the metals are bound and not causing toxicity. If SEM exceeds AVS, but the available metal concentrations do not exceed their chronic toxic values, then toxicity is again considered unlikely. Finally, metal partitioning to sediment organic matter and sediment surfaces

will be evaluated with partition coefficients, as with organic compounds. If the following three criteria are met, potential metal toxicity is indicated (Ankley et al, 1996).

1. $\sum_i [SEM_i] \geq [AVS]$
2. $\sum_i \frac{[SEM_i] - [AVS]}{K_{d,oc,i} * f_{oc} * [FCV_i]} \geq 1$
3. $\sum_i \frac{[SEM_i]}{K_{d,min,i} [FCV_{i,d}]} \geq 1$

where $[SEM_i]$ is the concentration of simultaneously extractable metal i , $[AVS]$ is the concentration of acid volatile sulfide, $K_{d,oc}$ is the metal distribution coefficient between sediment organic carbon and pore water, f_{oc} is the organic carbon content of the sediment, $K_{d,min}$ is a minimum metal distribution coefficient between sediment surfaces and pore water, and $[FCV]$ is the final chronic value for toxicity of each metal.

Other Sediment Quality Indices for Metals

In the absence of the site-specific data described above, several different types of other sediment quality indices have been developed. The approaches described for other sediment quality indices of organic compounds have also been applied to metals. These approaches are the same and will not be repeated here.

Summary of Sediment Quality Indices for Metals

Various sediment quality indices are available and each of the indices was developed with a given set of assumptions. As discussed, equilibrium partitioning-based indices for metals are based upon specific sets of site-specific data. In the study, total metals, AVS, SEM and organic carbon data were collected for the sediments. Specific data analysis methodologies that will be applied to the sediment data for metals will be based upon analysis of all of the site-specific data collected, including indigenous benthic organisms.

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Table 1. Equilibrium Partitioning-Based Sediment Quality Screening Indices at 1% Organic Carbon, in µg/kg Sediment

| Organic Compound | Tier 1 | Tier 2 | Predicted | Acute |
|------------------------------------|---------------|---------------|------------------|--------------|
| 1,1,1-Trichloroethane | | 170 | 30 | 26,441 |
| 1,1,2,2-Tetrachloroethane | | 940 | 1,366 | 12,089 |
| 1,1,2-Trichloroethane | | | 1,257 | 10,157 |
| 1,1-Dichloroethane | | | 27 | 2,417 |
| 1,1-Dichloroethene | | | 31 | 7,259 |
| 1,2-Dichlorobenzene | | 340 | 328 | |
| 1,2-Dichloroethane | | | 256 | 1,184 |
| 1,2-Dichloropropane | | | 2,075 | |
| 1,3-Dichlorobenzene | | 1,700 | 1,664 | |
| 1,4-Dichlorobenzene | | 350 | 344 | |
| 2,4-Dinitrotoluene | | | 293 | |
| 2,6-Dinitrotoluene | | | | 10,341 |
| 2-Chloroethyl Vinyl Ether | | | | 9,727 |
| 2-Chloronaphthalene | | | | 267,345 |
| 2-Methylnaphthalene | | | 157 | |
| 3,3'-Dichlorobenzidine | | | | 20,603 |
| 4,4'-DDD | | | 110 | |
| 4,4'-DDE | | | 6,187 | |
| 4,4'-DDT | | | 26 | 11,047,126 |
| 4-Bromophenyl phenyl ether | | 1,300 | 1,248 | |
| 4-Chlorophenyl phenyl ether | | | | 456,209 |
| Acenaphthene | 2,320 | | 1,718 | 395,891 |
| Acenaphthylene | | | | 30,620 |
| Acrolein | | | 0.005 | |
| Acrylonitrile | | | 1.330 | 46 |
| Alpha-Chlordane | | | 65 | 421,670,625 |
| Anthracene | | | 215 | 7,968 |
| Azobenzene (1,2-diphenylhydrazine) | | | 21 | |
| Benzene | | 57 | 160 | 147,632 |
| Benzidine | | | 1.66 | 24 |
| Benzo(a)anthracene | | | 107 | 10,350,786 |
| Benzo(a)pyrene | | | 143 | 30,698,790 |
| Benzo(b)fluoranthene | | | | 27,372 |
| Benzo(g,h,i)perylene | | | | 7,716 |
| Benzo(k)fluoranthene | | | | 17,418 |
| bis(2-Chloroethoxy)methane | | | | |
| bis(2-Chloroethyl)ether | | | 368 | |
| bis(2-Chloroisopropyl)ether | | | | |
| bis(2-Ethylhexyl)phthalate | | | 885363 | |
| Bromodichloromethane | | | 7426 | |
| Bromoform | | 650 | 1307 | |
| Bromomethane | | | 18 | |
| Butyl benzyl phthalate | | 11000 | 10933 | |

| Organic Compound | Tier 1 | Tier 2 | Predicted | Acute |
|---------------------------|---------------|---------------|------------------|--------------|
| Carbon tetrachloride | | 1200 | 225 | 45,470 |
| Chlorobenzene | | 820 | 413 | 50,361 |
| Chloroethane | | | | 7,937 |
| Chloroform | | | 22 | 745 |
| Chloromethane | | | 432 | |
| Chrysene | | | | 2,809 |
| cis-1,3-Dichloropropene | | | 0.05 | 205 |
| Dibenzo(a,h)anthracene | | | | 15,087 |
| Dibromochloromethane | | | 8701 | |
| Diethyl phthalate | | 630 | 606 | |
| Di-n-butyl phthalate | | 11000 | 11860 | 81,322,597 |
| Di-n-octylphthalate | | | 885363 | |
| Dioxins/furans TEQ | | | 0.26 | |
| Ethylbenzene | | 4800 | 90 | 66,435 |
| Fluoranthene | 2960 | | 6601 | 17,144,309 |
| Fluorene | | 540 | 538 | |
| Gamma-Chlordane | | | 65 | 291,925,818 |
| Heptachlor Epoxide | | | 2.96 | |
| Hexachlorobenzene | | | 13570 | |
| Hexachlorobutadiene | | | 171 | |
| Hexachloroethane | | 1000 | 1021 | |
| Mean Avg. Aroclor PCB | | | 97 | 80,898,414 |
| Mean Avg. Toxaphene | | 100 | 28 | |
| Methylene Chloride | | | 374 | 1,223 |
| Naphthalene | | 470 | 239 | 239,431 |
| Phenanthrene | 2380 | | 1859 | 17,412,134 |
| Pyrene | | | | 939 |
| Trans-1,3-Dichloropropene | | | 230 | |
| Trichloroethene | | 1600 | 215 | |
| Vinyl Chloride | | | | 691 |

Table 2. Non-Equilibrium Partitioning-Based Sediment Quality Screening Indices, in µg/kg sediment.

| Contaminant | ER-L | ER-M | AET-L | AET-H | TEL | PEL |
|----------------------------|-------------|-------------|--------------|--------------|------------|------------|
| 1,2-Dichlorobenzene | - | - | 50 | 50 | - | - |
| 1,4-Dichlorobenzene | - | - | 110 | 120 | - | - |
| 2-Methylnaphthalene | 70 | 670 | 670 | 1900 | 20.2 | 201 |
| 4,4'-DDD | 2 | 20 | 16 | 43 | 1.22 | 7.81 |
| 4,4'-DDE | 2.2 | 27 | 9 | 15 | 2.07 | 374.17 |
| 4,4'-DDT | 1 | 7 | 34 | 34 | 1.19 | 4.77 |
| Acenaphthene | 16 | 500 | 500 | 2000 | 6.71 | 88.9 |
| Acenaphthylene | 44 | 640 | 1300 | 1300 | 5.87 | 127.87 |
| Alpha-Chlordane | 0.5 | 6 | - | - | 2.26 | 4.79 |
| Anthracene | 85.3 | 1100 | 960 | 13000 | 46.85 | 245 |
| Arsenic | 8200 | 70000 | 57000 | 700000 | 7240 | 41600 |
| Benzo(a)anthracene | 261 | 1600 | 1600 | 5100 | 74.8 | 693 |
| Benzo(a)pyrene | 430 | 1600 | 1600 | 3600 | 88.8 | 763 |
| Benzo(b)fluoranthene | - | - | 3600 | 9900 | - | - |
| Benzo(g,h,i)perylene | - | - | 720 | 2600 | - | - |
| Benzo(k)fluoranthene | - | - | 3600 | 9900 | - | - |
| Bis(2-ethylhexyl)phthalate | 182 | - | 1300 | 1900 | 182 | 2650 |
| Butyl benzyl phthalate | - | - | 900 | 900 | - | - |
| Cadmium | 1200 | 9600 | 5100 | 9600 | 676 | 4210 |
| Chromium | 81000 | 370000 | 260000 | 270000 | 52300 | 160000 |
| Chrysene | 384 | 2800 | 2800 | 9200 | 108 | 846 |
| Copper | 34000 | 270000 | 390000 | 1300000 | 18700 | 108000 |
| Dibenzo(a,h)anthracene | 63.4 | 260 | 230 | 970 | 6.22 | 135 |
| Diethyl phthalate | - | - | 200 | 200 | - | - |
| Ethylbenzene | - | - | 10 | 37 | - | - |
| Fluoranthene | 600 | 5100 | 2500 | 30000 | 113 | 1494 |
| Fluorene | 19 | 540 | 540 | 3600 | 21.2 | 144 |
| Gamma-Chlordane | 0.5 | 6 | - | - | 2.26 | 4.79 |
| Heptachlor Epoxide | - | - | - | - | 0.6 | 2.67 |
| Hexachlorobenzene | - | - | 22 | 230 | - | - |
| Hexachlorobutadiene | - | - | 11 | 270 | - | - |
| Lead | 46700 | 218000 | 450000 | 660000 | 30240 | 112180 |
| Mean Avg. Aroclor PCB | 22.7 | 180 | 1000 | 3100 | 21.6 | 188.79 |
| Mercury | 150 | 710 | 590 | 2100 | 130 | 700 |
| Naphthalene | 160 | 2100 | 2100 | 2700 | 34.6 | 391 |
| Nickel | 20900 | 51600 | 110000 | - | 15900 | 42800 |
| Phenanthrene | 240 | 1500 | 1500 | 6900 | 86.7 | 544 |
| Pyrene | 665 | 2600 | 3300 | 16000 | 153 | 1398 |
| Silver | 1000 | 3700 | 3100 | - | 730 | 1770 |
| Zinc | 150000 | 410000 | 410000 | 1600000 | 124000 | 271000 |

**APPENDIX H
STREAM HABITAT FORMS**

Part I - Stream Physical Characteristics Worksheet

Observers: W. Skalak, J. Peimado Date: 04-19-01 Time: 1300 Weather conditions: Overcast, 70°F

Stream: BML Location of site: Station 1 BML Length of stream reach: _____

Stream Segment No.: _____ Observed Stream Uses: _____ Aesthetics (circle one): (1) wilderness (2) natural **(3) common** (4) offensive

Stream Type (Circle One): perennial or intermittent w/ perennial pools Stream Bends: No. Well Defined _____; No. Moderately Defined _____; No. Poorly Defined _____

Channel Obstructions/Modifications: _____ No. of Riffls: _____ Channel Flow Status (circle one): high moderate low no flow

Riparian Vegetation (%):

Left Bank: Trees _____ Shrubs _____ Grasses, Forbs 80 Cult. Fields _____ Other _____
 Right Bank: Trees _____ Shrubs _____ Grasses, Forbs _____ Cult. Fields _____ Other _____

| | | | | | | | | | | |
|----------------------|--|---------------------|---------------------------------|--|--|--|---|----------------------|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | 15 | 0 | Thalweg Depth: _____ Depth at Sample Point _____ | | | | | | 0 |
| | Habitat Type (Circle One) Rifle Run Glide Pool | | | Dominant Substrate Type | | | Dominant Types Riparian Vegetation: | | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | | | Silty Bottom | | | Left Bank: Cattails and grasses around bank Right Bank: Cattails and grasses around bank | | | |
| | Width of Natural Buffer Vegetation (m) LB: _____ RB: 5ft | | | Instream Cover Types: | | | | | | |
| | | | | % Instream Cover | | | | | | |

| | | | | | | | | | | |
|----------------------|--|---------------------|---------------------------------|---|--|--|-------------------------------------|----------------------|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | | | Thalweg Depth: _____ | | | | | | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | | | Dominant Substrate Type | | | Dominant Types Riparian Vegetation: | | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | | | | | | Left Bank: Right Bank: | | | |
| | Width of Natural Buffer Vegetation (m) LB: _____ RB: _____ | | | Instream Cover Types: | | | | | | |
| | | | | % Instream Cover | | | | | | |

Part I - Stream Physical Characteristics Worksheet

Observers: W. Skalak, J. Peimado Date: 04-19-01 Time: 1335 Weather conditions: Overcast, 70°F

Stream: BML Location of site: Station 2 Length of stream reach:

Stream Segment No.: _____ Observed Stream Uses: _____ Aesthetics (circle one): (1) wilderness (2) natural **(3) common** (4) offensive

Stream Type (Circle One): perennial or intermittent w/ perennial pools Stream Bends: No. Well Defined _____; No. Moderately Defined _____; No. Poorly Defined _____

Channel Obstructions/Modifications: _____ No. of Riffles: _____ Channel Flow Status (circle one): high moderate low no flow

Riparian Vegetation (%):

Left Bank: Trees _____ Shrubs _____ Grasses, Forbs 75 Cult. Fields _____ Other _____
 Right Bank: Trees _____ Shrubs _____ Grasses, Forbs _____ Cult. Fields _____ Other _____

| | | | | | | | | | | |
|----------------------|--|---------------------|---------------------------------|--|--|--|--|--|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | 15 | <5 | Thalweg Depth: _____ Depth at Sample Point _____ | | | | 15 | <5 | 0 |
| | Habitat Type (Circle One) Rifle Run Glide Pool | | | Dominant Substrate Type Silty Bottom | | | | Dominant Types Riparian Vegetation: Left Bank: Cattails, grasses Right Bank: _____ | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | | | Width of Natural Buffer Vegetation (m) LB: _____ RB: 5ft | | | | Instream Cover Types: % Instream Cover | | |

| | | | | | | | | | | |
|----------------------|--|---------------------|---------------------------------|--|--|--|--|--|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | | | Thalweg Depth: _____ | | | | | | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | | | Dominant Substrate Type | | | | Dominant Types Riparian Vegetation: Left Bank: _____ Right Bank: _____ | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | | | Width of Natural Buffer Vegetation (m) LB: _____ RB: _____ | | | | Instream Cover Types: % Instream Cover | | |

Part I - Stream Physical Characteristics Worksheet

Observers: W. Skalak, J. Peimado Date: 04-19-01 Time: 1400 Weather conditions: Overcast, 70°F

Stream: BML Location of site: Station 3 Length of stream reach:

Stream Segment No.: _____ Observed Stream Uses: _____ Aesthetics (circle one): (1) wilderness (2) natural **(3) common** (4) offensive

Stream Type (Circle One): perennial or intermittent w/ perennial pools Stream Bends: No. Well Defined _____; No. Moderately Defined _____; No. Poorly Defined _____

Channel Obstructions/Modifications: _____ No. of Riffles: _____ Channel Flow Status (circle one): high moderate low no flow

Riparian Vegetation (%):

Left Bank: Trees _____ Shrubs _____ Grasses, Forbs 80 Cult. Fields _____ Other _____
 Right Bank: Trees _____ Shrubs _____ Grasses, Forbs _____ Cult. Fields _____ Other _____

| | | | | | | | | | | |
|----------------------|---|---------------------|---------------------------------|---|-----------------------|-------|--|---|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | 20 | | Thalweg Depth | Depth at Sample Point | 3.5ft | | 20 | <10 | 10 |
| | Habitat Type (Circle One) Rifle Run Glide Pool | | | Dominant Substrate Type | | | | Dominant Types Riparian Vegetation: | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | | | Silty unstable bottom | | | | Left Bank: Grass and some cattails Right Bank: | | |
| | Width of Natural Buffer Vegetation (m) LB: _____ RB: _____ | | | | | | | Instream Cover Types: | | |
| | | | | | | | | % Instream Cover | | |

| | | | | | | | | | | |
|----------------------|---|---------------------|---------------------------------|--|--|--|--|-------------------------------------|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | | | Thalweg Depth: | | | | | | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | | | Dominant Substrate Type | | | | Dominant Types Riparian Vegetation: | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | | | Width of Natural Buffer Vegetation (m) LB: _____ RB: _____ | | | | Left Bank: Right Bank: | | |
| | | | | | | | | Instream Cover Types: | | |
| | | | | | | | | % Instream Cover | | |

Appendix H
Stream Habitat Summary

| Sample Location Site Number | Units | Bryan Municipal Lake Station 1 | Bryan Municipal Lake Station 2 | Bryan Municipal Lake Station 3 |
|--------------------------------|---------|--------------------------------------|--------------------------------------|--------------------------------------|
| Date | | 04/19/01 | 04/19/01 | 04/19/01 |
| Aesthetics | | Common | Common | Common |
| Stream Bends | | | | |
| Obstructions | | | | |
| Riffles | | | | |
| Flow Status | | | | |
| Riparian Vegetation: | | | | |
| Trees | % | | | |
| Shrubs | % | | | |
| Grass, Forbs | % | 80 | 75 | 80 |
| Cultivated Fields | % | | | |
| Stream Width | (ft) | 3 | 4 | 4 |
| Maximum Depth | (ft) | | | |
| In-Stream Vegetation Type | % | | | |
| In-Stream Cover | | | | |
| Dominant Substrate Type | | Silty bottom | Silty bottom | Silty unstable bottom |
| Bank Erosion | % | | <5 | <10 |
| Average Bank Slope | degrees | 15 | 15 | 20 |
| Tree Canopy | % | 0 | 0 | 10 |

Part I - Stream Physical Characteristics Worksheet

Observers: _____ Date: 04-19-01 Time: 1500 Weather conditions: Overcast, 70°F

Stream: FFL Location of site: Station 4 Length of stream reach: _____

Stream Segment No.: _____ Observed Stream Uses: _____ Aesthetics (circle one): (1) wilderness (2) natural (3) common (4) offensive

Stream Type (Circle One): perennial or intermittent w/ perennial pools Stream Bends: No. Well Defined _____; No. Moderately Defined _____; No. Poorly Defined _____

Channel Obstructions/Modifications: _____ No. of Riffls: _____ Channel Flow Status (circle one): high moderate low no flow

Riparian Vegetation (%):

Left Bank: Trees 10 Shrubs _____ Grasses, Forbs 90 Cult. Fields _____ Other _____
 Right Bank: Trees _____ Shrubs _____ Grasses, Forbs _____ Cult. Fields _____ Other _____

| | | | | | | | | | | |
|----------------------|--|-------------------------|------------------------------------|--|--|------------------|--|----------------------|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | 20 | <10 | Thalweg Depth: _____ Depth at Sample Point _____ | | | | | | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | Dominant Substrate Type | | Dominant Types Riparian Vegetation: | | | | % Gravel or Larger | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | Silty | Left Bank: Cattails Right Bank: | | | | | | 0 | |
| | Width of Natural Buffer Vegetation (m) LB: _____ RB: 17ft | Instream Cover Types: | | | | % Instream Cover | | | | |

| | | | | | | | | | | |
|----------------------|--|--|---------------------------------|---|--|--|--|----------------------|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | | | Thalweg Depth: _____ | | | | | | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | Dominant Substrate Type | | Dominant Types Riparian Vegetation: | | | | % Gravel or Larger | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | Width of Natural Buffer Vegetation (m) LB: _____ RB: _____ | Left Bank: Right Bank: | | | | | | % Instream Cover | |

Part I - Stream Physical Characteristics Worksheet

Observers: W. Skalak, J. Peimado Date: 04-19-01 Time: 1525 Weather conditions: Overcast, 70°F

Stream: FFL Location of site: Station 5 Length of stream reach:

Stream Segment No.: _____ Observed Stream Uses: _____ Aesthetics (circle one): (1) wilderness (2) natural (3) common (4) offensive

Stream Type (Circle One): perennial or intermittent w/ perennial pools Stream Bends: No. Well Defined _____; No. Moderately Defined _____; No. Poorly Defined _____

Channel Obstructions/Modifications: _____ No. of Riffles: _____ Channel Flow Status (circle one): high moderate low no flow

Riparian Vegetation (%):

Left Bank: Trees _____ Shrubs 20 Grasses, Forbs 80 Cult. Fields _____ Other _____
 Right Bank: Trees _____ Shrubs _____ Grasses, Forbs _____ Cult. Fields _____ Other _____

| | | | | | | | | | | |
|----------------------|--|-------------------------|---------------------------------|--|--|--|--|----------------------|----------------------------------|--------------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | 25 | 15 | Thalweg Depth: _____ Depth at Sample Point _____ 4ft _____ | | | | | | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | Dominant Substrate Type | | Dominant Types Riparian Vegetation: | | | | | | % Gravel or Larger |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | Silty Bottom | | Left Bank: Cattails, small trees Right Bank: | | | | | | <5% |
| | Width of Natural Buffer Vegetation (m) LB: RB: 5ft | Instream Cover Types: | | | | | | | | % Instream Cover |

| | | | | | | | | | | |
|----------------------|--|--|---------------------------------|---|--|--|--|----------------------|----------------------------------|--------------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | | | Thalweg Depth: _____ | | | | | | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | Dominant Substrate Type | | Dominant Types Riparian Vegetation: | | | | | | % Gravel or Larger |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | Width of Natural Buffer Vegetation (m) LB: RB: | | Left Bank: Right Bank: | | | | | | % Instream Cover |
| | Width of Natural Buffer Vegetation (m) LB: RB: | Instream Cover Types: | | | | | | | | % Instream Cover |

Part I - Stream Physical Characteristics Worksheet

Observers: W. Skalak, J. Peimado Date: 04-19-01 Time: 1550 Weather conditions: Overcast, 70°F

Stream: FFL Location of site: Station 6 Length of stream reach:

Stream Segment No.: _____ Observed Stream Uses: _____ Aesthetics (circle one): (1) wilderness (2) natural (3) common (4) offensive

Stream Type (Circle One): perennial or intermittent w/ perennial pools Stream Bends: No. Well Defined _____; No. Moderately Defined _____; No. Poorly Defined _____

Channel Obstructions/Modifications: _____ No. of Riffles: _____ Channel Flow Status (circle one): high moderate low no flow

Riparian Vegetation (%):

Left Bank: Trees 20 Shrubs _____ Grasses, Forbs 80 Cult. Fields _____ Other _____

Right Bank: Trees _____ Shrubs _____ Grasses, Forbs _____ Cult. Fields _____ Other _____

| | | | | | | | | | | |
|----------------------|--|---------------------|---------------------------------|--|-----------------------|------|----|--|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | 20 | 10 | Thalweg Depth | Depth at Sample Point | 11ft | 30 | 10 | 0 | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | | | Dominant Substrate Type Silty Bottom | | | | Dominant Types Riparian Vegetation: Left Bank: Cattails, Small Trees Right Bank: | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | | | Width of Natural Buffer Vegetation (m) LB: RB: 5ft | | | | Instream Cover Types: % Instream Cover | | |

| | | | | | | | | | | |
|----------------------|--|---------------------|---------------------------------|--|--|--|--|--|----------------------------------|-----------------|
| Location of Transect | Stream Width (m) | Left Bank Slope (°) | Left Bank Erosion Potential (%) | Stream Depths (m) at Points Across Transect | | | | Right Bank Slope (°) | Right Bank Erosion Potential (%) | Tree Canopy (%) |
| | | | | Thalweg Depth: | | | | | | |
| | Habitat Type (Circle One) Rifle Run Glide Pool | | | Dominant Substrate Type | | | | Dominant Types Riparian Vegetation: Left Bank: Right Bank: | | |
| | Algae or Macrophytes (Circle One) Abundant Common Rare Absent | | | Width of Natural Buffer Vegetation (m) LB: RB: | | | | Instream Cover Types: % Instream Cover | | |

Appendix H
Stream Habitat Summary
Finfeather Lake

| Sample Location Site Number | Units | Finfeather Lake Station 4 | Finfeather Lake Station 5 | Finfeather Lake Station 6 |
|--------------------------------|---------|------------------------------|------------------------------|------------------------------|
| Date | | 04/19/01 | 04/19/01 | 04/19/01 |
| Aesthetics | | | | |
| Stream Bends | | | | |
| Obstructions | | | | |
| Riffles | | | | |
| Flow Status | | | | |
| Riparian Vegetation: | | | | |
| Trees | % | 10 | 20 | 20 |
| Shrubs | % | 90 | 80 | 80 |
| Grass, Forbs | % | | | |
| Cultivated Fields | % | | | |
| Stream Width | (ft) | 4 | 4 | 11 |
| Maximum Depth | (ft) | | | |
| In-Stream Vegetation Type | | | | |
| In-Stream Cover | % | | | |
| Dominant Substrate Type | | | | |
| Bank Erosion | % | Silty <10 | Silty bottom 15 | Silty bottom 10 |
| Average Bank Slope | degrees | 20 | 25 | 20-30 |
| Tree Canopy | % | | | 0 |